

# Ocean and Coastal Management

## Multi-zone marine protected areas: assessment of ecosystem and fisheries benefits using multiple ecosystem models

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<b>Abstract:</b>	<p>The current alarming state of many coastal ecosystems calls for the development of tools to support recovery of exploited stocks, ensure their sustainable exploitation and protect marine ecosystems. Multi-zone marine protected areas (MPAs) are often advocated to reconcile conservation and fisheries benefits. However, whether such types of MPAs can really provide both benefits is still uncertain. Here, we analysed three existing Northwestern Mediterranean multi-zone MPAs (Cerbère-Banyuls, Cap de Creus and Medes Islands), each including a fully protected area, and one or more partially protected areas, using a comparative temporal ecosystem modelling approach. Unprotected areas surrounding the MPAs were also included. Our results showed similarities in ecosystem structure and functioning between the three multi-zone MPAs, potentially due to similarities in MPA configuration, enforcement and time since protection. The MPA providing the smallest benefits was the most recent and least enforced. Small-scale and recreational fisheries had strong negative impacts on target species and were driven by the competition for resources between both fisheries. Temporal increases of benefits were small, whenever detected, like slight recoveries of some target species and ecological indicators, mostly in Cerbère-Banyuls and Medes Islands MPAs. Our results confirm the benefits of protection to coastal marine resources and ecosystems when areas are enforced but highlight the current limitations of the three MPAs due to their small size and the significant impacts of small-scale and recreational fisheries. This study illustrates the capability to evaluate protection effects of small multi-zone MPAs with an ecosystem modelling perspective and represents the baseline to develop future scenarios of alternative management</p>

	options to foster ecosystem recovery and resource rebuilding in the studied MPAs.
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11<sup>th</sup> of November 2019

Barcelona, Spain

To the editor of *Ocean and Coastal Management*,

Dear editor,

Please find attached our manuscript **“Multi-zone marine protected areas: assessment of ecosystem and fisheries benefits using multiple ecosystem models”** to be considered for publication in *Ocean and Coastal management*. Our manuscript is intended to be a Research Article.

In this study we modelled three relatively well studied MPAs from the Northwestern Mediterranean Sea with a historical perspective: Cerbère-Banyuls, Cap de Creus and Medes Islands. Our aims were to: (1) quantify the main structural and functional traits of the three multi-zone MPAs, and (2) assess how exploitation and protection regimes affected these marine ecosystems over time, using a comparative approach to identify commonalities and differences between MPAs. The food web models were developed using the Ecopath with Ecosim (EwE) modelling approach. The baseline models of each MPA were built taking into account the contribution of each zone to the whole MPA based on a previous study that is already submitted (Vilas et al., submitted. *Aquatic Conservation: Marine and Freshwater Ecosystems*).

Our results showed similarities in ecosystem structure and functioning between the three multi-zone MPAs, potentially due to similarities in MPA configuration, enforcement and time since protection. Small-scale and recreational fisheries had strong negative impacts on target species and were driven by the competition for resources between both fisheries. Temporal increases of benefits were small, whenever detected, like slight recoveries of some target species and ecological indicators, mostly in Cerbère-Banyuls and Medes Islands MPAs. Our results confirm the benefits of protection to coastal marine resources and ecosystems when areas are enforced but highlight the current limitations of the three MPAs due to their small size and the significant impacts of small-scale and recreational fisheries.

This study presents the first development of MPAs temporal models that are nearly located, potentially constituting a network of protected areas with an historical

perspective in the Mediterranean Sea. This study represents an important step forward in assessing multi-zone MPAs putative effects over time on recovering marine resources and ecosystems in the Northwestern Mediterranean Sea.

In this study we used recent developments of the EwE approach following the best practices. Specifically, we used the PREBAL analysis (Link et al., 2010. *Ecological Modelling* 12, 1580-1591) to ensure that model parameters obeyed general ecologic principles and to guide the balancing procedure. In addition, we used the ECOIND plug-in (Coll et al., 2017. *Environmental Modelling and Software* 89, 120-130) and the Ecological Network Analysis module in EwE to calculate several ecological indicators. We also addressed uncertainty of input data on results by using the Monte Carlo uncertainty routine and the ECOIND and ECOSAMPLER plug-ins (Steenbeek et al., 2018. *SoftwareX* 7, 198-204).

I confirm that all coauthors of the manuscript agreed to be listed and approve the submitted version of the manuscript.

We confirm that the manuscript was not submitted to other journals and it is original.

We hope this study is of interest for *Ocean and Coastal Management* and we look forward to hearing from you,

Yours sincerely,



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**Multi-zone marine protected areas: assessment of ecosystem and fisheries benefits using multiple ecosystem models**

## **Multi-zone marine protected areas: assessment of ecosystem and fisheries benefits using multiple ecosystem models**

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3 **benefits using multiple ecosystem models**

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32

33 **Abstract**

34 The current alarming state of many coastal ecosystems calls for the development of tools to  
35 support recovery of exploited stocks, ensure their sustainable exploitation and protect  
36 marine ecosystems. Multi-zone marine protected areas (MPAs) are often advocated to  
37 reconcile conservation and fisheries benefits. However, whether such types of MPAs can  
38 really provide both benefits is still uncertain. Here, we analysed three existing  
39 Northwestern Mediterranean multi-zone MPAs (Cerbère-Banyuls, Cap de Creus and Medes  
40 Islands), each including a fully protected area, and one or more partially protected areas,  
41 using a comparative temporal ecosystem modelling approach. Unprotected areas  
42 surrounding the MPAs were also included. Our results showed similarities in ecosystem  
43 structure and functioning between the three multi-zone MPAs, potentially due to  
44 similarities in MPA configuration, enforcement and time since protection. The MPA  
45 providing the smallest benefits was the most recent and least enforced. Small-scale and  
46 recreational fisheries had strong negative impacts on target species and were driven by the  
47 competition for resources between both fisheries. Temporal increases of benefits were  
48 small, whenever detected, like slight recoveries of some target species and ecological  
49 indicators, mostly in Cerbère-Banyuls and Medes Islands MPAs. Our results confirm the  
50 benefits of protection to coastal marine resources and ecosystems when areas are enforced  
51 but highlight the current limitations of the three MPAs due to their small size and the  
52 significant impacts of small-scale and recreational fisheries. This study illustrates the  
53 capability to evaluate protection effects of small multi-zone MPAs with an ecosystem  
54 modelling perspective and represents the baseline to develop future scenarios of alternative  
55 management options to foster ecosystem recovery and resource rebuilding in the studied  
56 MPAs.

57 **Keywords:** Northwestern Mediterranean Sea; Ecopath model; Marine Protected Areas,  
58 Small-scale fisheries; Recreational fisheries.

59

## 60 1. Introduction

61 Coastal and marine ecosystems provide multiple ecosystem goods and services, including  
62 seafood, coastal protection, carbon sequestration, recreational and other cultural services,  
63 and support for human livelihoods and well-being ([Costanza et al., 1997](#); [Martínez et al.,  
64 2007](#)). However, the production of these ecosystem goods and services can be degraded as  
65 human demand for resources and the associated activities increase in number, frequency,  
66 magnitude and spatial coverage ([MEA, 2005](#); [Worm et al., 2006](#); [Pauly and Zeller, 2016](#)).  
67 Currently, marine ecosystems are under increasing threat from a range of stressors  
68 including overfishing, climate change, biological invasions, pollution, aquaculture and  
69 habitat modification, directly or indirectly caused by multiple human activities ([Costello et  
70 al., 2010](#); [Halpern et al., 2015](#)). The interaction and cumulative effects of multiples  
71 stressors are highly complex and their combined impacts on marine ecosystems are largely  
72 unknown or even highly uncertain ([Crain et al., 2009](#); [Côté et al., 2016](#)).

73 The Mediterranean Sea is not an exception. Although it is considered a hotspot of  
74 biodiversity ([Coll et al., 2010](#)), the Mediterranean Sea has a long history of human  
75 disturbance and exploitation ([Lotze et al., 2011](#)) with dramatic changes to its ecosystems  
76 taking place for long time, but especially during the last 50 years ([Azzurro et al., 2011](#);  
77 [Maynou et al., 2011](#); [Piroddi et al., 2017](#)). Because of intense human pressures, the  
78 Mediterranean Sea has been defined as a sea “under siege”, and classified among the most  
79 impacted ecoregions of the world ([Coll et al., 2012](#); [Micheli et al., 2013](#); [Halpern et al.,  
80 2015](#)).

81 The recognition of the interactive nature of stressors and the interconnection between  
82 marine organisms, multiple human impacts, and ecosystem functioning and services has led  
83 to increasing demand for an ecosystem-based management (EBM) ([Rosenberg and  
84 McLeod, 2005](#); [Leslie and McLeod, 2007](#)). The overall objective of this approach is to  
85 maintain ecosystems as a whole in healthy, productive and resilient conditions allowing  
86 them to provide the needed ecosystem services to society ([McLeod et al., 2005](#)). Within  
87 this context, Marine Protected Areas (MPAs hereafter) are considered to be a key  
88 management tool for EBM as they are expected to mitigate human impacts, recover  
89 exploited resources and promote their sustainable use, conserve or restore habitats and

90 biodiversity, maintain and enhance ecosystem services and reduce conflicts between users  
91 ([Lester et al., 2009](#); [Halpern et al., 2010](#); [Leenhardt et al., 2015](#)). In fact, MPAs are being  
92 increasingly used worldwide for conservation and fisheries management purposes  
93 ([Boonzaier and Pauly, 2016](#); [Morgan et al., 2018](#)). MPAs conservation benefits largely vary  
94 due to their intrinsic features, the level of enforcement, their conservation goals and  
95 organization ([Guidetti et al., 2008](#); [Di Franco et al., 2016](#); [Giakoumi et al., 2018](#); [Scianna et](#)  
96 [al., 2019](#)). A MPA is a discrete area of the sea established to achieve the long-term  
97 conservation of natural resources therein ([Claudet, 2011](#)). The level of protection within  
98 MPAs can vary from areas of full protection or no-take areas (FPA), where all extractive  
99 activities (e.g. fishing) are prohibited while some non-extractive activities (e.g. diving) can  
100 be allowed; to areas of partial protection (PPA), where some human activities (e.g. small-  
101 scale and recreational fisheries) are allowed but regulated ([Horta e Costa et al., 2016](#)). In  
102 addition, networks of MPAs could provide greater benefits than the sum of individual  
103 MPAs benefits ([Gaines et al., 2010](#); [Grorud-Colvert et al., 2014](#)). A network of MPAs is a  
104 “collection of individual MPAs that operates synergistically, at different spatial scales and  
105 with a range of protection levels that have been designed to fulfill ecological objectives  
106 more effectively than individual MPAs could alone” ([IUCN-WCPA, 2008](#)).

107 Within the EBM, ecological modelling has emerged as a suitable tool to integrate available  
108 ecological information into a coherent form to obtain insights about how ecosystems are  
109 structured, functioning, impacted by human activities and environment, and delivering  
110 ecosystem services ([Link, 2010a](#); [Christensen and Maclean, 2011](#)). Although quantitative  
111 models have become a key tool to assess the extent to which MPAs are able to achieve their  
112 conservation objectives ([Fulton et al., 2015a](#)), the understanding of MPA effects on  
113 ecological processes and ecosystem function is still scarce ([Cheng et al., 2019](#)). For  
114 instance, it is mostly unknown if protection benefits inside fully protected areas extend to  
115 partially areas and beyond, and if those benefits have the potential to cascade through the  
116 ecosystem components. There is little evidence that multi-zone MPAs (i.e., those including  
117 both FPAs and PPAs) provide fisheries benefits inside partially protected areas and it is  
118 unknown what are the drivers of multi-zone MPAs. One of the most used approaches  
119 within the EBM is the Ecopath with Ecosim (EwE) modelling toolbox, which has been

120 widely applied to model aquatic food webs and assess the impact of human activities  
121 ([Heymans et al., 2014](#); [Colléter et al., 2015](#)).

122 Previously, several studies have modelled MPAs in the Mediterranean Sea ([Albouy et al.,](#)  
123 [2010](#); [Libralato et al., 2010](#); [Valls et al., 2012](#); [Prato et al., 2016](#); [Vilas et al., Submitted](#)).  
124 However, none of the existing studies modelled simultaneously the multi-zone nature of  
125 Mediterranean MPAs nor a series of MPAs that are nearly located, potentially constituting  
126 a network of protected areas. In addition, the study of long-term effects of MPAs has not  
127 been pursued due to difficulties to gather historical data about protection effects. In this  
128 study we modelled three relatively well studied MPAs from the northwestern  
129 Mediterranean Sea with a historical perspective: Cerbère-Banyuls, Cap de Creus and Medes  
130 Islands. Our aims were to: (1) quantify the main structural and functional traits of the three  
131 multi-zone MPAs, and (2) assess how exploitation and protection regimes affected these  
132 marine ecosystems over time, using a comparative approach to identify commonalities and  
133 differences between MPAs.

134

## 135 2. Materials & Methods

### 136 2.1. Study area

137 We studied three MPAs in the Northwestern Mediterranean Sea: Cerbère-Banyuls (France),  
138 Cap de Creus (Spain) and Medes Islands (Spain) (Fig. 1 and Table 1). These MPAs include  
139 one Fully Protected Area (FPA) and one (Cerbère-Banyuls and Medes Islands) or more  
140 (Cap de Creus) Partially Protected Areas (PPAs) (Fig. 1). While in FPA all extractive  
141 activities are forbidden, in PPAs both small-scale and recreational fisheries are allowed  
142 except in Medes Islands MPA, where recreational fisheries are prohibited.

143 The three MPAs comprise a similar depth range between 0 and 60 meters, and some  
144 vulnerable and/or endemic species inhabit their areas such as the Mediterranean seagrass  
145 (*Posidonia oceanica*), dusky grouper (*Epinephelus marginatus*) and red coral (*Corallium*  
146 *rubrum*). In addition to the protected zone scheme (FPA and PPA), we included in our  
147 study the immediate unprotected areas (UPAs) surrounding each MPA (Fig. 1 and Table 1),  
148 in order to add the adjacent ecosystem without protection where some species inhabiting  
149 the MPAs can theoretically forage and move to ([Di Lorenzo et al., 2016](#)). The boundaries  
150 of the unprotected areas were selected taking into account that they had similar features to  
151 the MPA, and they were adjacent to the MPA (Fig. 1).

### 152 2.2. Modelling approach – Ecopath and Ecosim

153 The three MPA models were developed using Ecopath with Ecosim approach (EwE 6.6  
154 version) ([Christensen and Walters, 2004](#); [Christensen et al., 2008](#)). We used the static  
155 Ecopath model to provide a quantitative representation of the food web as a “snapshot” in  
156 terms of flows and biomasses for a defined period of time. To develop the food-web model  
157 we used functional groups, which could be ontogenetic fractions of a species, single species  
158 or groups of species sharing common ecological and behavioral traits. The baseline models  
159 of each MPA were developed taking into account the contribution of each zone to the  
160 whole MPA based on a previous study ([Vilas et al., Submitted](#)). After the static models of  
161 the three MPAs were developed, we used the time-dynamic model Ecosim, which describes  
162 the temporal dynamics of species biomass and flows over time by accounting for changes  
163 in predation, consumption rate, fishing and the environment ([Walters et al., 1997](#);

164 [Christensen and Walters, 2004](#)). A short description of the modelling approach is given in  
165 the Supplementary Online Material (SOM thereafter) 1; and a detailed explanation of the  
166 algorithms and equations of the EwE approach are given in [Christensen and Walters \(2004\)](#)  
167 and [Heymans et al. \(2016\)](#).

## 168 **2.3. Functional groups and input data**

### 169 **2.3.1. Functional groups**

170 A meta-web structure defined for the Western Mediterranean Sea model ([Coll et al., 2019b](#))  
171 developed under the Safenet Project (Sustainable fisheries in EU Mediterranean waters  
172 through a network of MPAs) was used and adapted to our study area. Specifically, we  
173 removed those functional groups (hereafter FGs), which did not occur in the study area.  
174 The final food-web structure contained 64 FGs for Cerbère-Banyuls MPA and 67 FGs for  
175 Cap de Creus MPA and Medes Islands MPA (SOM 2).

### 176 **2.3.2. Ecopath input data and balancing procedure**

177 The Ecopath models represented a situation of the Cerbère-Banyuls MPA for 2013, while  
178 the Cap de Creus MPA and the Medes Islands MPA represented an average situation for  
179 the period 2005-2008 and 2000-2004, respectively. These periods of time to model the  
180 MPAs were selected considering the best available biomass data (SOM 2 for details on the  
181 parameterization of each functional group). Input parameters were obtained using similar  
182 procedures as those documented in [Coll et al. \(2006\)](#), [Corrales et al. \(2015\)](#), [Piroddi et al.](#)  
183 [\(2015\)](#) and [Vilas et al. \(Submitted\)](#). Biomass and catch data were scaled using data from  
184 models of each management unit (FPA, PPA and UPA) developed previously ([Coll et al.,](#)  
185 [2019c](#); [Vilas et al., Submitted](#)) by the percentage of the area of each management unit to the  
186 whole MPA (Table 1).

187 Biomass estimates were obtained from different sources from the study area or surrounding  
188 areas (SOM 2 for details on parameterization of each functional group). The three MPA  
189 models shared biomass data on some FGs (marine mammals, seabirds, sea turtles, pelagic  
190 fish, some invertebrates' groups, primary producers, zooplankton and phytoplankton) due  
191 to the lack of local data and/or the closeness among these MPAs. Fish and invertebrate

192 biomasses were estimated using data from (1) underwater visual census (UVC) ([Claudet,](#)  
193 [2013](#); [Hereu Fina et al., 2017](#)), (2) bottom trawl surveys performed in adjacent areas  
194 (Mediterranean International Trawl Surveys, MEDITS; see [Bertrand et al. \(2002\)](#)), and (3)  
195 additional information found in the literature. MEDITS trawling survey data was used to  
196 extract biomass estimates and was weighted by the bathymetry strata where species  
197 occurred. We extracted MEDITS trawling survey data from the closest stations to each  
198 MPA in order to have biomass data for those species that were not present during UVC.  
199 This assumption was assessed previously by expert knowledge from Safenet project ([Coll](#)  
200 [et al., 2019c](#)). For several pelagic fish species, available data from MEDITS and MEDIAS  
201 (acoustic) surveys were used. As these surveys did not fully cover coastal areas, scaling  
202 factors based on species depth distribution from AQUAMAPS ([Kaschner et al., 2016](#)) were  
203 used to correct biomass estimates of the pelagic fish groups.

204 Production (P/B, year<sup>-1</sup>) and consumption (Q/B, year<sup>-1</sup>) rates were estimated either using  
205 empirical equations (Heymans et al., 2016), taken from literature or from other models  
206 developed in the northwestern Mediterranean Sea ([Coll et al., 2006](#); [Corrales et al., 2015](#);  
207 [Coll et al., 2019c](#)). Additionally, local body lengths of reef-associated species obtained  
208 from UVC ([Calò et al., 2018](#)) were used to estimate those rates using empirical equations.

209 The trophic information to populate the diet matrix was compiled using published studies  
210 on stomach content analyses, giving preference to local or surrounding areas (SOM 2). Due  
211 to the small sizes of the investigated MPA and the capacity of some species to move greatly  
212 ([Gell and Roberts, 2003](#); [Grüss et al., 2011](#)), we set a fraction of the diet composition as  
213 import for all MPA based on the time that these species spent foraging outside the areas and  
214 biological and ecological traits such as size, behavior and ecology of species of each  
215 functional group (Table 1 in SOM 3).

216 Catch data were obtained from different sources (database, literature and unpublished data)  
217 (SOM 2) and included recreational and small-scale catch. Small-scale fisheries in Cerbère-  
218 Banyuls were obtained from a local study ([Prats, 2016](#)), while for Cap de Creus and Medes  
219 Islands, landings were obtained from an official dataset of the regional government of  
220 Catalonia managed by the Institute of Marine Sciences (ICM-CSIC) ([Tudó, 2017](#)). For Cap  
221 de Creus, these landings were from Llançà, Port de la Selva, Cadaqués and Roses harbours,



222 where the main fleet operating is coming from. These landings were scaled by the number  
223 of boats and months fishing inside the MPA ([Gómez et al., 2006](#)). For Medes Islands,  
224 landings were from l’Estartit harbour. To these official landings, we incorporated  
225 percentages of discards from a literature review based on the Spanish Mediterranean and  
226 Gulf of Cadiz region ([Coll et al., 2014](#)). Regarding recreational fisheries catch, we used  
227 information from [Ivanhoff et al. \(2010\)](#) (Cerbère-Banyuls), [Lloret et al. \(2008a\)](#) and [Lloret](#)  
228 [et al. \(2008b\)](#) (Cap de Creus) and [Sacanell \(2012\)](#) (Medes Islands).

229 To achieve mass-balance of the baseline models, we applied a manual procedure following  
230 a top-down approach modifying appropriate input parameters (Table 2 in SOM 3) (starting  
231 from the groups with higher trophic levels) and following the best practice guidelines  
232 provided in the literature ([Heymans et al., 2016](#)). In addition, the PREBAL analysis was  
233 used to ensure that the model parameters obeyed general ecologic principles and to guide  
234 the balancing procedure ([Link, 2010b](#)). Detailed explanations of these procedures are given  
235 in SOM 3 and the final diet matrixes are provided in SOM 4.

236 The pedigree routine ([Christensen and Walters, 2004](#)) was used to quantify the uncertainty  
237 associated with the input parameters and the quality of the models and to validate choices  
238 made in balancing the model. A detailed explanation of this routine is given in SOM 3.

#### 239 **2.3.4. Ecosim historical time series fitting.**

240 The model representing Cerbère-Banyuls MPA ecosystem during the 2013–2017 period  
241 was fitted to time series of historical data, while for Cap de Creus MPA and the Medes  
242 Islands MPA, the Ecosim model was fitted to time series of 2008-2017 and 2004-2017,  
243 respectively (Table 1 in SOM 5). Overall, we used historical fishing effort trends to drive  
244 the fishing fisheries of the models, while biomass and catch time series were used to  
245 calibrate the model and compare predicted to observed results (Table 1 in SOM 5),  
246 respectively, following previous studies in the Mediterranean Sea and best practices of  
247 EwE ([Coll et al., 2013](#); [Heymans et al., 2016](#); [Corrales et al., 2017](#)).

248 Available time series of fishing activities included data on nominal fishing effort, expressed  
249 in number of boats per year for the small-scale fisheries and in hooks per day for the  
250 recreational fisheries. Data for the small-scale fisheries were obtained from [IFREMER](#)

251 [\(2015\)](#) (Cerbère-Banyuls) and from the Ministry of Agriculture, Livestock, Fisheries and  
252 Food of the Catalan Government (Cap de Creus and Medes Islands), while data for  
253 recreational fisheries were collected from [IFREMER \(2015\)](#) and [Ivanhoff et al. \(2010\)](#)  
254 (Cerbère-Banyuls MPA) and from [Lloret et al. \(2008a\)](#), [Lloret et al. \(2008b\)](#), [Font and](#)  
255 [Lloret \(2010\)](#) and [Font and Lloret \(2011\)](#) (Cap de Creus and Medes Islands MPAs).

256 Available relative and absolute observed biomass for most of the demersal groups were  
257 obtained from MEDITS survey and UVC ([Claudet, 2013](#); [Hereu Fina et al., 2017](#); [Calò et](#)  
258 [al., 2018](#)). No historical catch data were available for Cerbère-Banyuls MPA (Table 1 in  
259 SOM 5). Absolute observed landings data for Cap de Creus and Medes Islands MPA were  
260 obtained from an official dataset of the regional government of Catalonia managed by the  
261 Institute of Marine Sciences (ICM-CSIC) ([Tudó, 2017](#)). Regarding recreational fisheries,  
262 historical catches of Cap de Creus and Medes MPA were reconstructed based on their  
263 historical effort trends.

264 To fit the Ecosim model of the three MPAs, we used the Stepwise Fitting Procedure ([Scott](#)  
265 [et al., 2016](#)), which automates the model fitting procedure described by [Mackinson et al.](#)  
266 [\(2009\)](#) and [Heymans et al. \(2016\)](#). The fitting procedure tests alternative hypotheses related  
267 to the impact of fishing, changes in predator-prey dynamics (vulnerabilities), changes in  
268 primary production (production anomalies) or all of the above together (Table 2)  
269 ([Mackinson et al., 2009](#); [Heymans et al., 2016](#)). A primary production anomaly is a forcing  
270 function applied to the primary production rate (in our study both phytoplankton and  
271 benthic primary producers) that may represent historical productivity changes impacting  
272 biomasses through the ecosystem. During the fitting procedure, vulnerabilities and  
273 production anomalies were estimated to improve model fits by comparing model  
274 predictions to observed data using the sum of squares (SS) statistics. The fitting procedure  
275 finds the statistically “best fit” model based on Akaike’s Information Criterion (AIC),  
276 which penalizes for estimating too many parameters based on the number of time series  
277 available for estimating the SS ([Mackinson et al., 2009](#); [Heymans et al., 2016](#)). In this  
278 study, the maximum number of parameters that could be estimated were 19 (Cerbère-  
279 Banyuls), 51 (Cap de Creus) and 48 (Medes Islands), respectively.

280 To choose the best final model, we manually evaluated whether the parameterization  
281 process led to credible and sensible behavior ([Heymans et al., 2016](#)), following previous  
282 studies ([e.g., Corrales et al., 2017](#)).

## 283 **2.4. Model analyses and ecological indicators**

### 284 **2.4.1. Ecological indicators of initial conditions**

285 To analyse the food-web structure of the three MPAs, we used the biomasses of selected  
286 FGs, trophic flows and trophic levels (TLs) within the flow diagram. The TL was also used  
287 to analyze the ecological position of the FG of the three MPA models ([Lindeman, 1942](#);  
288 [Stergiou and Karpouzi, 2001](#))

289 Several ecological indicators were also computed to describe the structure and functioning  
290 of the ecosystems and were divided into four groups using the ECOIND plug-in ([Coll and](#)  
291 [Steenbeek, 2017](#)):

292 (1) Biomass-based: calculated from the biomass of components included in the food-  
293 web model, they could provide valuable information to evaluate MPAs effectiveness  
294 ([Micheli et al., 2004](#); [Claudet et al., 2008](#)). We included Total Biomass (TB,  $t \cdot km^{-2} \cdot year^{-1}$ ),  
295 Biomass of Fish species (FB,  $t \cdot km^{-2} \cdot year^{-1}$ ), and Kempton's Q diversity index (QI).

296 (2) Trophic-based: reflect the TLs for different groups of the food web, provide  
297 information on the structure of the ecosystem and are used to quantify the impact of fishing  
298 ([Rochet and Trenkel, 2003](#)). We selected TL of the community (TLc), TL of the  
299 community, TL of the community including organisms with  $TL \geq 3.25$  (TL3.25) and TL of  
300 the community including organisms with  $TL \geq 4$  (TL4).

301 (3) Species and size-based: based on species traits and conservation status, they could  
302 offer insights of the effects of MPAs ([Claudet et al., 2010](#)). We selected biomass of IUCN-  
303 endangered species biomass in the community ( $t \cdot km^{-2} \cdot year^{-1}$ ), mean length of fish in the  
304 community (ML, cm) and mean life span of fish community (MLS, year).

305 (4) Catch-based: based on catch, they reflect the fishing strategy of the fisheries and are  
306 used to quantify the impact of fishing ([Hilborn and Walters, 1992](#); [Pauly et al., 1998](#)). We

307 included total catch (TC  $t \cdot km^{-2} \cdot year^{-1}$ ), trophic level of the catch (TLc), intrinsic  
308 vulnerability index of catch (VI).

309 The Mixed Trophic Impact (MTI) analysis was used to quantify the direct and indirect  
310 impact in the food web that a hypothetical increase in the biomass of one functional group  
311 would have on the biomasses of all the other functional groups in the food web, including  
312 the fishing fleets ([Ulanowicz and Puccia, 1990](#); [Christensen et al., 2008](#)). To evaluate the  
313 impact of small-scale fisheries on the MPAs, the MTI was used to quantify the direct and  
314 indirect impact of each fishery on the functional groups for the studied MPAs, and their  
315 potential competition and trade-offs between them.

316 Finally, to identify the key species within the ecosystem (both keystone and structuring  
317 species), we calculated the keystone index (KSi) developed by [Valls et al. \(2015\)](#) of the  
318 most important reef functional groups. A keystone group is defined as a predator species  
319 with a high and wide impact on the food web despite its low biomass ([Paine, 1966, 1969](#);  
320 [Valls et al., 2015](#)).

#### 321 **2.4.2. Time-dynamic analyses**

322 Once the fitting procedure was completed, we used the best fit model to examine biomass  
323 and catch time series predicted by the model to explore the dynamics of selected functional  
324 groups. We selected four target species due to their role in terms of biomasses and  
325 commercial interest: the common two-banded seabream (*Diplodus vulgaris*), white  
326 seabream (*D. sargus sargus*), common dentex (*Dentex dentex*) and groupers (this group  
327 was mainly represented by *Epinephelus marginatus*). These species play an important role  
328 in the ecosystem as high trophic level predators or intermediate trophic species and are of  
329 great importance in small-scale and recreational fisheries. In addition, the previous  
330 selection of ecological indicators from the ECOIND plug-in was used to describe  
331 ecological changes in the ecosystem over time ([Coll and Steenbeek, 2017](#)).

332 We addressed the impact of uncertainty in Ecopath input parameters on Ecosim outputs  
333 (biomass and catch trends, and ecological indicators) by using the Monte Carlo (MC)  
334 uncertainty routine and the ECOSAMPLER plug-in ([Heymans et al., 2016](#); [Coll and](#)  
335 [Steenbeek, 2017](#); [Steenbeek et al., 2018](#)). We ran 500 MC simulations based on the

336 coefficient of variation obtained from the pedigree routine, which assesses the quality of the  
337 input data (SOM 3). Results from the MC simulations were used to plot the confidence  
338 intervals of the selected ecological indicators in Ecopath and to plot the 5<sup>th</sup> and 95<sup>th</sup>  
339 percentile confidence intervals for the fitted biomass and catch trends and for ecological  
340 indicators. Finally, we used the Spearman's rank correlation to evaluate if selected  
341 modelled results (time series of biomass, catch and ecological indicators) decreased or  
342 increased in time, following previous studies ([e.g., Corrales et al., 2017](#)).

### 343 **3. Results**

#### 344 **3.1. Structure and functional traits of the three MPAs**

345 The pedigree index values of the three MPA models showed similar values, ranging from  
346 0.41 to 0.51. The highest pedigree values were obtained for Cerbère-Banyuls (0.51),  
347 followed by Medes Islands (0.45) and Cap de Creus (0.41).

348 The visualization of trophic links and flows between functional groups highlighted the  
349 complexity of these coastal ecosystems due to the large number of trophic links between  
350 functional groups and the important role of detritus (FG 65) and other macro-benthos (FG  
351 51) in transferring energy up to the food web (Fig. 2).

352 The functional groups of the models ranged from trophic level (TL) of 1 for primary  
353 producers (FG. 60-67) and detritus groups (FG 66-67) to TL = 4.2 for both groups of  
354 dolphins (FG 1-2) (Fig. 2 and Table 3 in SOM 3). Invertebrates groups were classified with  
355 a TL between 2 and 3.5, with benthopelagic cephalopods showing higher TLs. Fish had  
356 TLs between 3 and 4, with the exception of salema (FG 33) and mugilidae (FG 34), which  
357 showed lower TL due to their herbivorous and detritivore behaviors. Overall, similar TLs  
358 were found for the three MPAs.

359 Total and fish biomass displayed similar patterns, with the highest biomass values found  
360 for Cerbère-Banyuls and Medes Islands and lower values in Cap de Creus (Fig. 3).  
361 Conversely, Kempton's Q Index presented a higher value in Cap de Creus, followed by  
362 Medes Islands and Banyuls-Cerbère (Fig. 3). With the exception of TL of the community 4,  
363 trophic based indicators also presented higher values for Cerbère-Banyuls, followed by

364 Medes Islands and Cap de Creus (Fig. 3). Species and size-based indicators showed that  
365 Cerbère-Banyuls presented the highest values for ML of fish community and IUCN species  
366 B, followed by Cap de Creus and Medes Islands (Fig. 3), while MLS of fish community  
367 was higher in Cap de Creus, followed by Cerbère-Banyuls and Medes Islands (Fig. 3).  
368 Total catch showed similar values between Cap de Creus and Medes Islands, while  
369 Cerbère-Banyuls had the lowest value (Fig. 3). TL of the catch and the Intrinsic  
370 Vulnerability Index presented similar values in Cerbère-Banyuls and Cap de Creus, while  
371 Medes Islands had the lowest value (Fig. 3).

372 The keystone index identified groupers (FG 27), common dentex (FG 24), other  
373 commercial medium demersal fishes (FG 31) and non-commercial medium demersal fishes  
374 (FG 32) as keystone species in the three MPAs, followed by red scorpionfish (*Scorpaena*  
375 *scrofa*) (FG 25) and Scorpaenidae (FG 26).

376 The MTI analysis highlighted that the small-scale fisheries had the most widespread  
377 negative impact on many FG of all MPAs, especially in Cerbère-Banyuls, while the impact  
378 of recreational fisheries was more prominent in Cap de Creus and Medes Islands (Fig. 5).  
379 This analysis showed strong negative impact of fisheries on target species such as other  
380 large pelagic species (FG 8), common dentex (FG 24), and groupers (FG 27) while  
381 competitors or preys of those species may be positively impacted such as white seabream  
382 (FG 22), common two-banded seabream (FG 23) and brown meagre (*Sciaena umbra*) (FG  
383 28) (Fig. 5). Results also highlighted that each fishery had a negative impact on itself due to  
384 self-competition for resources according to the MTI results and that competition between  
385 fisheries was complex (Fig 5). While small-scale fisheries had positive impacts on  
386 recreational fisheries in Cerbère-Banyuls MPA, recreational fisheries had slightly negative  
387 impacts on small-scale fisheries (Fig. 5a). On the contrary, small-scale fisheries had  
388 negative impacts on recreational fisheries in Cap de Creus and Medes Islands MPAs, while  
389 recreational fisheries had positive impacts on the small-scale fisheries (Fig. 5b and c).

### 390 **3.2. Temporal changes of MPAs**

391 The best-fitted food web temporal models were obtained when trophic interactions, fishing,  
392 and primary production anomaly were included in the model configuration for all MPAs

393 (Step 8 in Table 3). However, for all MPAs the best model was not able to reproduce the  
394 trends of white seabream, common two-banded seabream, common dentex and groupers  
395 satisfactorily, which we selected as target groups of this study. Therefore, we moved  
396 through the fitting procedure analysis to find the model that was able to reproduce the  
397 trends of most of the groups (and specifically the target groups) and was highly significant.  
398 We finally choose a model fit with 12 vulnerabilities and 3 spline points, 15 vulnerabilities  
399 and 3 spline points, and 11 vulnerabilities and 5 spline points for Cerbère-Banyuls, Cap de  
400 Creus and Medes Islands models respectively, as the best options (Step 8 in Table 3).

401 Observed biomass and catch time series were satisfactorily reproduced by model  
402 predictions for most of the target groups (Fig. 6 and SOM 6, 7 and 8) when using the  
403 chosen fitted model. The temporal models showed a non-significant biomass pattern of  
404 white seabream in Cerbère-Banyuls, firstly decreasing and later increasing, and in Cap de  
405 Creus, where firstly increased and later decreased, while in Medes Islands it presented a  
406 significant decreasing biomass trend (Fig 6). The models showed similar biomass patterns  
407 for common two-banded seabream, highlighting a non-significant trend in Cerbère-Banyuls  
408 and Cap de Creus and a significant decreasing trend in Medes Islands (Fig. 6). We observed  
409 a significant decreasing biomass trend of common dentex in Cerbère-Banyuls, while in Cap  
410 de Creus and Medes this group showed non-significant biomass trends (Fig. 6). The results  
411 highlighted a non-significant biomass pattern of groupers in Cerbère-Banyuls (Fig 6), while  
412 they significantly increased in Cap de Creus and Medes Islands (Fig 6). However, in Medes  
413 Islands the model did not capture well the overall declining trend of observations for  
414 groupers biomass.

415 Regarding the temporal changes of ecological indicators, the Kempton's Index showed a  
416 non-significant trend in Cerbère-Banyuls and Cap de Creus, while in Medes Islands, the  
417 Kempton's Index significantly increased during the simulated period (Fig. 7). The TL of  
418 the community presented a non-significant pattern in Cerbère-Banyuls and Cap de Creus,  
419 while in Medes Islands significantly decreased (Fig. 7). IUCN species biomass significantly  
420 declined in Cerbère-Banyuls, while in Cap de Creus and Medes Islands presented non-  
421 significant patterns (Fig. 7). Total catches highlighted non-significant trends in Cerbère-

422 Banyuls and Cap de Creus (in the last one firstly increasing and then decreasing), while in  
423 Medes Islands, total catches significantly decreased (Fig. 7).

#### 424 **4. Discussion**

425 Due to the deteriorating condition of many coastal ecosystems, there is a pressing need to  
426 better understand how current management options assist in recovering and conserving  
427 these ecosystems. Within this context, we developed a food-web modelling approach to  
428 quantitatively assess multi-zone MPAs putative effects over time on recovering marine  
429 resources and ecosystems in the Northwestern Mediterranean Sea.

430 Overall, ecological indicators showed similar patterns, with highest values in Banyuls-  
431 Cerbère MPA, followed by Medes Islands and Cap de Creus MPAs. This may be related to  
432 differences in the ecological effectiveness of the three MPAs, which is partly explained by  
433 MPA design, management and implementation features (e.g. the extend of area protected,  
434 fully protected area enforcement, time since protection, MPA organization) ([Claudet et al.,  
435 2008](#); [Guidetti et al., 2008](#); [Edgar et al., 2014](#); [Giakoumi et al., 2017](#); [Di Franco et al.,  
436 2018](#)). The lack of enforcement is one of the most relevant issues concerning MPAs in the  
437 Mediterranean context ([Fenberg et al., 2012](#)). Within this context, while Cerbère-Banyuls  
438 and Medes MPAs are considered to have a high level of enforcement ([Sala et al., 2012](#);  
439 [Giakoumi et al., 2017](#)), Cap de Creus has been considered a MPA only on paper due to a  
440 lack of sufficient enforcement ([Lloret et al., 2008a](#); [Lloret et al., 2008b](#)). In fact, total  
441 catches were higher in Cap the Creus than in the other two MPA; with Cerbère-Banyuls  
442 presenting the lowest total catches. Total catches are expected to be higher in well designed  
443 and enforced MPAs ([Halpern et al., 2009](#); [Di Lorenzo et al., 2016](#)), so our results could be  
444 explained by the allowed and/or real fishing effort and the small effects of these MPAs in  
445 PPAs and adjacent areas. [Horta e Costa et al. \(2016\)](#) presented a novel classification system  
446 for MPAs by scoring allowed uses in each management unit based on their impacts on  
447 biodiversity. In this scale, Cerbère-Banyuls obtained a rate of 4.7, being a highly protected  
448 area and Medes 6.4 being less well protected ([Horta e Costa et al., 2016](#)). Although Cap de  
449 Creus was not included in this study, we could expect a higher MPA index because of its  
450 smaller FPA and allowed uses. These features (smallest FPA and weak enforcement) and  
451 the fact that Cap de Creus is the newest (Table 1) MPA in the study, could suggest that the



452 year of establishment, enforcement and the size of protected areas have a strong effect on  
453 MPA effects, in line with other studies ([Claudet et al., 2008](#); [Guidetti et al., 2008](#)).

454 Our results also indicated that both the small-scale and recreational fisheries can have a  
455 notable impact affecting organisms from lower to higher trophic levels. According to our  
456 study, small-scale fisheries had the largest negative ecological impacts. In fact, this fishery  
457 tends to develop its main activity inside the MPA or in surrounding areas ([Goñi et al.,  
458 2008](#); [Stelzenmüller et al., 2008](#)). Recreational fisheries seemed to have larger impacts in  
459 the Spanish MPAs (Cap de Creus and Medes Islands MPA), while they had a negligible  
460 impact in Cerbère-Banyuls MPA, although the impact of recreational fisheries could be  
461 similar than the small-scale fisheries, as highlighted in coastal waters of Cap de Creus by  
462 [Lloret et al. \(2008a\)](#). This could be related to cultural reasons, differences in enforcement of  
463 each MPA and/or the reliability of catch data between study sites. In addition, our results  
464 highlighted that despite the recreational fisheries seemed to have an overall lower negative  
465 impact, it had stronger negative impacts on vulnerable species such as common dentex and  
466 groupers in Cap de Creus and Medes Islands MPAs, in line of the observed impact of  
467 small-scale and recreational fisheries on vulnerable species ([Lloret et al., 2019](#)), which is an  
468 important fact to consider when establishing plans for these species. Our results also  
469 highlighted the competition for resources between both (small-scale and recreational)  
470 fisheries. Competition between both fisheries has become an important issue in coastal  
471 areas (especially in surrounding areas of MPAs), as they target similar species and fishing  
472 grounds ([Chuenpagdee and Jentoft, 2018](#); [Lloret et al., 2018](#)). In small and touristic areas  
473 such as the ones in the present study, a strong competition between small-scale and  
474 recreational fishers could be expected due to the decline of the small-scale fisheries and the  
475 increasing importance of recreational fisheries as a leisure activity ([Gómez and Lloret,  
476 2017](#)).

477 The temporal dynamic model of Cerbère-Banyuls predicted an increasing biomass pattern  
478 of white seabream and common two-banded seabream while common dentex and groupers  
479 largely decreased. The decrease of common dentex and groupers could be related to the  
480 impact of fishing, although fishing effort did not increase during the study period and initial  
481 fishing mortalities were not high. The increase of both seabreams could be explained by the

482 reduction of their predators. In fact, common dentex and groupers are the most important  
483 predators of these species in terms of total consumption.

484 In contrast, the temporal dynamic model of Cap de Creus predicted a decreasing biomass  
485 pattern of white seabream, common two-banded seabream and common dentex while  
486 groupers increased. These results suggest a recovery of groupers population in Cap de  
487 Creus ([Hereu Fina et al., 2017](#)), despite low level of enforcement, that could cause biomass  
488 reductions of their preys such as both seabreams. Similarly, the temporal dynamic model of  
489 Medes Islands highlighted decreasing trends for white seabream, common two-banded  
490 seabream and common dentex, while groupers slightly increased. These results evidence  
491 that under protection the food-web effects can play an important role and there are winners  
492 and losers as a result of the new ecological state, where results of protection can include  
493 cascading effects of predators on preys species ([Edgar et al., 2014](#); [Cheng et al., 2019](#)) as  
494 well as an increase in competitive interactions ([Micheli et al., 2004](#)).

495 Ecological indicators are important within the EBM framework because they are  
496 quantitative representations of ecosystem status and provide a mean for evaluating the  
497 impact of human activities in marine ecosystems and the effectiveness of management  
498 measures ([Shin and Shannon, 2010](#); [Tam et al., 2017](#)). In our study, ecological indicators  
499 showed overall contrasting results, as some indicators indicated recovery (e.g. Kempton's  
500 Q in the three MPAs) while others showed degradation patterns (e.g., TL of the community  
501 in the three MPAs). These contrasting results could respond to the limitation in historical  
502 data to represent ecosystem dynamics in the study area and could indicate limited recovery  
503 of species and ecosystems in the three MPAs.

504 The most notable effects of MPAs are an increase in abundance, average size and biomass  
505 inside protected areas ([Lester et al., 2009](#); [Giakoumi et al., 2017](#)). Due to the increased  
506 density inside the MPA, adults and juveniles may migrate to adjacent areas (spillover  
507 effect) and therefore, increase fisheries yields in adjacent areas ([Goñi et al., 2008](#); [Di  
508 Lorenzo et al., 2016](#)). Larger abundance and sizes would imply an increase in reproductive  
509 potential of species and eggs and larvae from recovered populations could then be exported  
510 to external unprotected locations, including adjacent ones ([Gell and Roberts, 2003](#);  
511 [Harrison et al., 2012](#)). In addition, MPAs could restore ecosystem functioning ([Cheng et al.,](#)

512 [2019](#)). Although some of our temporal results showed potential recoveries (some target  
513 species, increases of predators and declines of prey, and some ecological indicators), our  
514 study evidences that they are still far from what we would expect from the temporal  
515 protection effects of MPAs, highlighting an overall modest historical positive effect of  
516 protection on these MPAs. These results are in line with [Hereu Fina et al. \(2017\)](#), which  
517 have pointed out illegal fishing as one of the main reasons for a lack of recovery and have  
518 called for further enhance the enforcement in Mediterranean MPAs.

519 Overall the input data used was of acceptable quality compared to the distribution of  
520 pedigree values in other existing models ([Morissette, 2007](#); [Lassalle et al., 2014](#)), although  
521 they are among the lowest values in the Mediterranean Sea ([Corrales et al., 2015](#)). This  
522 could be due to the challenges of modelling small and local coastal areas, where specific  
523 data is a strong requirement. In fact, limitations of available data are a common concern for  
524 most Mediterranean MPAs ([Prato et al., 2016](#); [Vilas et al., Submitted](#)) despite there are  
525 monitoring programs inside these MPAs.

526 In general, there is a lack of biomass estimates for many functional groups, especially  
527 regarding benthic invertebrates. Obtaining realistic estimates of total catch (official and  
528 Illegal, Unregulated and Unreported (IUU)) is a challenging task worldwide ([Pauly and  
529 Zeller, 2016](#)) and even more in the Mediterranean Sea ([Coll et al., 2014](#); [Pauly et al., 2014](#)).  
530 In our study this challenge is higher due to the importance of recreational fishers and illegal  
531 fishing in coastal areas, specifically in MPA and surrounding areas (e.g., [Lloret et al.,  
532 2008b](#); [Ben Lamine et al., 2018](#)).

533 The ability of ecosystem models to replicate trends increases with data availability, quality,  
534 and the length of historical data provided ([Giron-Nava et al., 2017](#)). Within this context, in  
535 Cerbère-Banyuls a maximum of four data points of biomass per functional group were  
536 available for the fitting procedure. The Medes Islands MPA model had the longest available  
537 time series, and thus presented a better and more informative fit, while the Cap de Creus  
538 MPA model showed an intermediate situation. This limited the capability to calibrate and  
539 validate the models and to track ecosystem dynamic status. Therefore, our results highlight  
540 the need to further monitor MPAs within the Mediterranean Sea and should inform future  
541 scientific research objectives in the area.

542 Because of the caveats explained above, uncertainties in model outputs were high. Despite  
543 these limitations, the models presented in this study were developed using the best available  
544 data, including several sources of information including unpublished and *ad hoc* field data,  
545 expert knowledge and published data, and following the best practices in ecosystem  
546 modelling development ([Heymans et al., 2016](#)). These models represent a useful tool to  
547 integrate such amount of information into a coherent picture of the ecosystems and confirm  
548 the capability of EwE to evaluate protection effects within an ecological perspective in  
549 small MPAs even in cases of data limitations. As new information is generated, the present  
550 models can be updated and further developed, improving model quality and becoming  
551 increasingly valuable tools for evaluating management options in the three MPAs.

552 Human activities have been concentrated in coastal ecosystems, resulting in a major  
553 modification of these marine areas ([Halpern et al., 2015](#)). Since coastal areas account for a  
554 large amount of the global ecosystem services ([Costanza et al., 1997](#)), there is a need for  
555 the restoration of coastal marine ecosystems in order to ensure their ecological role. In  
556 addition, in recent decades, marine ecosystems have been increasingly impacted by other  
557 stressors, directly or indirectly induced by multiple anthropogenic activities ([Halpern et al.,  
558 2015](#)). For example, climate change and alien species are already heavily impacting  
559 Mediterranean ecosystems, especially in the Eastern Mediterranean Sea ([Lejeusne et al.,  
560 2010](#); [Katsanevakis et al., 2014](#)). Organisms and ecosystems already stressed by fishing are  
561 more vulnerable to these additional stressors ([Occhipinti-Ambrogi and Savini, 2003](#);  
562 [Poloczanska et al., 2016](#)). MPAs, when properly managed, have demonstrated to be an  
563 effective tool to protect target species and habitats to local stressors such as fishing ([Sala et  
564 al., 2017](#)), while their role in promoting resilience to regional and global stressors such  
565 biological invasions and climate change is debated ([Giakoumi and Pey, 2017](#); [Roberts et  
566 al., 2017](#); [Giakoumi et al., 2019](#)). In highly impacted and crowded areas like the  
567 Northwestern Mediterranean Sea, the establishment of well-designed networks of well-  
568 enforced MPA is necessary to achieve “clean, healthy and productive” oceans (Good  
569 Environmental Status) according to the Marine Strategy Framework Directive ([Fenberg et  
570 al., 2012](#)).

571 Within the EBM, ecosystem models and ecological forecasts have become an essential  
572 analytical and decision-making tool despite their limitations due to high uncertainties and  
573 complex ecosystem characteristics ([Link et al., 2012](#); [Collie et al., 2014](#); [Maris et al., 2017](#)).  
574 They have the potential to provide insights of possible future impacts on marine ecosystems  
575 and can offer guidance to decision-makers by evaluating the trade-off between different  
576 management units and identify those measures that have the potential to meet conservation  
577 objectives ([Fulton et al., 2015b](#); [Acosta et al., 2016](#)). This study is the baseline to develop  
578 future scenarios of alternative management options in order to maximize the impacts of  
579 MPAs to their surrounding areas and fisheries sustainability by alternative MPAs  
580 configurations. In addition, due to the vicinity of the three MPAs along a latitudinal  
581 gradient, this study is a part of a nested modelling approach with different geographic  
582 scales with the aim to assess the current effects of the actual MPA network and to perform  
583 simulations of alternative MPA network configuration ([Coll et al., 2019a](#)).

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#### 588 **Tables legends**

589 **Table 1.** Surface area (km<sup>2</sup>) covered by management units (MU) and year of creation of  
590 each Marine Protected Area (MPA) of Cerbère-Banyuls, Cap de Creus and Medes Islands.  
591 FPA= Fully protected Area, PPA= partially protected area, UPA= unprotected area flanking  
592 the MPA.

593 **Table 2.** Overall fitting procedure applied to the three MPA models following the  
594 methodology suggested by [Mackinson et al. \(2009\)](#) and [Heymans et al. \(2016\)](#).

595 **Table 3.** Results of the fitting procedure of the three MPA of Cerbère-Banyuls, Cap de  
596 Creus and Medes Islands. The table shows the statistically “best” model for each step. Vs =  
597 number of vulnerabilities estimated, PPsp = number of primary production spline points, k  
598 = number of parameters (Vs + PPsp), %IF = improved fit compared to the baseline AICc.  
599 The “best” models are highlighted in bold.

600 **Figure legends**

601 **Figure. 1.** The study area encompassing the three multi-zone MPAs in the Northwestern  
602 Mediterranean Sea with the fully protected areas (FPAs), partially protected areas (PPAs)  
603 and unprotected neighboring areas (UPAs).

604 **Figure. 2.** Flow diagram of Cerbère-Banyuls (a), Cap de Creus (b) and Medes Island (c)  
605 MPA models. The numbers identify the functional groups of the model (listed in SOM 2).  
606 The size of each circle is proportional to the biomass of the functional group. The thickness  
607 of the connecting lines is proportional to the magnitude of their trophic flows.

608 **Figure. 3.** Ecological indicators estimated for the three multi-zone MPA models of  
609 Cerbère-Banyuls, Cap de Creus and Medes Islands. Boxplot shows the distribution of  
610 values for an ecological indicator derived from the Monte Carlo routine while the dot  
611 represents the value of the indicator in the baseline Ecopath balanced model.

612 **Figure. 4.** Functional groups plotted against Keystone Index (KS) (Valls et al., 2015) and  
613 trophic level for Cerbère-Banyuls (a), Cap de Creus (b) and Medes (c) multi-zone MPA  
614 models. The numbers identify the functional groups of the model (listed in SOM 2). The  
615 size of the circles is proportional to the biomass of the functional group.

616 **Figure. 5.** Mixed Trophic Impact (MTI) analysis of the three MPA applied to the fisheries  
617 in a) Cerbère-Banyuls, b) Cap de Creus, and c) Medes Islands multi-zone MPA models.

618 **Figure. 6.** Predicted (solid lines) versus observed (dots) biomass ( $t \cdot km^{-2}$ ) values for target  
619 species of Cerbère-Banyuls, Cap de Creus and Medes Islands multi-zone MPAs models.  
620 Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo  
621 routine. Rho and p-values result from Spearman's rank correlation test.

622 **Figure. 7.** Temporal trends of ecological indicators of Cerbère-Banyuls, Cap de Creus and  
623 Medes Islands multi-zone MPAs models. Blue shadows represent the 5% and 95%  
624 percentiles obtained using the Monte Carlo routine. Rho and p-values result from  
625 Spearman's rank correlation test.

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## Modelling approach – Ecopath and Ecosim

The static Ecopath model provides a quantitative representation of the food web as a “snapshot” in terms of flows and biomasses for a defined period of time. To develop the food-web model we used functional groups, which consist of ontogenetic fractions of a species, single species or groups of species sharing common ecological and behavioral traits.

Ecopath is a mass-balanced model that is parameterized based on two master equations describing the production (Eq. 1) and consumption (Eq. 2) of each functional group (Christensen and Walters, 2004; Christensen et al., 2008).

Production = predation mortality + fishing mortality + other mortality + biomass accumulation + net migration (Eq. 1)

Consumption = production + respiration + unassimilated food (Eq. 2)

For each functional group, three of the four basic parameters (biomass ( $B$ ), production ( $P/B$ ) and consumption ( $Q/B$ ) rates, and ecotrophic efficiency ( $EE$ )) are required and the fourth is estimated. In addition, for each functional group the diet composition is required as well as the catch by fleet. A detailed explanation of the algorithms and equations of the approach and its main advantages and limitations are described in Christensen and Walters (2004) and Heymans et al. (2016).

Ecosim describes the temporal dynamics of species biomass and flows over time by accounting for changes in predation, consumption rate, fishing and the environment (Walters et al., 1997; Christensen and Walters, 2004). Ecosim uses a set of differential equations to describe biomass dynamics, expressed as:

$$\frac{dB_i}{dt} = \left(\frac{P}{Q}\right)_i \cdot \sum Q_{ji} - \sum Q_{ij} + I_i - (M_i + F_i + e_i) \cdot B_i \quad (\text{Eq. 3})$$

where  $dB_i/dt$  is the growth rate of group (i) during time t in terms of its biomass  $B_i$ ;  $(P/Q)_i$  is the net growth efficiency of group (i);  $M_i$  is the non-predation mortality rate;  $F_i$  is the fishing mortality rate;  $e_i$  is the emigration; and  $I_i$  is the immigration rate (Christensen and Walters, 2004).

Consumption rates ( $Q_{ij}$ ) are calculated based on the “foraging arena” theory, which divides the biomass of a prey into a vulnerable and a non-vulnerable fraction and the transfer rate or vulnerability between the two fractions determines the trophic flow between the predator and the prey. The vulnerability concept incorporates density-dependency and expresses how far a group is from its carrying capacity (Christensen and Walters, 2004; Christensen et al., 2008). Default values of vulnerability ( $v_{ij} = 2$ ) represents a mixed trophic flow, a low value ( $v_{ij} < 2$ ) indicates a “bottom-up” flow and a situation closer to carrying capacity, while a high value ( $v_{ij} > 2$ ) indicates a “top-down” flow and a situation further away from carrying capacity (Walters and Martell, 2004; Ahrens et al., 2012). For each predator-prey interaction, consumption rates are calculated as:

$$Q_{ij} = \frac{a_{ij} * v_{ij} * B_i * P_j * T_i * T_j * M_{ij} / D_j}{v_{ij} + v_{ij} * T_i * M_{ij} + a_{ij} * M_{ij} * P_i * T_j / D_j} * f(Env_{function}, t) \quad (\text{Eq. 4})$$

where  $a_{ij}$  is the rate of effective search for prey (i) by predator (j),  $T_i$  represents prey relative feeding time,  $T_j$  is the predator relative feeding time,  $B_i$  is prey biomass,  $P_j$  is predator abundance,  $M_{ij}$  is the mediation forcing effects, and  $D_j$  represents effects of handling time as a limit to consumption rate (Christensen et al., 2008; Ahrens et al., 2012). Environmental response functions ( $f(Env_{function}, t)$ ) can be used to account for external drivers that change over time, such as temperature. A detailed explanation of the algorithms and equations of the EwE approach are given in Christensen and Walters (2004) and Heymans et al. (2016).

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**Table 1.** Species composition, methods and references used to estimate the basic input parameters of the three MPA Ecopath models of Cerbère-Banyuls, Cap de Creus and Medes Islands.

Basic Input parameters	Method	Source
<b>1. Bottlenose dolphins (BOD): <i>Tursiops truncatus</i></b>		
<b>B</b>		(Forcada et al., 2004)
<b>P/B</b>	Life history table	(Barlow and Boveng, 1991)
<b>Q/B</b>	From modified energy requirement equation: $E = aW^{0.714}$	(Blanco et al., 2001; Hunter, 2006; Kastelein et al., 2003; Pauly et al., 1998; Pavan et al., 2013)
<b>D</b>		(Blanco et al., 2001)
<b>2. Striped dolphins (STD): <i>Stenella coeruleoalba</i></b>		
<b>B</b>		(Forcada and Hammond, 1998)
<b>P/B</b>	Life history table	(Barlow and Boveng, 1991)
<b>Q/B</b>	From modified energy requirement equation: $E = aW^{0.714}$	(Di-Méglio et al., 1996; Hunter, 2006; Marsili et al., 1997; Pauly et al., 1998; Pavan et al., 2013)
<b>D</b>		(Aznar et al., 2017)
<b>3. Endangered and pelagic seabirds (ENS): <i>Calonectris diomedea</i>, <i>Hydrobates pelagicus melitensis</i>, <i>Puffinus yelkouan</i>, <i>Puffinus mauretanicus</i></b>		
<b>B</b>		(Abelló and Oro, 1998; Amengual et al., 1999; Arcos et al., 2012; Arroyo et al., 2016; Bourgeois et al., 2011; du Rau et al., 2015; Gallo-Orsi, 2003; Genovart et al., 2016; Grémillet et al., 2014; Paracuellos and Jerez, 2003; Paracuellos and Nevado, 2003; Pettex et al., 2017; Thibault et al., 1996; Zotier et al., 1999)
<b>P/B</b>	Mortality (z) = $-\ln(\text{Survival rate})$	(Krebs, 1989)
<b>Q/B</b>	Yearly food intake (including breeding and no breeding seasons)/B	(Karpouzi et al., 2007; Paleczny, 2012)
<b>D</b>		(Arcos and Oro, 2002; Bourgeois et al., 2011; Granadeiro et al., 1998)
<b>4. Gulls and cormorants (GUC): <i>Larus audouinii</i>, <i>Larus genei</i>, <i>Larus melanocephalus</i>, <i>Larus michahellis</i>, <i>Phalacrocorax aristotelis</i>, <i>Phalacrocorax carbo</i></b>		
<b>B</b>		(Abelló and Oro, 1998; Amengual et al., 1999; Arcos et al., 2012; Arroyo et al., 2016; Bourgeois et al., 2011; du Rau et al., 2015; Gallo-Orsi, 2003; Genovart et al., 2016; Grémillet et al., 2014; Paracuellos and Jerez, 2003; Paracuellos and Nevado, 2003; Pettex et al., 2017; Thibault et al., 1996; Zotier et al., 1999)
<b>P/B</b>	Mortality (z) = $-\ln(\text{Survival rate})$	(Krebs, 1989)
<b>Q/B</b>	Yearly food intake (including breeding and no breeding seasons)/B	(Karpouzi et al., 2007; Paleczny, 2012)
<b>D</b>		(Goutner, 1994; Morat et al., 2014; Talmat-Chaouchi et al., 2014; Van Eerden and Munsterman, 1986)
<b>5. Terns (TER): <i>Sterna albifrons</i>, <i>Sterna caspia</i>, <i>Sterna hirundo</i>, <i>Sterna nilotica</i>, <i>Sterna sandvicensis</i></b>		
<b>B</b>		(Abelló and Oro, 1998; Amengual et al., 1999; Arcos et al., 2012;

		Arroyo et al., 2016; Bourgeois et al., 2011; du Rau et al., 2015; Gallo-Orsi, 2003; Genovart et al., 2016; Grémillet et al., 2014; Paracuellos and Jerez, 2003; Paracuellos and Nevado, 2003; Pettex et al., 2017; Thibault et al., 1996; Zotier et al., 1999)
<b>P/B</b>	Mortality (z)= -ln(Survival rate)	(Krebs, 1989)
<b>Q/B</b>	Yearly food intake (including breeding and no breeding seasons)/B	(Karpouzi et al., 2007; Paleczny, 2012)
<b>D</b>		(Dies et al., 2005)
<b>6. Loggerhead turtle (LGT): <i>Caretta caretta</i></b>		
<b>B</b>		(de Segura et al., 2006; Lauriano et al., 2011)
<b>P/B</b>	Mortality (z)= -ln(Survival rate)	(Casale et al., 2015; Casale and Heppell, 2016; Krebs, 1989)
<b>Q/B</b>	Yearly food intake/B	(Hatase and Tsukamoto, 2008)
<b>D</b>		(Cardona et al., 2012a)
<b>7. Non-commercial large pelagic fish (NLP): <i>Cetorhinus maximus</i>, <i>Mobula mobular</i>, <i>Mola mola</i></b>		
<b>B</b>		(di Sciara et al., 2015; Mancusi et al., 2005) and International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Fortuna et al., 2014; Parker and Stott, 1965; Powell, 2003)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); a and b parameters from (Coull et al., 1989; Matthews and Parker, 1950)
<b>D</b>		(Grémillet et al., 2017)
<b>8. Other large pelagic fish (OLP): <i>Coryphaena hippurus</i>, <i>Lichia amia</i>, <i>Seriola dumerili</i></b>		
<b>B</b>		(Morales-Nin and Azevedo, 2004) and International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Gatt et al., 2015; Morey et al., 2003)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); a and b parameters from (Gatt et al., 2015; Morey et al., 2003)
<b>D</b>		(Cardona et al., 2012b; Coetzee, 1982)
<b>9. Mackerels (MCK): <i>Scomber scombrus</i>, <i>Scomber colias</i>, <i>Scomber spp.</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Campillo, 1992; Velasco et al., 2011)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); a and b parameters from (Merella et al., 1997; Torres et al., 2012)
<b>D</b>		(Apostolidis and Stergiou, 2014; Cardona et al., 2012b)
<b>10. Horse mackerels (HRM): <i>Trachurus trachurus</i>, <i>Trachurus mediterraneus</i>, <i>Trachurus picturatus</i></b>		

<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Vasconcelos et al., 2006; Zupa et al., 2006)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Torres et al., 2012)
<b>D</b>		(Buttay, 2009; Cardona et al., 2012b; Cresson et al., 2014)
<b>11. Other medium pelagic fish (OMP): <i>Alosa fallax</i>, <i>Caranx rhonchus</i>, <i>Pomatomus saltatrix</i>, <i>Sphyræna sphyraena</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Allam et al., 2004; Backus, 1962; Coull et al., 1989; Douchement, 1981; Quignard and Douchement, 1991)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Quignard and Douchement, 1991; Stergiou and Moutopoulos, 2001a; Torres et al., 2012)
<b>D</b>		(Dhieb et al., 2001; Kalogirou et al., 2012; Karachle, 2008; Mostarda et al., 2007)
<b>12. European sardine (ESA): <i>Sardina pilchardus</i></b>		
<b>B</b>		Mediterranean International Acoustic Survey (MEDIAS) (MEDIAS, 2012), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (De Ranieri, 2011)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Demirhindi, 1961)
<b>13. European anchovy (EAN): <i>Engraulis encrasicolus</i></b>		
<b>B</b>		Mediterranean International Acoustic Survey (MEDIAS) (MEDIAS, 2012), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Colloca et al., 2013)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Tudela and Palomera, 1997)
<b>14. Other small pelagic fish (OSP): <i>Cubiceps gracilis</i>, <i>Spicara flexuosa</i>, <i>Spicara maena</i>, <i>Spicara smaris</i>, <i>Spicara spp.</i>, <i>Sprattus sprattus</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean

		(MEDITS) (Bertrand et al., 2002), and scaled with Aquamaps (Kaschner et al., 2016)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Apostolidis and Stergiou, 2014; Passelaigne, 1974; Zavodnik, 1969)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997; Sinovčić et al., 2004)
<b>D</b>		(Cresson et al., 2014; Khoury, 1987; Tičina et al., 2000)
<b>15. Anglerfish (ANG): <i>Lophius budegassa</i>, <i>Lophius piscatorius</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Scientific and for Fisheries (STECF), 2013; Tsimenidis, 1980)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Apostolidis and Stergiou, 2014; Merella et al., 1997)
<b>D</b>		(López et al., 2016; Macpherson, 1981)
<b>16. European conger (ECO): <i>Conger conger</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Campillo, 1992)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)
<b>D</b>		(Bell, 1983)
<b>17. European hake (EHK): <i>Merluccius merluccius</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Campillo, 1992)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Aldebert and Recasens, 1995)
<b>D</b>		(Mellon-Duval et al., 2017)
<b>18. Poor cod (PCO): <i>Trisopterus capelanus</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Planas and Vives, 1952)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)

<b>D</b>		(Biagi et al., 1992)
<b>19. Common pandora (CPA): <i>Pagellus erythrinus</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>B Medes</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Planas and Vives, 1952)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Cresson et al., 2014)
<b>20. Sparidae (SPA): <i>Boops boops</i>, <i>Diplodus annularis</i>, <i>Diplodus cervinus</i>, <i>Diplodus puntazzo</i>, <i>Oblada melanura</i>, <i>Pagrus pagrus</i>, <i>Sparus aurata</i>, <i>Spondylisoma cantharus</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco 2017) and k parameter from (Derbal and Kara, 2006; Froggia, 1984; Girardin, 1978; Quignard, 1986; Vassilopoulou, 1989)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992; Merella et al., 1997)
<b>D</b>		(Bell, 1983; Benchalel et al., 2010; Chessa et al., 2004; Cresson et al., 2014; Labropoulou and Papadopoulou-Smith, 1999; Lenfant and Olive, 1998; Rosocchi and Nouaze, 1985; Sala and Ballesteros, 1997)
<b>21. White seabream (WSE): <i>Diplodus sargus</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco 2017) and k parameter from (Man-Wai and Quignard, 1982)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Cresson et al., 2014)
<b>22. Common two-banded seabream: <i>Diplodus vulgaris</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)

<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco 2017) and k parameter from (Girardin, 1978)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Bell, 1983)
<b>23. Common dentex (DEN): <i>Dentex dentex</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco 2017) and k parameter from (Morales-Nin and Moranta, 1997)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morales-Nin and Moranta, 1997)
<b>D</b>		(Morales-Nin and Moranta, 1997)
<b>24. Red scorpionfish (RSC): <i>Scorpaena scrofa</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Bradai and Bouain, 1988)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003) and $L_{\max}$ from UVC (Di Franco, 2017)
<b>D</b>		(Bell, 1983)
<b>25. Scorpaenidae (SCO): <i>Chelidonichthys cuculus</i>, <i>Chelidonichthys obscurus</i>, <i>Eutrigla gunardus</i>, <i>Lepidotrigla cavillona</i>, <i>Scorpaena elongate</i>, <i>Scorpaena notata</i>, <i>Scorpaena porcus</i>, <i>Scorpaena spp.</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013) and International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Bradai and Bouain, 1988; Ordines et al., 2009)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003)
<b>D</b>		(Bell, 1983; Colloca et al., 1994; Labropoulou and Plaitis, 1997;

		Moreno-Amich, 1994; Ordines et al., 2012)
<b>26. Groupers: <i>Epinephelus marginatus</i></b>		
<b>B Banyuls</b>		Underwater Visual Census (UVC) (Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Bouchereau et al., 1999; Girardin, 1978)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly 1990); <i>a</i> and <i>b</i> parameters from (Morey et al. 2003)
<b>D</b>		(Linde et al., 2004)
<b>27. Brown meagre: <i>Sciaena umbra</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Hereu Fina and Quintana Pou, 2012; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Chauvet, 1991)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003)
<b>D</b>		(Karachle, 2008)
<b>28. Labridae and serranidae (LAS): <i>Coris julis</i>, <i>Labrus merula</i>, <i>Labrus viridis</i>, <i>Serranus cabrilla</i>, <i>Serranus scriba</i>, <i>Symphodus cinereus</i>, <i>Symphodus doderleini</i>, <i>Symphodus mediterraneus</i>, <i>Symphodus ocellatus</i>, <i>Symphodus roissali</i>, <i>Symphodus rostratus</i>, <i>Symphodus spp.</i>, <i>Symphodus tinca</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Cheung et al., 2013; Froese and Binohlan, 2003; Gordoia et al., 2000; Papaconstantinou et al., 1994)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Gordoia et al., 2000; Merella et al., 1997; Morey et al., 2003; Stergiou and Moutopoulos, 2001b; Valle et al., 2003)
<b>D</b>		(Arculeo et al., 1993; Bell, 1983; Kabasakal, 2001)
<b>29. Flatfish (FLA): <i>Arnoglossus thori</i>, <i>Arnoglossus laterna</i>, <i>Buglossidium luteum</i>, <i>Citharus linguatula</i>, <i>Lepidorhombus whiffiagonis</i>, <i>Scophthalmus rhombus</i>, <i>Solea solea</i>, <i>Symphurus nigrescens</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Deniel, 1990; García-Rodríguez and Acón, 2000; Giovanardi

	Mortality (z)= F+M	and Piccinetti, 1984; Landa et al., 1996; Stergiou and Politou, 1995; Vianet et al., 1989)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997; Morey et al., 2003; Vianet et al., 1989)
<b>D</b>		(Bell, 1983; Cresson et al., 2014; Darnaude et al., 2004; De Juan et al., 2007; Fanelli et al., 2009; Macpherson, 1981; Morte et al., 1999; Pellegrini and Barghigiani, 1989; Wyche and Shackley, 1986)
<b>30. Other commercial medium demersal fish (CMD): <i>Dicentratus labrax</i>, <i>Phycis phycis</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(Hereu Fina et al., 2017)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and <i>k</i> parameter from (Morey et al., 2003; Nony, 1983; Stergiou and Politou, 1995)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Cresson et al., 2014; Macpherson, 1981)
<b>31. No commercial medium demersal fish (NMD): <i>Muraena helena</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and <i>k</i> parameter from (Matić-Skoko et al., 2014)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003)
<b>D</b>		(Sallami et al., 2014)
<b>32. Salema (SAL): <i>Sarpa salpa</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and <i>k</i> parameter from (Verlaque, 1985)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003)
<b>D</b>		(Verlaque, 1985)
<b>33. Mugilidae (MUG): <i>Chelon labrosus</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)



<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Morey et al., 2003)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Campillo, 1992)
<b>D</b>		(Blanco et al., 2003)
<b>34. Red mullet (RMU): <i>Mullus barbatus</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>B Medes</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Levi et al., 1992)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)
<b>D</b>		(Esposito et al., 2014)
<b>35. Surmullet (SUR): <i>Mullus surmuletus</i></b>		
<b>B Banyuls</b>		(Claudet, 2013)
<b>B Cap de Creus</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>B Medes</b>		(García-Rubies and i Limousin, 1990; Macpherson et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Campillo, 1992)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)
<b>D</b>		(Bell, 1983)
<b>36. No commercial small demersal fish (NSD): <i>Blennidae, Chromis chromis, Gobiis sp.</i></b>		
<b>B</b>		(García-Rubies and i Limousin, 1990)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\max}$ from UVC (Di Franco, 2017) and k parameter from (Maiorano et al., 2010; Merella et al., 1997; Morey et al., 2003; Sangun et al., 2007)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045$ $\cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Lamprakis et al., 2003; Morey et al., 2003)
<b>D</b>		(Bell, 1983; Karachle, 2008; Pinnegar et al., 2000; Zander and Berg, 1984)
<b>37. Small-spotted catshark (SPC): <i>Scyliorhinus canicula</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Županovic, 1961)

<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)
<b>D</b>		(Macpherson, 1981)
<b>38. Rays and skates (RSK): <i>Raja asterias</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Abdel-Aziz, 1992)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Merella et al., 1997)
<b>D</b>		(Navarro et al., 2013)
<b>39. Torpedos (TOR): <i>Torpedo marmorata</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>	$\log M = -0.0066 - 0.279 \cdot \log L_{\infty} + 0.6543 \cdot \log k + 0.4634 \cdot \log T$ Mortality (z)= F+M	Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Duman and Baştusta, 2013)
<b>Q/B</b>	$\log Q/B = 6.37 - 1.5045 \cdot \log T' - 0.168 \cdot \log W_{\infty} + 0.1399 \cdot Pf + 0.2765 \cdot HD$	Empirical equation from (Pauly, 1990); <i>a</i> and <i>b</i> parameters from (Morey et al., 2003)
<b>D</b>		(Barría et al., 2015)
<b>40. Coastal benthic cephalopods (CBC): <i>Octopus vulgaris</i>, <i>Sepia officinalis</i></b>		
<b>B</b>		(Valls et al., 2012)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>	$Q/B = 0.0683 + 0.0474(W)$	(Rodhouse and Nigmatullin, 1996); W from (Palomera et al., 2015)
<b>D</b>		(Castro and Guerra, 1990; Quetglas et al., 1998)
<b>C</b>		(Ivanoff et al., 2010; Prats, 2016)
<b>41. Benthopelagic cephalopods (BPC): <i>Alloteuthis media</i>, <i>Alloteuthis spp.</i>, <i>Alloteuthis subulata</i>, <i>Illex coindetii</i>, <i>Loligo spp.</i>, <i>Loligo vulgaris</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>	$Q/B = 0.0683 + 0.0474(W)$	(Rodhouse and Nigmatullin, 1996); W from (Palomera et al., 2015)
<b>D</b>		(Martínez-Baena et al., 2016; Rosas-Luis and Sánchez, 2015; Valls et al., 2015)
<b>42. Other benthic cephalopods (BCE): <i>Eledone cirrhosa</i>, <i>Eledone moschata</i>, <i>Sepia elegans</i>, <i>Sepia spp.</i>, <i>Sepietta oweniana</i>, <i>Sepiola intermedia</i>, <i>Sepiola spp.</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>	$Q/B = 0.0683 + 0.0474(W)$	(Rodhouse and Nigmatullin, 1996); W from (Palomera et al.,

		2015)
<b>D</b>		(Castro and Guerra, 1990; Šifner and Vrgoč, 2009; Valls, 2017)
<b>43. Bivalves (BIV): <i>Bivalvia</i> spp, <i>Pinna nobilis</i></b>		
<b>B Banyuls</b>		(De Juan et al., 2011; Preuvost, 2011)
<b>B Cap de Creus</b>		<i>estimated</i>
<b>B Medes</b>		<i>estimated</i>
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Sarà et al., 2003)
<b>44. Gastropods (GAS): <i>Gastropoda</i> spp</b>		
<b>B Banyuls</b>		(De Juan et al., 2011)
<b>B Cap de Creus</b>		<i>estimated</i>
<b>B Medes</b>		<i>estimated</i>
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Heymans et al., 2016)
<b>D</b>		(Mazzella and Russo, 1989)
<b>45. European lobster (LOB): <i>Palinurus elephas</i></b>		
<b>B Banyuls</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>B Cap de Creus</b>		(Hereu Fina and Quintana Pou, 2012)
<b>B Medes</b>		(Hereu Fina and Quintana Pou, 2012)
<b>P/B</b>		(Heymans et al., 2016)
<b>Q/B</b>		Empirical equation from (Pauly, 1980); $L_{\infty}$ and k parameter from (Sardà, 2000)
<b>D</b>		(Goñi et al., 2001)
<b>46. Other commercial decapods (OCD): <i>Maja squinado</i>, <i>Squilla mantis</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Bernárdez et al., 2000; Froglija, 1989)
<b>47. Non-commercial decapods (NDC): <i>Dardanus arrosor</i>, <i>Dromia personata</i>, <i>Ethusa mascarone</i>, <i>Goneplax rhomboides</i>, <i>Liocarcinus depurator</i>, <i>Maja brachydactyla</i>, <i>Medorippe lanata</i>, <i>Pagurus excavatus</i>, <i>Pagurus</i> spp, <i>Pilumnus</i> spp, <i>Pisa armata</i></b>		
<b>B</b>		<i>estimated</i>
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Abelló and Cartes, 1987)
<b>48. Purple sea urchin (PSU): <i>Paracentrotus lividus</i></b>		
<b>B Banyuls</b>		(Schvartz and Labbe, 2012)
<b>B Cap de Creus</b>		(Hereu Fina and Quintana Pou, 2012)

<b>B Medes</b>		(Hereu Fina and Quintana Pou, 2012)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Murillo-Navarro and Jiménez-Guirado, 2012)
<b>49. Other sea urchins (OSU): <i>Arbacia lixula</i>, <i>Sphaerechinus granularis</i></b>		
<b>B Banyuls</b>		(Schvartz and Labbe, 2012)
<b>B Cap de Creus</b>		(Hereu Fina and Quintana Pou, 2012)
<b>B Medes</b>		(Hereu Fina and Quintana Pou, 2012)
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Rodríguez, 1972)
<b>50. Sea cucumbers (SCU): <i>Holothuria spp</i></b>		
<b>B</b>		<i>estimated</i>
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Coulon and Jangoux, 1993)
<b>51. Other macro-benthos (OMB): <i>Actinaria sp</i>, <i>Ascidacea sp.</i>, <i>Asteroidea sp.</i>, <i>Ophiuroidea sp.</i>, <i>Polychaeta sp.</i>, <i>Porifera sp</i></b>		
<b>B</b>		<i>estimated</i>
<b>P/B</b>		(Brey, 1990)
<b>Q/B</b>		(Pauly et al., 1993b)
<b>D</b>		(Rodríguez, 1972)
<b>52. Jellyfish (JLL): <i>Aequorea forskalea</i>, <i>Aurelia aurita</i>, <i>Chrysaora hysoscella</i>, <i>Hydrozoa</i>, <i>Pelagia noctiluca</i>, <i>Rhizostoma pulmo</i>, <i>Pleurobrachia pileus</i></b>		
<b>B</b>		International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>		(Coll, 2018)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Sabatés et al., 2010)
<b>53. Salps (SAL): <i>Thalia democratic</i>, <i>Salpa fusiformis</i>, <i>Salpa maxima</i>, <i>Pyrosoma</i>, <i>Pyrosoma atlanticum</i></b>		
<b>B</b>		(Pascual Torner, 2016) and International Bottom Trawl Survey in the Mediterranean (MEDITS) (Bertrand et al., 2002)
<b>P/B</b>		(Coll, 2018)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Pascual Torner, 2016)
<b>54. Red coral (RCO): <i>Corallium rubrum</i></b>		
<b>B Banyuls</b>		(Septentrion Environement, 2015)
<b>B Cap de Creus</b>		(Coll, 2018)
<b>B Medes</b>		(Coll, 2018)
<b>P/B</b>		(Coll, 2018)
<b>Q/B</b>		(Coll, 2018)

<b>D</b>		(Tsounis et al., 2006)
<b>55. Other corals and gorgonians (CGO):</b> <i>Bryozoa</i> , <i>Alcyonium acaule</i> , <i>Alcyonium palmatum</i> , <i>Alcyonium spp.</i> , <i>Eunicella cavolini</i> , <i>Eunicella filiformis</i> , <i>Eunicella singularis</i>		
<b>B Banyuls</b>		(Schvartz and Labbe, 2012)
<b>B Cap de Creus</b>		(Coll, 2018)
<b>B Medes</b>		(Coll, 2018)
<b>P/B</b>		(Coll, 2018)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Rossi, 2001)
<b>56. Macro zooplankton (MAZ)</b>		
<b>B</b>		(Gaudy et al., 2003)
<b>P/B</b>		(Heymans et al., 2016)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Båmstedt and Karlson, 1998)
<b>57. Meso and micro zooplankton (MMZ)</b>		
<b>B</b>		(Gaudy et al., 2003)
<b>P/B</b>		(Coll, 2018)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Calbet et al., 2002)
<b>58. Suprabenthos (SUB)</b>		
<b>B</b>		<i>estimated</i>
<b>P/B</b>		(Corbera et al., 2013)
<b>Q/B</b>		(Coll, 2018)
<b>D</b>		(Cartes et al., 2001)
<b>59. Mediterranean seagrass (POS):</b> <i>Posidonia oceanica</i>		
<b>B Banyuls</b>		(Ferrari, 2006)
<b>B Cap de Creus</b>		(Hereu Fina and Quintana Pou, 2012)
<b>B Medes</b>		(Hereu Fina and Quintana Pou, 2012)
<b>P/B</b>		(Coll, 2018)
<b>60. Other seagrasses:</b>		
<b>B Banyuls</b>		<i>nonpresent</i>
<b>B Cap de Creus</b>		(EMODnet, 2014)
<b>B Medes</b>		(EMODnet, 2014)
<b>P/B</b>		(Coll, 2018)
<b>61. Erected algae (EAL):</b>		
<b>B Banyuls</b>		<i>nonpresent</i>
<b>B Cap de Creus</b>		(Giannoulaki, 2013)
<b>B Medes</b>		(Giannoulaki, 2013)
<b>P/B</b>		(Coll, 2018)
<b>62. Seaweeds (SEA):</b>		

<b>B Banyuls</b>		(Mellado, 2006)
<b>B Cap de Creus</b>		(Giannoulaki, 2013)
<b>B Medes</b>		(Giannoulaki, 2013)
<b>P/B</b>		(Coll, 2018)
<b>63. Small phytoplankton (SMP):</b>		
<b>B</b>		(Joint Research Centre, 2017a)
<b>P/B</b>		(Joint Research Centre, 2017b)
<b>64. Large phytoplankton (LPH):</b>		
<b>B</b>		(Joint Research Centre, 2017a)
<b>P/B</b>		(Joint Research Centre, 2017b)
<b>65. Detritus (DET):</b>		
<b>B</b>		(Pauly et al., 1993a)
<b>66. Discards and by-catch (DIB) – (except FPA)</b>		
<b>B Banyuls</b>		(Prats, 2016)
<b>B Cap de Creus</b>		(Tudó, 2017)
<b>B Medes</b>		(Tudó, 2017)
<b>67. Recreational fishery (RFI) – (except FPA)</b>		
<b>C Banyuls</b>		(Ivanoff et al., 2010)
<b>C Cap de Creus</b>		(Sacanell, 2012)
<b>C Medes</b>		(Sacanell, 2012)
<b>68. Artisanal fishery (AFI) – (except FPA)</b>		
<b>C Banyuls</b>		(Prats, 2016)
<b>C Cap de Creus</b>		(Tudó, 2017)
<b>C Medes</b>		(Tudó, 2017)

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**Table 1.** Values used to calculate import estimations in the diet of the functional groups based on their behaviour and size for the management units models of the three MPAs models in the NW Mediterranean Sea (Vilas et al., in preparation). These values were slightly decreased in the three models of this study taking into account expert knowledge of the project and the fact that the modelled areas are higher (whole MPA's).

<b>Behavior</b>	<b>value</b>
<b>demersal</b>	0.2
<b>benthopelagic</b>	0.5
<b>pelagic</b>	0.8
<b>sessile</b>	0
<b>benthic</b>	0.1
<b>passive pelagic</b>	0.3

<b>size</b>	<b>value</b>
<b>small</b>	0.2
<b>medium</b>	0.5
<b>large</b>	0.8

**Table 2.** Initial (in bold) and estimated input data of the three MPA models in the NW Mediterranean Sea (Ba = Cerbère-Banyuls; C = Cap de Creus; M = Medes Islands). B = final biomass ( $t \cdot km^{-2}$ ); P/B = production/biomass ( $year^{-1}$ ); Q/B = consumption/biomass ( $year^{-1}$ ); EE = ecotrophic efficiency; P/Q = production/consumption ratio; landings and discards ( $t \cdot km^{-2} \cdot year^{-1}$ ).

Functional group	B			P/B			Q/B			EE			P/Q			Landings			Discards		
	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M
1 Bottlenose dolphins	<b>0.01</b>	<b>0.01</b>	<b>0.01</b>	<b>0.07</b>	<b>0.07</b>	<b>0.03</b>	<b>12.81</b>	<b>12.81</b>	<b>12.81</b>	0.00	0.00	0.00	0.01	0.01	0.00	-	-	-	-	-	-
2 Striped dolphins	<b>0.01</b>	<b>0.01</b>	<b>0.01</b>	<b>0.03</b>	<b>0.03</b>	<b>0.03</b>	<b>14.11</b>	<b>14.11</b>	<b>14.11</b>	0.00	0.00	0.00	0.00	0.00	0.00	-	-	-	-	-	-
3 Endangered and pelagic seabirds	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.41</b>	<b>0.41</b>	<b>0.41</b>	<b>73.75</b>	<b>73.75</b>	<b>73.75</b>	0.00	0.00	0.00	0.01	0.01	0.01	-	-	-	-	-	-
4 Gulls and cormorants	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.31</b>	<b>0.31</b>	<b>0.31</b>	<b>70.80</b>	<b>70.80</b>	<b>70.80</b>	0.00	0.00	0.00	0.00	0.00	0.00	-	-	-	-	-	-
5 Terns	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.29</b>	<b>0.29</b>	<b>0.29</b>	<b>85.40</b>	<b>85.40</b>	<b>85.40</b>	0.00	0.00	0.00	0.00	0.00	0.00	-	-	-	-	-	-
6 Loggerhead turtles	<b>0.01</b>	<b>0.01</b>	<b>0.01</b>	<b>0.20</b>	<b>0.20</b>	<b>0.20</b>	<b>2.52</b>	<b>2.52</b>	<b>2.52</b>	0.00	0.00	0.00	0.08	0.08	0.08	-	-	-	-	-	-
7 Non-commercial large pelagic fishes	<b>0.01</b>	<b>0.01</b>	<b>0.01</b>	<b>0.14</b>	<b>0.14</b>	<b>0.14</b>	<b>1.63</b>	<b>1.63</b>	<b>1.63</b>	0.00	0.00	0.00	0.08	0.08	0.08	-	-	-	-	-	-
8 Other large pelagic fishes	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	0.41	0.50	0.41	<b>4.13</b>	<b>4.96</b>	<b>4.13</b>	0.02	0.89	0.45	<b>0.10</b>	<b>0.10</b>	<b>0.10</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	-	-
9 Mackerels	<b>0.00</b>	<b>0.00</b>	<b>0.01</b>	0.88	0.88	<b>0.86</b>	<b>7.31</b>	<b>7.32</b>	<b>7.31</b>	0.98	0.34	0.58	<b>0.12</b>	<b>0.12</b>	0.12	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	<b>0.00</b>	<b>0.00</b>
10 Horse mackerels	<b>0.01</b>	<b>0.02</b>	<b>0.00</b>	0.85	0.86	0.87	<b>7.10</b>	<b>7.19</b>	<b>7.26</b>	0.75	0.49	0.73	<b>0.12</b>	<b>0.12</b>	<b>0.12</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>
11 Other medium pelagic fishes	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	0.71	<b>0.77</b>	<b>0.69</b>	<b>6.47</b>	<b>5.19</b>	<b>6.47</b>	0.46	0.10	0.60	<b>0.11</b>	0.15	0.11	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	-	-
12 European sardine	<b>0.03</b>	<b>0.05</b>	<b>0.06</b>	1.63	1.64	1.64	<b>10.87</b>	<b>10.91</b>	<b>10.91</b>	0.99	0.71	0.53	<b>0.15</b>	<b>0.15</b>	<b>0.15</b>	-	-	-	-	-	-
13 European anchovy	<b>0.02</b>	<b>0.01</b>	<b>0.01</b>	1.63	1.63	1.63	<b>10.84</b>	<b>10.88</b>	<b>10.88</b>	0.97	0.91	0.63	<b>0.15</b>	<b>0.15</b>	<b>0.15</b>	-	-	-	-	-	-
14 Round sardinella	-	<b>0.00</b>	<b>0.00</b>	-	1.38	1.38	-	<b>9.18</b>	<b>9.18</b>	-	0.85	0.77	-	<b>0.15</b>	<b>0.15</b>	-	-	-	-	-	-
15 Other small pelagic fish	<b>0.07</b>	<b>0.04</b>	<b>0.03</b>	<b>1.21</b>	<b>1.31</b>	<b>1.30</b>	<b>9.85</b>	<b>12.04</b>	<b>9.85</b>	0.85	0.96	0.77	0.12	0.11	0.13	<b>0.00</b>	<b>0.00</b>	-	-	<b>0.00</b>	-
16 Anglerfish	<b>0.05</b>	<b>0.01</b>	<b>0.01</b>	0.47	<b>0.58</b>	<b>0.58</b>	<b>4.30</b>	<b>4.30</b>	<b>4.30</b>	0.34	0.64	0.63	<b>0.11</b>	0.13	0.13	<b>0.01</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	-
17 European conger	<b>0.02</b>	<b>0.01</b>	<b>0.00</b>	0.56	<b>0.57</b>	<b>0.64</b>	<b>4.63</b>	<b>4.63</b>	<b>4.63</b>	0.16	0.63	0.74	<b>0.12</b>	0.12	0.14	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	<b>0.00</b>	<b>0.00</b>
18 European hake	<b>0.03</b>	<b>0.02</b>	<b>0.04</b>	0.67	<b>0.71</b>	<b>0.59</b>	<b>5.59</b>	<b>5.62</b>	<b>5.62</b>	0.43	0.74	0.63	<b>0.12</b>	0.13	0.11	<b>0.00</b>	<b>0.01</b>	<b>0.01</b>	<b>0.00</b>	-	-
19 Poor cod	<b>0.02</b>	<b>0.01</b>	<b>0.01</b>	1.16	1.16	1.16	<b>9.65</b>	<b>9.69</b>	<b>9.69</b>	0.53	0.90	0.73	<b>0.12</b>	<b>0.12</b>	<b>0.12</b>	<b>0.00</b>	<b>0.00</b>	<b>0.00</b>	-	<b>0.00</b>	<b>0.00</b>

20	Common pandora	0.11	0.02	0.05	0.75	0.88	0.82	6.27	6.27	6.27	0.88	0.90	0.91	0.12	0.14	0.13	0.00	0.01	0.01	0.00	0.00	0.00
21	Sparidae	18.27	3.70	10.37	1.08	1.03	1.03	8.99	8.58	8.60	0.83	0.54	0.41	0.12	0.12	0.12	0.01	0.03	0.03	0.00	0.00	0.00
22	White seabream	2.06	5.32	5.95	0.86	0.96	0.86	7.18	8.01	7.18	0.99	0.19	0.13	0.12	0.12	0.12	0.00	0.01	0.00	-	0.00	0.00
23	Common two-banded seabream	4.17	4.70	5.31	0.84	0.96	0.88	7.02	7.96	7.32	0.82	0.24	0.14	0.12	0.12	0.12	0.00	0.00	0.00	-	-	-
24	Common dentex	2.24	0.99	0.74	0.58	0.62	0.62	5.26	5.13	5.13	0.25	0.06	0.06	0.11	0.12	0.12	0.00	0.01	0.00	0.00	0.00	0.00
25	Red scorpionfish	2.17	0.74	0.75	0.71	0.72	0.71	5.93	6.01	5.88	0.69	0.23	0.19	0.12	0.12	0.12	0.00	0.01	0.00	0.00	0.00	0.00
26	Scorpaenidae	0.27	0.20	0.27	0.85	0.87	0.95	7.06	7.00	7.06	0.99	0.65	0.46	0.12	0.12	0.13	0.00	0.00	0.03	0.00	0.00	0.00
27	Groupers	12.53	2.10	1.69	0.18	0.23	0.22	3.54	3.91	3.91	0.00	0.01	0.00	0.05	0.06	0.06	0.00	0.00	0.00	-	-	0.00
28	Brown meagre	3.47	0.26	0.24	0.78	0.76	0.76	6.48	6.35	6.35	0.16	0.29	0.37	0.12	0.12	0.12	0.00	0.00	0.00	-	0.00	-
29	Labridae and serranidae	7.87	4.59	12.86	1.44	1.30	1.24	10.29	10.84	10.29	0.91	0.65	0.39	0.14	0.12	0.12	0.01	0.00	0.00	-	0.00	-
30	Flatfishes	0.07	0.02	0.02	1.31	1.17	1.35	10.94	10.37	9.49	0.42	0.66	0.92	0.12	0.11	0.14	0.00	0.00	0.01	0.00	0.00	0.00
31	Other commercial medium demersal fish	11.48	0.27	0.21	0.53	0.58	0.48	5.25	5.25	5.25	0.05	0.60	0.70	0.10	0.11	0.09	0.00	0.01	0.01	-	0.00	0.00
32	No commercial medium demersal fish	5.38	0.12	0.21	0.47	0.48	0.49	4.70	4.70	4.70	0.05	0.49	0.25	0.10	0.10	0.11	0.00	0.00	0.00	-	0.00	0.00
33	Salema	35.43	7.65	33.28	1.40	1.34	1.40	9.32	8.90	9.32	0.17	0.10	0.02	0.15	0.15	0.15	0.00	0.00	0.00	-	0.00	0.00
34	Mugilidae	16.98	0.81	2.95	0.91	1.00	0.88	6.07	6.66	5.84	0.28	0.21	0.05	0.15	0.15	0.15	0.00	0.01	0.00	-	0.00	0.00
35	Red mullet	0.30	0.02	0.01	1.05	1.05	1.07	7.47	7.47	7.47	0.81	0.80	0.88	0.14	0.14	0.14	0.00	0.01	0.00	-	0.00	0.00
36	Surmulet	0.75	0.06	0.84	0.92	0.98	0.95	6.58	6.58	6.58	0.80	0.97	0.10	0.14	0.15	0.14	0.00	0.00	0.01	-	0.00	0.00
37	No commercial small demersal fish	3.35	1.20	3.35	1.42	1.35	1.34	11.85	9.89	9.89	0.94	0.86	0.51	0.12	0.14	0.14	0.00	0.00	-	-	-	-
38	Small-spotted catshark	0.01	0.04	0.00	0.68	0.65	0.65	6.20	6.20	6.20	0.26	0.02	0.26	0.11	0.10	0.10	0.00	0.00	0.00	-	0.00	-
39	Rays and skates	0.03	0.04	0.03	0.71	0.68	0.79	5.94	5.97	5.97	0.03	0.02	0.16	0.12	0.11	0.13	0.00	0.00	0.00	-	0.00	0.00
40	Torpedos	0.01	0.01	0.00	0.51	0.50	0.50	4.98	5.00	5.00	0.14	0.01	0.04	0.10	0.10	0.10	0.00	-	-	0.00	-	-
41	Coastal benthic cephalopods	3.00	0.45	0.23	1.57	1.61	1.57	11.01	11.01	11.01	0.89	0.80	0.91	0.14	0.15	0.14	0.00	0.08	0.01	-	0.01	0.00
42	Benthopelagic cephalopods	0.07	0.01	0.02	1.72	1.72	1.72	16.53	16.53	16.53	0.29	0.83	0.80	0.10	0.10	0.10	0.00	0.00	0.00	-	0.00	0.00
43	Other benthic cephalopods	0.17	0.02	0.01	1.66	1.66	1.66	11.97	11.97	11.97	0.76	0.89	0.86	0.14	0.14	0.14	-	0.00	0.00	-	0.00	0.00
44	Bivalves	12.39	8.51	13.42	1.20	1.20	1.20	5.00	5.00	5.00	0.80	0.80	0.80	0.24	0.24	0.24	-	-	-	-	-	-





**Table 3.** Output estimates the 3 modelled MPA in the NW Mediterranean Sea (Ba = Cerbère-Banyuls; C = Cap de Creus; M = Medes Islands). TL = Trophic Level; F = fishing mortality (year<sup>-1</sup>); M2= predation mortality (year<sup>-1</sup>); M0 = other natural mortality (year<sup>-1</sup>); F/Z = exploitation rate (fishing mortality (F) / total mortality (Z)); FD = flow to detritus (t·km<sup>-2</sup>·year<sup>-1</sup>).

Functional group	TL			F			M2			M0			F/Z			FD		
	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M
1 Bottlenose dolphins	4.18	4.16	4.23	-	-	-	-	-	-	0.07	0.07	0.03	-	-	-	0.04	0.04	0.04
2 Striped dolphins	4.16	4.19	4.21	-	-	-	-	-	-	0.03	0.03	0.03	-	-	-	0.03	0.03	0.03
3 Endangered and pelagic seabirds	3.98	4.00	4.00	-	-	-	-	-	-	0.41	0.41	0.41	-	-	-	0.01	0.01	0.01
4 Gulls and cormorants	4.02	4.05	4.05	-	-	-	-	-	-	0.31	0.31	0.31	-	-	-	0.00	0.00	0.00
5 Terns	3.98	3.98	3.99	-	-	-	-	-	-	0.29	0.29	0.29	-	-	-	0.00	0.00	0.00
6 Loggerhead turtles	3.31	3.31	3.31	-	-	-	-	-	-	0.20	0.20	0.20	-	-	-	0.01	0.01	0.01
7 Non-commercial large pelagic fishes	3.75	3.75	3.75	-	-	-	-	-	-	0.14	0.14	0.14	-	-	-	0.00	0.00	0.00
8 Other large pelagic fishes	3.99	4.08	4.02	0.01	0.44	0.19	-	-	-	0.40	0.06	0.23	0.02	0.89	0.45	0.00	0.00	0.00
9 Mackerels	3.50	3.51	3.51	0.27	0.19	0.05	0.59	0.58	0.45	0.02	0.58	0.36	0.31	0.14	0.06	0.00	0.01	0.02
10 Horse mackerels	3.47	3.48	3.48	0.21	0.09	0.20	0.43	0.44	0.43	0.21	0.44	0.24	0.25	0.09	0.23	0.02	0.03	0.01
11 Other medium pelagic fishes	3.90	3.93	3.94	0.03	0.37	0.29	0.29	0.49	0.12	0.39	0.69	0.28	0.05	0.24	0.42	0.00	0.00	0.01
12 European sardine	2.95	2.95	2.95	-	-	-	1.61	1.16	0.86	0.02	0.48	0.77	-	-	-	0.06	0.12	0.17
13 European anchovy	3.01	3.01	3.01	-	-	-	1.58	2.38	1.03	0.05	0.14	0.60	-	-	-	0.05	0.03	0.04
14 Round sardinella	-	3.12	3.12	-	-	-	-	1.67	1.06	-	0.20	0.32	-	-	-	-	0.00	0.00
15 Other small pelagic fish	3.02	3.02	3.02	0.01	0.01	-	1.03	1.25	1.00	0.18	0.05	0.31	0.00	0.01	-	0.15	0.11	0.07
16 Anglerfish	3.95	4.00	4.00	0.16	0.37	0.36	0.00	0.00	0.00	0.31	0.21	0.21	0.34	0.63	0.63	0.06	0.01	0.01
17 European conger	3.87	3.97	3.96	0.05	0.30	0.37	0.04	0.06	0.11	0.47	0.21	0.17	0.09	0.52	0.57	0.03	0.01	0.00
18 European hake	3.91	3.94	3.93	0.02	0.39	0.27	0.27	0.14	0.10	0.38	0.18	0.22	0.03	0.55	0.46	0.04	0.03	0.06
19 Poor cod	3.35	3.35	3.35	0.00	0.01	0.08	0.61	1.04	0.77	0.54	0.12	0.31	0.00	0.01	0.07	0.05	0.02	0.02

<b>20</b>	<b>Common pandora</b>	3.28	3.25	3.28	0.04	0.30	0.31	0.62	0.49	0.43	0.09	0.09	0.08	0.05	0.34	0.38	0.15	0.03	0.06
<b>21</b>	<b>Sparidae</b>	2.96	2.98	3.00	0.00	0.01	0.00	0.90	0.55	0.42	0.18	0.47	0.61	0.00	0.01	0.00	36.13	8.09	24.18
<b>22</b>	<b>White seabream</b>	3.20	3.25	3.27	0.00	0.00	0.00	0.86	0.18	0.11	0.00	0.78	0.75	0.00	0.00	0.00	2.97	12.68	13.00
<b>23</b>	<b>Common two-banded seabream</b>	3.10	3.14	3.14	0.00	0.00	0.00	0.69	0.23	0.13	0.15	0.72	0.75	0.00	0.00	0.00	6.48	10.89	11.77
<b>24</b>	<b>Common dentex</b>	3.77	3.89	3.89	0.00	0.01	0.00	0.15	0.03	0.04	0.43	0.58	0.58	0.00	0.01	0.00	3.32	1.59	1.19
<b>25</b>	<b>Red scorpionfish</b>	3.47	3.51	3.59	0.00	0.01	0.01	0.49	0.16	0.13	0.22	0.55	0.57	0.00	0.01	0.01	3.06	1.30	1.31
<b>26</b>	<b>Scorpaenidae</b>	3.52	3.52	3.53	0.00	0.02	0.11	0.84	0.55	0.33	0.01	0.30	0.51	0.00	0.02	0.11	0.39	0.34	0.51
<b>27</b>	<b>Groupers</b>	3.78	3.96	3.95	0.00	0.00	0.00	-	-	-	0.18	0.22	0.22	0.00	0.01	0.00	11.12	2.11	1.70
<b>28</b>	<b>Brown meagre</b>	3.23	3.30	3.32	0.00	0.00	0.00	0.13	0.22	0.28	0.65	0.54	0.48	0.00	0.00	0.00	6.76	0.47	0.42
<b>29</b>	<b>Labridae and serranidae</b>	3.10	3.11	3.11	0.00	0.00	0.00	1.31	0.84	0.48	0.13	0.46	0.75	0.00	0.00	0.00	17.19	12.05	36.14
<b>30</b>	<b>Flatfishes</b>	3.23	3.24	3.24	0.01	0.23	0.50	0.54	0.55	0.73	0.77	0.40	0.11	0.01	0.19	0.37	0.20	0.04	0.04
<b>31</b>	<b>Other commercial medium demersal fish</b>	3.25	3.50	3.51	0.00	0.05	0.03	0.02	0.30	0.31	0.50	0.23	0.15	0.00	0.08	0.06	17.80	0.35	0.25
<b>32</b>	<b>No commercial medium demersal fish</b>	3.51	3.80	3.81	0.00	0.00	0.02	0.03	0.23	0.10	0.44	0.25	0.37	0.00	0.01	0.04	7.44	0.15	0.28
<b>33</b>	<b>Salema</b>	2.09	2.09	2.09	0.00	0.00	0.00	0.24	0.13	0.02	1.16	1.21	1.38	0.00	0.00	0.00	107.13	22.84	107.80
<b>34</b>	<b>Mugilidae</b>	2.20	2.30	2.30	0.00	0.01	0.00	0.25	0.20	0.04	0.66	0.79	0.83	0.00	0.01	0.00	31.77	1.73	5.90
<b>35</b>	<b>Red mullet</b>	3.15	3.16	3.15	-	0.34	0.33	0.85	0.50	0.61	0.20	0.20	0.13	-	0.33	0.31	0.51	0.03	0.02
<b>36</b>	<b>Surmulet</b>	3.24	3.25	3.24	-	0.04	0.01	0.73	0.90	0.08	0.19	0.03	0.85	-	0.04	0.01	1.12	0.08	1.82
<b>37</b>	<b>No commercial small demersal fish</b>	2.98	2.99	2.99	0.00	0.00	-	1.34	1.17	0.68	0.08	0.19	0.66	0.00	0.00	0.00	8.22	2.59	8.86
<b>38</b>	<b>Small-spotted catshark</b>	3.62	3.67	3.63	0.03	0.01	0.14	0.15	0.00	0.03	0.51	0.64	0.48	0.04	0.01	0.22	0.01	0.08	0.01
<b>39</b>	<b>Rays and skates</b>	3.73	3.74	3.75	0.01	0.01	0.13	0.02	0.00	0.00	0.69	0.66	0.66	0.01	0.02	0.16	0.06	0.08	0.06
<b>40</b>	<b>Torpedos</b>	3.93	3.98	3.99	0.05	-	-	0.02	0.00	0.02	0.45	0.50	0.48	0.10	-	-	0.02	0.01	0.00
<b>41</b>	<b>Coastal benthic cephalopods</b>	3.28	3.36	3.36	0.00	0.18	0.05	1.40	1.11	1.38	0.17	0.32	0.14	0.00	0.11	0.03	7.13	1.14	0.54
<b>42</b>	<b>Benthopelagic cephalopods</b>	3.47	3.60	3.63	0.00	0.03	0.26	0.50	1.39	1.11	1.22	0.30	0.34	0.00	0.02	0.15	0.32	0.04	0.06
<b>43</b>	<b>Other benthic cephalopods</b>	3.25	3.25	3.25	-	0.00	0.05	1.25	1.48	1.37	0.40	0.18	0.23	-	0.00	0.03	0.48	0.06	0.04
<b>44</b>	<b>Bivalves</b>	2.07	2.07	2.07	-	-	-	0.96	0.96	0.96	0.24	0.24	0.24	-	-	-	27.77	19.06	30.06
<b>45</b>	<b>Gastropods</b>	2.11	2.12	2.11	-	-	-	1.31	1.31	1.31	0.33	0.33	0.33	-	-	-	17.77	7.05	13.99
<b>46</b>	<b>European lobster</b>	3.16	3.16	3.16	0.09	-	0.82	0.19	0.99	0.33	0.92	0.22	0.06	0.08	-	0.68	0.03	0.00	0.00

<b>47 Other commercial decapods</b>	3.08	3.07	3.08	0.16	0.12	0.07	2.73	1.37	2.75	0.15	1.51	0.06	0.05	0.04	0.03	0.01	0.02	0.01
<b>48 Non-commercial decapods</b>	2.72	2.75	2.74	-	-	-	3.18	3.18	3.18	0.80	0.80	0.80	-	-	-	100.97	27.34	46.10
<b>49 Purple sea urchin</b>	2.00	2.00	2.00	-	0.00	0.00	0.61	0.66	0.50	0.27	0.22	0.37	-	0.00	0.00	26.14	11.10	27.19
<b>50 Other sea urchin</b>	2.00	2.00	2.00	-	-	-	0.44	0.49	0.54	0.37	0.33	0.28	-	-	-	2.21	0.91	2.14
<b>51 Sea cucumbers</b>	2.00	2.00	2.00	-	-	-	0.71	0.71	0.71	0.18	0.18	0.18	-	-	-	8.18	2.52	4.62
<b>52 Other macro-benthos</b>	2.02	2.03	2.03	-	-	-	2.81	2.81	2.81	0.70	0.70	0.70	-	-	-	759.72	270.03	502.65
<b>53 Jellyfish</b>	3.07	3.07	3.07	-	-	-	1.13	1.50	1.30	0.86	0.48	0.69	-	-	-	0.01	0.01	0.01
<b>54 Salps and other gelatinous zooplankton</b>	2.04	2.04	2.04	-	-	-	0.58	0.60	0.61	1.45	1.43	1.41	-	-	-	0.02	0.02	0.02
<b>55 Red coral</b>	2.46	2.46	2.46	-	-	-	-	-	-	0.22	0.22	0.22	-	-	-	0.04	0.02	0.08
<b>56 Other corals and gorgonians</b>	2.17	2.17	2.17	-	-	-	0.07	0.08	0.04	0.15	0.14	0.18	-	-	-	0.49	0.13	0.58
<b>57 Macrozooplankton</b>	2.63	2.63	2.63	-	-	-	14.26	13.05	14.91	0.74	1.95	0.09	-	-	-	13.13	13.90	12.72
<b>58 Meso and microzooplankton</b>	2.00	2.00	2.00	-	-	-	36.82	23.19	38.40	4.88	18.51	3.30	-	-	-	197.86	248.85	191.96
<b>59 Suprabenthos</b>	2.17	2.16	2.15	-	-	-	7.38	6.56	6.56	0.82	1.64	1.64	-	-	-	412.88	186.24	384.65
<b>60 Mediterranean seagrass</b>	1.00	1.00	1.00	-	-	-	1.53	0.14	2.13	0.82	2.21	0.22	-	-	-	18.41	118.70	3.32
<b>61 Other seagrasses</b>	-	1	1	-	-	-	-	1.43	3.08	-	2.61	0.96	-	-	-	-	2.29	0.86
<b>62 Erected algae</b>	-	1	1	-	-	-	-	0.06	0.10	-	0.64	0.60	-	-	-	-	6.17	5.11
<b>63 Seaweeds</b>	1	1	1	-	-	-	0.28	0.19	0.25	0.42	0.51	0.45	-	-	-	3.19	2.15	3.43
<b>64 Small phytoplankton</b>	1	1	1	-	-	-	161.77	138.94	142.78	1.18	24.02	20.17	-	-	-	1.11	21.02	20.30
<b>65 Large phytoplankton</b>	1	1	1	-	-	-	39.51	28.03	35.08	89.86	101.34	94.28	-	-	-	759.42	797.91	853.34
<b>66 Detritus</b>	1	1	1	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-
<b>67 Discards and by-catch</b>	1	1	1	-	-	-	-	-	-	-	-	-	-	-	-	0.00	0.01	0.01

**Table 4.** Output estimates the 3 modelled MPA in the NW Mediterranean Sea (Ba = Cerbère-Banyuls; C = Cap de Creus; M = Medes Islands). R/A = Respiration/Assimilation ratio; R/B = Respiration/Biomass ratio (year<sup>-1</sup>); P/R = Production/Respiration ratio; NE = Net Efficiency.

Functional group	R/A			R/B			P/R			NE		
	Ba	C	M	Ba	C	M	Ba	C	M	Ba	C	M
1 Bottlenose dolphins	0.99	0.99	1.00	10.18	10.18	10.22	0.01	0.01	0.00	0.01	0.01	0.00
2 Striped dolphins	1.00	1.00	1.00	11.25	11.25	11.25	0.00	0.00	0.00	0.00	0.00	0.00
3 Endangered and pelagic seabirds	0.99	0.99	0.99	58.59	58.59	58.59	0.01	0.01	0.01	0.01	0.01	0.01
4 Gulls and cormorants	0.99	0.99	0.99	56.32	56.32	56.32	0.01	0.01	0.01	0.01	0.01	0.01
5 Terns	1.00	1.00	1.00	68.03	68.03	68.03	0.00	0.00	0.00	0.00	0.00	0.00
6 Loggerhead turtles	0.90	0.90	0.90	1.81	1.81	1.81	0.11	0.11	0.11	0.10	0.10	0.10
7 Non-commercial large pelagic fishes	0.89	0.89	0.89	1.16	1.16	1.16	0.12	0.12	0.12	0.11	0.11	0.11
8 Other large pelagic fishes	0.88	0.87	0.88	2.89	3.47	2.89	0.14	0.14	0.14	0.13	0.13	0.13
9 Mackerels	0.85	0.85	0.85	4.97	4.98	4.99	0.18	0.18	0.17	0.15	0.15	0.15
10 Horse mackerels	0.85	0.85	0.85	4.83	4.89	4.94	0.18	0.18	0.18	0.15	0.15	0.15
11 Other medium pelagic fishes	0.86	0.81	0.87	4.46	3.38	4.48	0.16	0.23	0.15	0.14	0.19	0.13
12 European sardine	0.81	0.81	0.81	7.06	7.09	7.09	0.23	0.23	0.23	0.19	0.19	0.19
13 European anchovy	0.81	0.81	0.81	7.05	7.07	7.07	0.23	0.23	0.23	0.19	0.19	0.19
14 Round sardinella	-	0.81	0.81	-	5.97	5.97	-	0.23	0.23	-	0.19	0.19
15 Other small pelagic fish	0.85	0.86	0.83	6.67	8.32	6.57	0.18	0.16	0.20	0.15	0.14	0.17
16 Anglerfish	0.86	0.83	0.83	2.97	2.86	2.86	0.16	0.20	0.20	0.14	0.17	0.17
17 European conger	0.85	0.84	0.83	3.15	3.13	3.06	0.18	0.18	0.21	0.15	0.16	0.17
18 European hake	0.85	0.84	0.87	3.80	3.78	3.90	0.18	0.19	0.15	0.15	0.16	0.13
19 Poor cod	0.85	0.85	0.85	6.56	6.59	6.59	0.18	0.18	0.18	0.15	0.15	0.15
20 Common pandora	0.85	0.83	0.84	4.26	4.14	4.20	0.18	0.21	0.20	0.15	0.18	0.16
21 Sparidae	0.85	0.85	0.85	6.11	5.83	5.84	0.18	0.18	0.18	0.15	0.15	0.15

<b>22</b>	<b>White seabream</b>	0.85	0.85	0.85	4.89	5.45	4.89	0.18	0.18	0.18	0.15	0.15	0.15
<b>23</b>	<b>Common two-banded seabream</b>	0.85	0.85	0.85	4.77	5.41	4.98	0.18	0.18	0.18	0.15	0.15	0.15
<b>24</b>	<b>Common dentex</b>	0.86	0.85	0.85	3.63	3.49	3.49	0.16	0.18	0.18	0.14	0.15	0.15
<b>25</b>	<b>Red scorpionfish</b>	0.85	0.85	0.85	4.03	4.09	4.00	0.18	0.18	0.18	0.15	0.15	0.15
<b>26</b>	<b>Scorpaenidae</b>	0.85	0.85	0.83	4.80	4.73	4.69	0.18	0.18	0.20	0.15	0.15	0.17
<b>27</b>	<b>Groupers</b>	0.94	0.93	0.93	2.65	2.90	2.90	0.07	0.08	0.08	0.06	0.07	0.07
<b>28</b>	<b>Brown meagre</b>	0.85	0.85	0.85	4.41	4.32	4.32	0.18	0.18	0.18	0.15	0.15	0.15
<b>29</b>	<b>Labridae and serranidae</b>	0.83	0.85	0.85	6.79	7.37	7.00	0.21	0.18	0.18	0.18	0.15	0.15
<b>30</b>	<b>Flatfishes</b>	0.85	0.86	0.82	7.44	7.12	6.24	0.18	0.16	0.22	0.15	0.14	0.18
<b>31</b>	<b>Other commercial medium demersal fish</b>	0.88	0.86	0.89	3.68	3.63	3.72	0.14	0.16	0.13	0.13	0.14	0.11
<b>32</b>	<b>No commercial medium demersal fish</b>	0.87	0.87	0.87	3.29	3.28	3.26	0.14	0.15	0.15	0.13	0.13	0.13
<b>33</b>	<b>Salema</b>	0.81	0.81	0.81	6.06	5.79	6.06	0.23	0.23	0.23	0.19	0.19	0.19
<b>34</b>	<b>Mugilidae</b>	0.81	0.81	0.81	3.94	4.33	3.79	0.23	0.23	0.23	0.19	0.19	0.19
<b>35</b>	<b>Red mullet</b>	0.83	0.83	0.82	4.93	4.93	4.91	0.21	0.21	0.22	0.18	0.18	0.18
<b>36</b>	<b>Surmulet</b>	0.83	0.81	0.82	4.34	4.28	4.32	0.21	0.23	0.22	0.18	0.19	0.18
<b>37</b>	<b>No commercial small demersal fish</b>	0.85	0.83	0.83	8.06	6.56	6.57	0.18	0.21	0.20	0.15	0.17	0.17
<b>38</b>	<b>Small-spotted catshark</b>	0.86	0.87	0.87	4.28	4.31	4.31	0.16	0.15	0.15	0.14	0.13	0.13
<b>39</b>	<b>Rays and skates</b>	0.85	0.86	0.84	4.04	4.10	3.99	0.18	0.16	0.20	0.15	0.14	0.16
<b>40</b>	<b>Torpedos</b>	0.87	0.88	0.87	3.47	3.50	3.50	0.15	0.14	0.14	0.13	0.13	0.13
<b>41</b>	<b>Coastal benthic cephalopods</b>	0.82	0.82	0.82	7.24	7.20	7.24	0.22	0.22	0.22	0.18	0.18	0.18
<b>42</b>	<b>Benthopelagic cephalopods</b>	0.87	0.87	0.87	11.50	11.50	11.51	0.15	0.15	0.15	0.13	0.13	0.13
<b>43</b>	<b>Other benthic cephalopods</b>	0.83	0.83	0.83	7.92	7.91	7.92	0.21	0.21	0.21	0.17	0.17	0.17
<b>44</b>	<b>Bivalves</b>	0.60	0.60	0.60	1.80	1.80	1.80	0.67	0.67	0.67	0.40	0.40	0.40
<b>45</b>	<b>Gastropods</b>	0.60	0.60	0.60	2.46	2.46	2.46	0.67	0.67	0.67	0.40	0.40	0.40
<b>46</b>	<b>European lobster</b>	0.71	0.75	0.75	3.02	3.63	3.63	0.40	0.33	0.33	0.29	0.25	0.25
<b>47</b>	<b>Other commercial decapods</b>	0.73	0.77	0.78	8.30	9.96	10.08	0.37	0.30	0.29	0.27	0.23	0.22
<b>48</b>	<b>Non-commercial decapods</b>	0.66	0.70	0.70	7.58	9.23	9.23	0.53	0.43	0.43	0.34	0.30	0.30
<b>49</b>	<b>Purple sea urchin</b>	0.76	0.76	0.76	2.72	2.72	2.72	0.32	0.32	0.32	0.24	0.24	0.24



## **PREBAL analysis and balancing procedure**

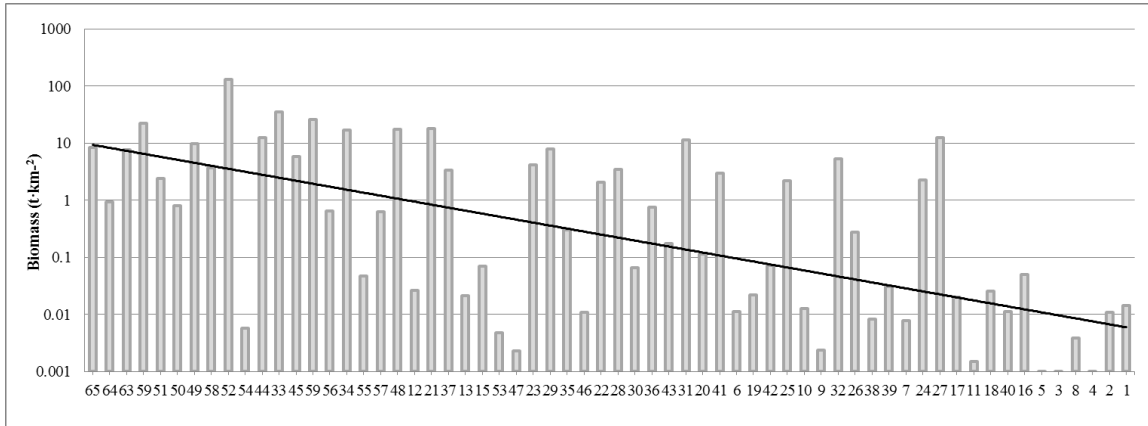
An Ecopath model is considered ecologically and thermodynamically balanced under the following conditions: (1) estimated  $EE < 1$  for all functional groups, (2) values of  $P/Q$  (production/consumption rate or gross efficiency of food conversion, GE) are between 0.1 and 0.35 with the exception of some fast growing groups, (3)  $R/A$  (respiration/food assimilation)  $< 1$ , (4)  $R/B$  (respiration/biomass) are between 1 and 10 for fishes and higher values for small organisms, (5)  $NE$  (net efficiency of food conversion)  $> GE$  and (6)  $P/R$  (production/respiration)  $< 1$  (Christensen et al., 2008; Heymans et al., 2016) (Table 2 in the manuscript and Table 3 in SOM 3).

Initial results of the three MPA models showed that the  $EE > 1$  for several groups. To achieve mass-balance, we applied a manual mass-balanced procedure following a top-down approach modifying appropriate input parameters (starting from the groups with higher trophic levels) and following the best practice guidelines provided in the literature (Heymans et al., 2016).

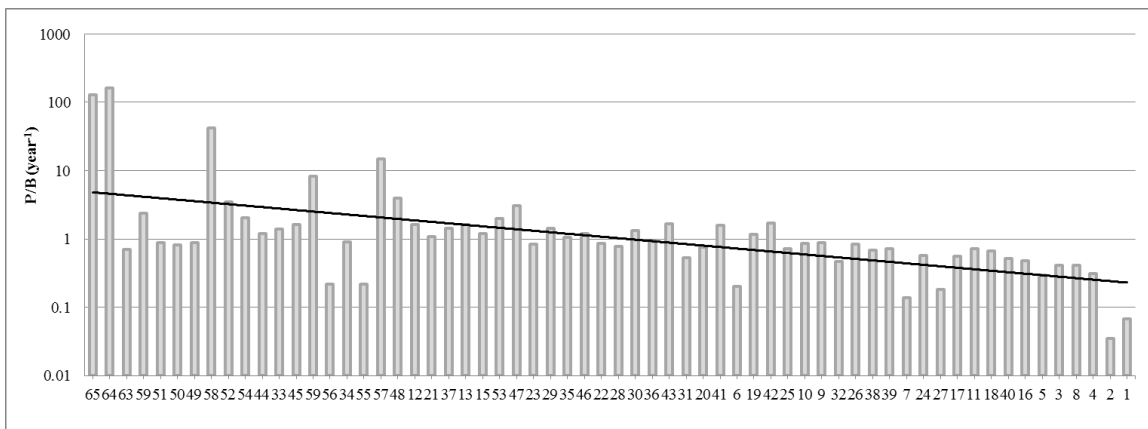
In order to ensure that the model parameters obeyed general ecologic principles and to guide the balancing procedure we used the PREBAL analysis (Link, 2010). This analysis highlighted that some  $P/B$  and  $Q/B$  values had to be adjusted since they were too low or too high based on their trophic levels based on previous modelling developments (Corrales et al., 2015; Corrales et al., 2017) (see Fig. 1, 2 and 3).

After testing the model parameters with PREBAL, biomass and diets were readjusted where needed, as in other Ecopath models developed in the Mediterranean Sea (Coll et al., 2006; Tsagarakis et al., 2010; Corrales et al., 2015). For pelagic fish groups (mackerels, horse mackerels, European sardine, European anchovy and other small pelagic fish), we added immigration biomass because studied areas are relatively small compared to the dispersal rate of those species (e.g., Papetti et al., 2013). Finally, the diet matrix was slightly adjusted to take into account the abundances of species in the ecosystem (SOM 4).

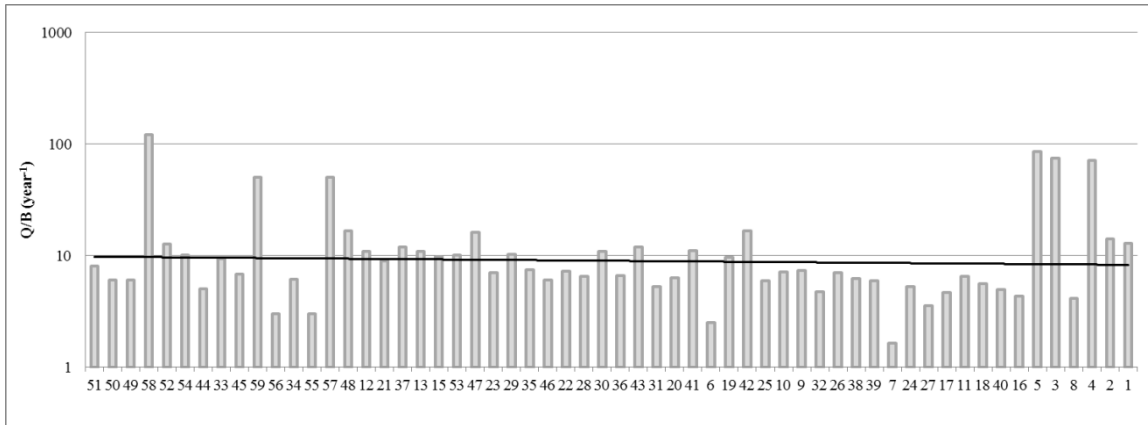
a)



b)



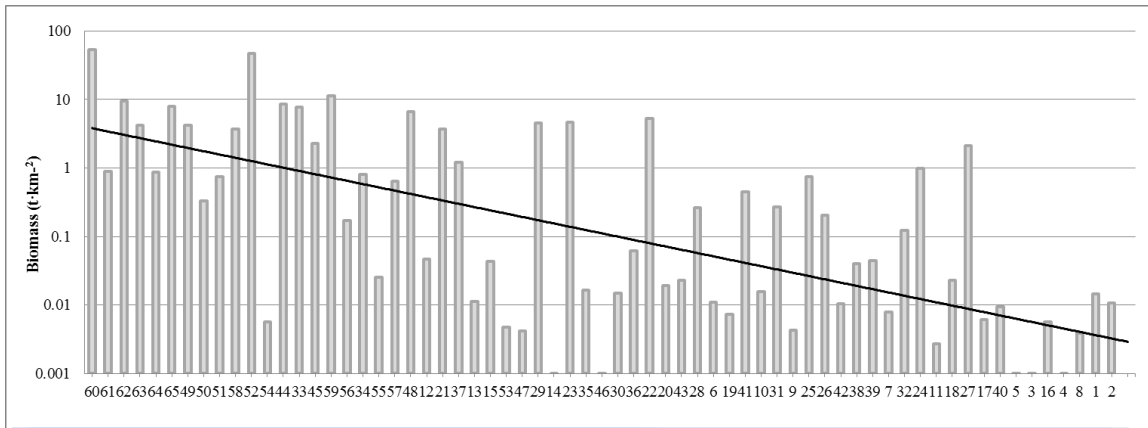
c)



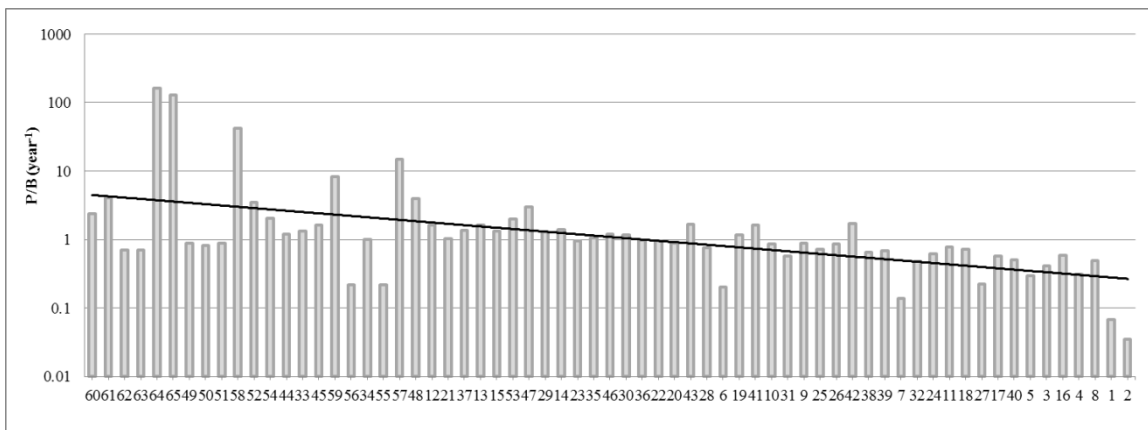
**Fig. 1.** PREBAL of the Cerbère-Banyuls MPA plotting Biomass estimates (t·km<sup>-2</sup>) (a), Production/Biomass ratio (year<sup>-1</sup>) (b), and Consumption/Biomass ratio (year<sup>-1</sup>) (c) on a log scale vs species ranked by trophic level, from the lowest to highest trophic level. Trend lines are given. Numbers in identify the functional groups of the model.

a)

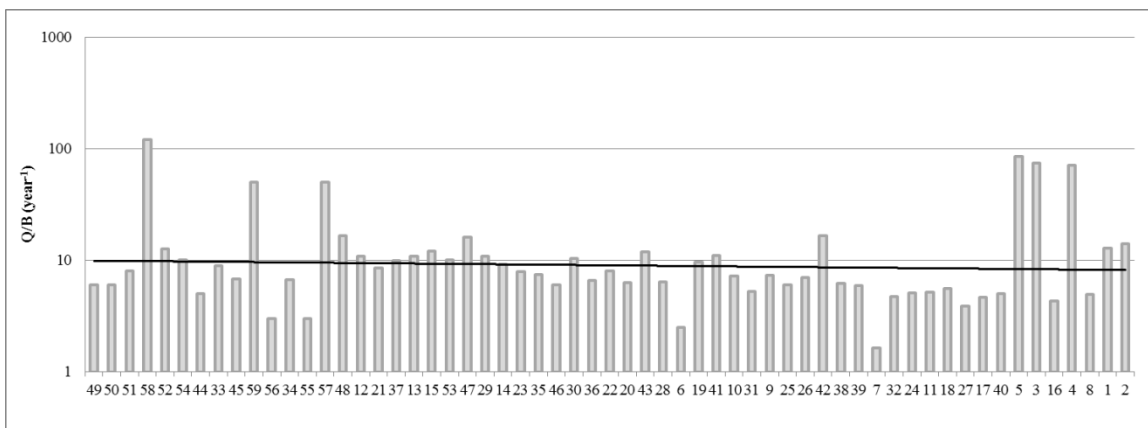




b)

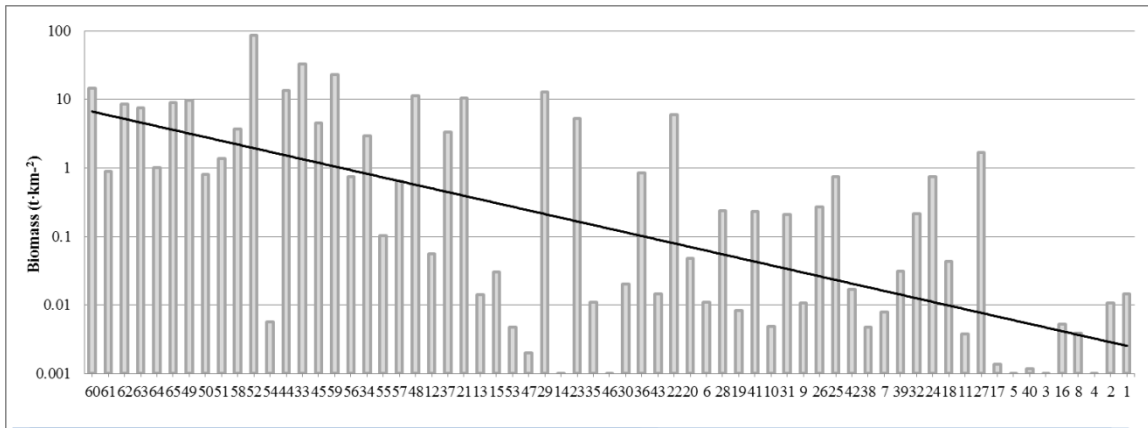


c)

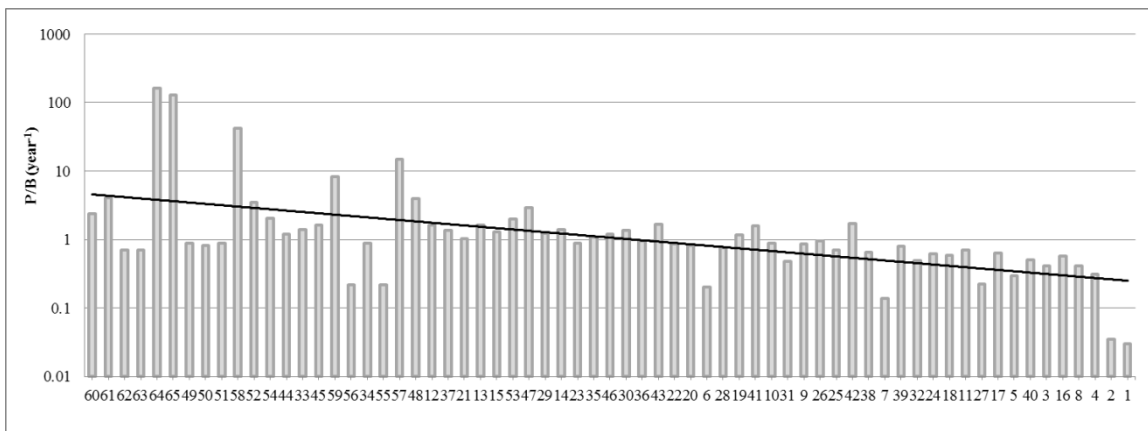


**Fig. 2.** PREBAL of the Cap de Creus MPA plotting Biomass estimates ( $t \cdot km^{-2}$ ) (a), Production/Biomass ratio ( $year^{-1}$ ) (b), and Consumption/Biomass ratio ( $year^{-1}$ ) (c) on a log scale vs species ranked by trophic level, from the lowest to highest trophic level. Trend lines are given. Numbers in identify the functional groups of the model.

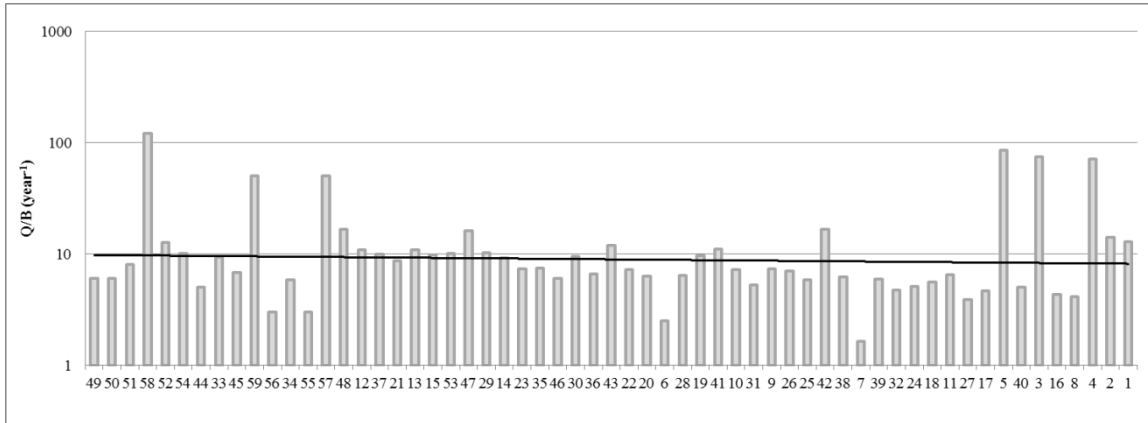
a)



b)



c)



**Fig. 3.** PREBAL of the Medes Islands MPA plotting Biomass estimates ( $t \cdot km^{-2}$ ) (a), Production/Biomass ratio ( $year^{-1}$ ) (b), and Consumption/Biomass ratio ( $year^{-1}$ ) (c) on a log scale vs species ranked by trophic level, from the lowest to highest trophic level. Trend lines are given. Numbers in identify the functional groups of the model.

## Pedigree routine

The pedigree routine (Christensen and Walters, 2004) was used to quantify the uncertainty associated with the input parameters and the quality of the models and to validate choices made in balancing the model. For each input data, pedigree values were assigned manually to record the degree of confidence associated with the data except for diet pedigree values, which were obtained from the predator point of view considering four diet features (region, year, type of data and method). With the information for each functional group, the pedigree index was calculated for the overall model. The pedigree index ranges between 0 (low quality) and 1 (high quality), allowing a description of the quality of the model that can be compared to other models. The confidence intervals for the pedigree analysis and index values used are described in Table 4.

**Table 4.** Confidence intervals and index values used to describe the uncertainty for each input parameter of the balanced Ecopath model (as default parameters).

<b>Parameter</b>	<b>Option</b>	<b>Index value</b>	<b>CI (%)</b>
<b>Biomass</b>			
	Estimated by Ecopath	0.0	±80
	From other model	0.0	±80
	Guesstimate	0.0	±80
	Approximate or indirect method	0.4	±50
	Sampling/locally, low precision	0.6	±40
	Sampling/locally, high precision	0.8	±10
<b>PB/ and Q/B</b>			
	Estimated by Ecopath	0.0	±80
	Guesstimated	0.1	±70
	From other model	0.2	±60
	Empirical relationship	0.5	±50
	Similar species, similar system, low precision	0.6	±40
	Similar species, same system, low precision	0.7	±30
	Same species, similar system, low precision	0.8	±20
	Same species, same system, high precision	1	±10
<b>Diet</b>			
	From other model	0.1	±80
	General knowledge of related groups/species	0.1	±70
	General knowledge for same groups/species	0.2	±60
	Qualitative diet composition study	0.5	±50

	Quantitative (but limited) diet composition study	0.6	±40
	Quantitative (detailed) diet composition study	0.7	±30
<b>Catch</b>			
	From other model	0.1	±90
	National statistics	0.5	±50
	Intermediate (local study and national statistics)	0.6	±40
	Local study, low precision/incomplete data	0.7	±30
	Local study, high precision/complete data	0.9	±10

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**Table 1.** Diet composition matrix for the Cerbère-Banyuls MPA model. Cells in grey are values lo

Prey \ predator	1	2	3	4	5	6	7	8
1 Bottlenose dolphins								
2 Striped dolphins								
3 Endangered and pelagic seabirds								
4 Gulls and cormorants								
5 Terns								
6 Loggerhead turtles								
7 Non-commercial large pelagic fishes								
8 Other large pelagic fishes								
9 Mackerels				0.01	0.01			0.01
10 Horse mackerels	0.01	0.01	0.01	0.02	0.01			0.01
11 Other medium pelagic fishes								
12 European sardine	0.04	0.03	0.04	0.05	0.08			0.02
13 European anchovy	0.03	0.03	0.05	0.02	0.02			0.01
14 Other small pelagic fish	0.03	0.03	0.03	0.01	0.08			0.01
15 Anglerfish								
16 European conger								
17 European hake	0.01							
18 Poor cod		0.01						
19 Common pandora		0.01						
20 Sparidae	0.01	0.02		0.02				0.08
21 White seabream								
22 Common two-banded seabream								
23 Common dentex								
24 Red scorpionfish								
25 Scorpaenidae								
26 Groupers								
27 Brown meagre								
28 Labridae and Serranidae				0.01				
29 Flatfishes				0.01				
30 Other commercial medium demersal fish	0.01	0.02		0.01				
31 No commercial medium demersal fish								
32 Salema								
33 Mugilidae								
34 Red mullet								
35 Surmulet								
36 No commercial small demersal fish		0.01	0.05	0.01				0.01
37 Small-spotted catshark								
38 Rays and skates								
39 Torpedos								
40 Coastal benthic cephalopods	0.01							
41 Benthopelagic cephalopods	0.03	0.03	0.02					0.01
42 Other benthic cephalopods	0.01	0.01						
43 Bivalves								
44 Gastropods								
45 European lobster								
46 Other commercial decapods				0.01				
47 Non-commercial decapods						0.01		

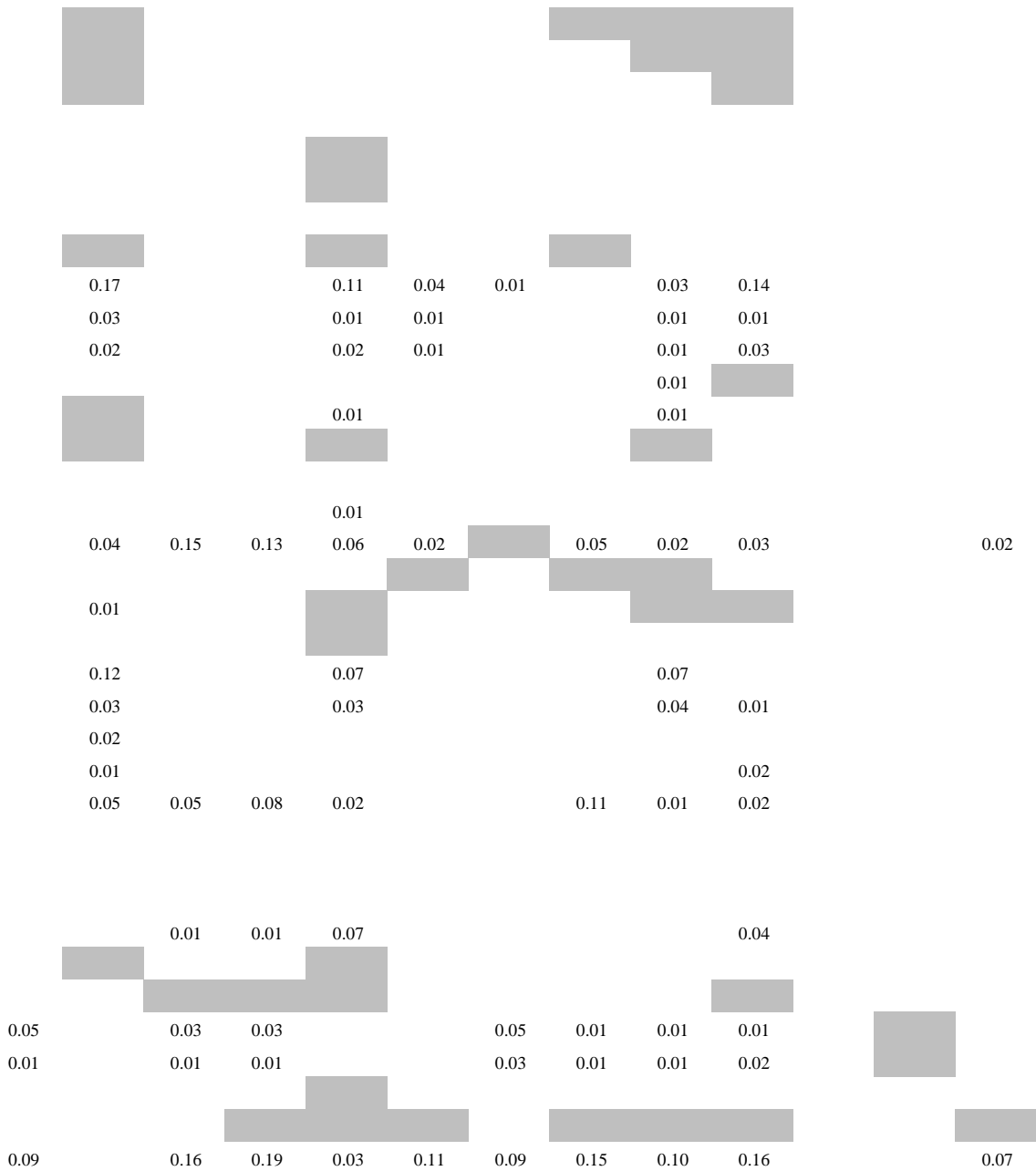


lower than 0.01. The numbers identify the functional groups of the model.

9	10	11	12	13	14	15	16	17	18	19	20	21
		0.01										
	0.01	0.07				0.01		0.03				
0.01	0.02	0.07				0.01		0.02				
	0.01	0.04						0.03				
						0.02						
		0.02				0.02		0.01				
		0.01										
		0.08				0.07	0.04	0.07			0.01	
						0.03	0.02	0.03				
						0.04	0.03	0.03				
								0.01				
						0.04	0.01	0.01				
						0.07	0.07	0.04			0.01	0.02
						0.01	0.04			0.03		
		0.02				0.08	0.20	0.01				
						0.01	0.02					
							0.01					
								0.01				
		0.09				0.07		0.06	0.05	0.01		0.01
						0.01						
	0.01	0.01	0.01					0.01				
						0.01		0.01				
							0.06			0.03	0.01	0.04
							0.02			0.05	0.01	0.01
						0.08	0.11	0.03	0.04	0.12	0.10	0.13











0.04	0.02	0.03	0.01
0.02		0.03	0.01
0.04	0.02	0.03	0.02



0.01			
0.05	0.08	0.02	0.03
0.02	0.01	0.01	
0.02	0.02	0.01	

0.01

0.03	0.06	0.02	0.06	0.03	0.01	0.02	0.02
		0.02	0.02				

0.01

0.01	0.05					0.01
0.04	0.12					
0.06	0.05	0.01	0.04	0.08	0.02	0.02



0.05			0.01	0.01	0.01	0.05	0.19	0.09	0.01
0.06			0.02	0.05	0.00	0.03	0.24	0.03	



0.17	0.07	0.20	0.17	0.08	0.21	0.09	0.11	0.09	0.06	0.00
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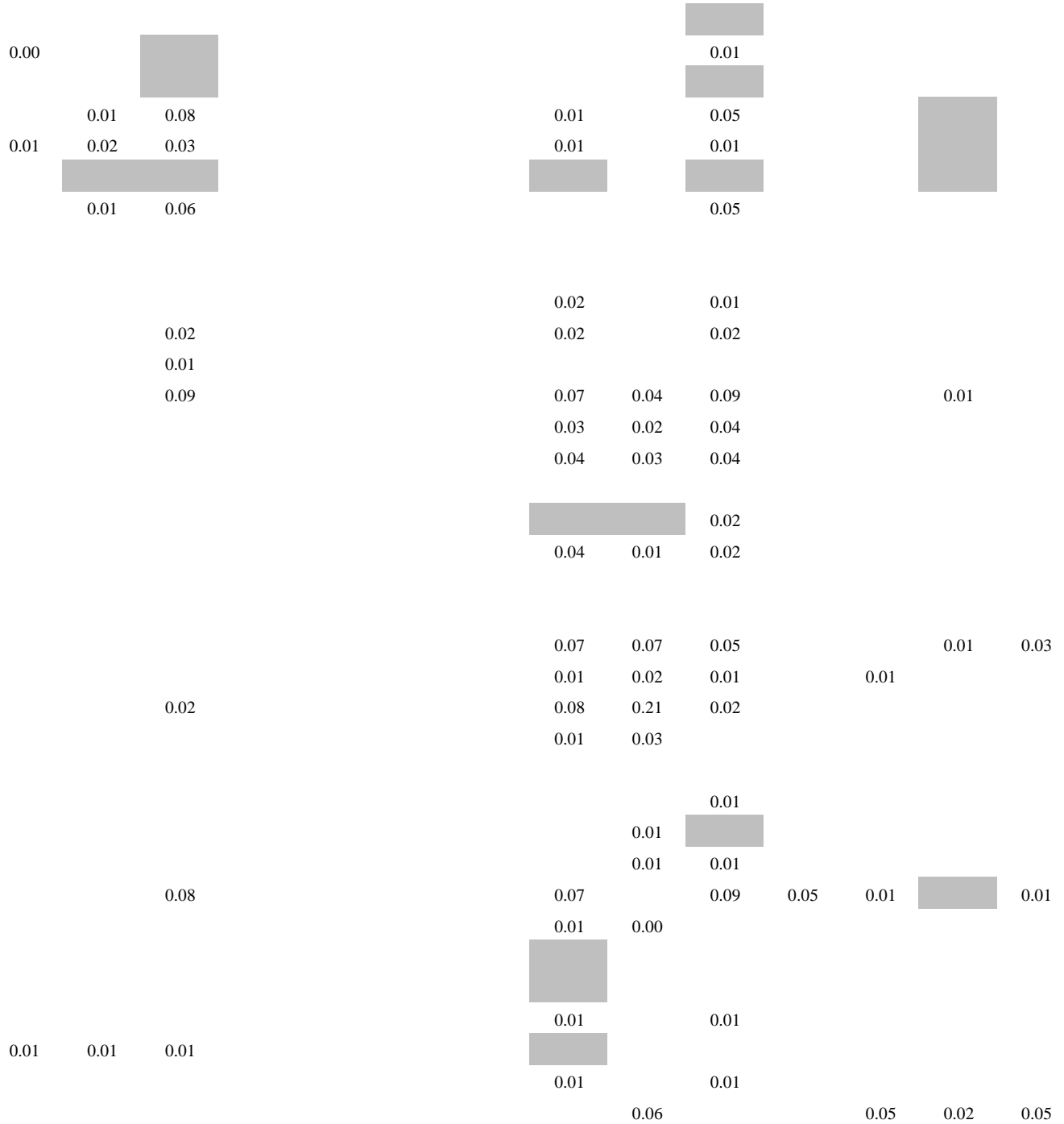




01. The numbers identify the functional groups of the model.

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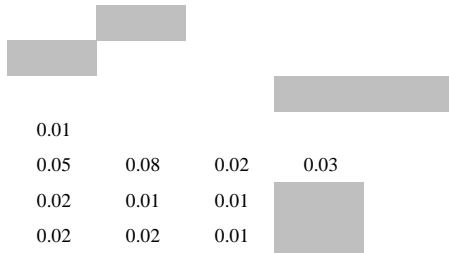
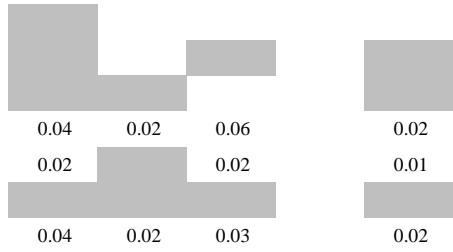




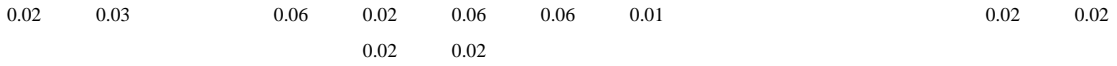




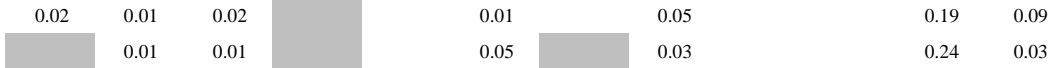
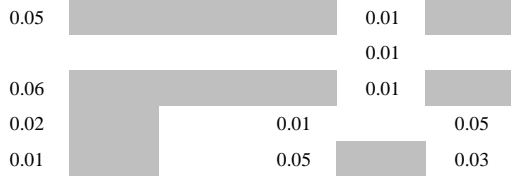
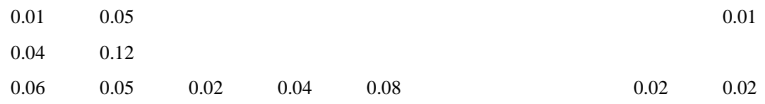




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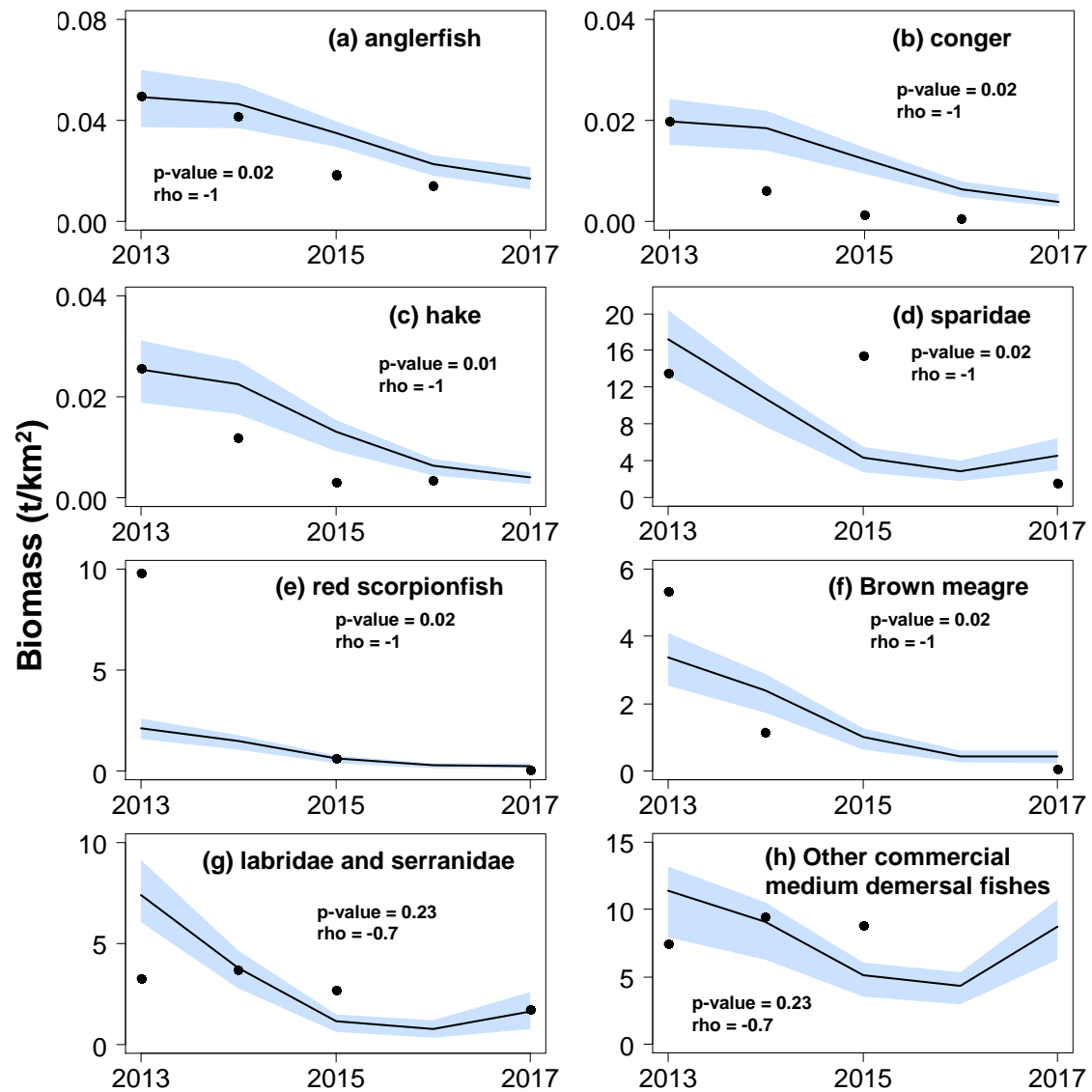
<b>23. Common two-banded seabream</b>	Relative biomass	2013-2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004, 2006-2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>24. Common dentex</b>	Relative biomass	2013-2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004-2009, 2014 and 2016	UVC (Hereu et al., 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>25. Red scorpionfish</b>	Relative biomass	2013, 2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004, 2006-2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>26. Scorpaenidae</b>	Relative biomass	2013, 2015 and 2017	UVC (Claudet, 2013)					X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>27. Groupers</b>	Relative biomass			2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	-		X	X
	Absolut catch			2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2016	ICM database and Sacanell (2012)		X
<b>28. Brown meagre</b>	Relative biomass	2013, 2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004, 2006-2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>29. Labridae and serranidae+</b>	Relative biomass	2013-2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004, 2006-2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>30. Flatfishes</b>	Relative biomass	2013-2015 and 2017	UVC (Claudet, 2013)	-		-		X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
	Absolut biomass	-		2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	2004-2006, 2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X

<b>31. Other commercial medium demersal fish</b>	Relative biomass	2013,2014 and 2015	UVC (Claudet, 2013)	2008, 2009, 2011, 2014 and 2016	UVC (Hereu et al., 2017)	2004-2009, 2011, 2014 and 2016	MEDITS (Bertrand et al., 2002) and UVC (Hereu, 2017)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>32. No commercial medium demersal fish</b>	Relative biomass	2013, 2014 and 2017	UVC (Claudet, 2013)	2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	-		X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)		X
<b>33. Salema</b>	Absolut biomass	2013, 2014 and 2017	UVC (Claudet, 2013)	-		-		X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004, 2005, 2009-2014, 2016 and 2017	ICM database and Sacanell (2012)		X
<b>34. Mugilidae</b>	Absolut biomass	2013 and 2017	UVC (Claudet, 2013)	-		-		X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2006, 2008-2015, 2016 and 2017	ICM database and Sacanell (2012)		X
<b>35. Red mullet</b>	Absolut biomass	2013-2016	UVC (Claudet, 2013)	2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	2004-2006, 2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2008 and 2012-2014	ICM database and Sacanell (2012)		X
<b>36. Surmullet</b>	Absolut biomass	2013-2015 and 2017	UVC (Claudet, 2013)	2008, 2009, 2011 and 2014	MEDITS (Bertrand et al., 2002)	-	MEDITS (Bertrand et al., 2002)	X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2011 and 2014	ICM database and Sacanell (2012)		X
<b>37. No commercial small demersal fish</b>	Absolut biomass	2013-2015 and 2017	UVC (Claudet, 2013)	2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	-		X	X
	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	-	ICM database and Sacanell (2012)		X
<b>38. Small-spotted catshark</b>	Absolut biomass	-		2008, 2011, 2012 and 2014	MEDITS (Bertrand et al., 2002)	2004 and 2012	MEDITS (Bertrand et al., 2002)	X	X
	Absolut catch	-		2008, 2009, 2011, 2012, 2014 and 2016	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004, 2005, 2009-2013 and 2017	ICM database and Sacanell (2012)		X
<b>39. Rays and skates</b>	Absolut biomass	-		2008, 2011 and 2014	MEDITS (Bertrand et al., 2002)	2004-2006, 2012 and 2014	MEDITS (Bertrand et al., 2002)	X	X

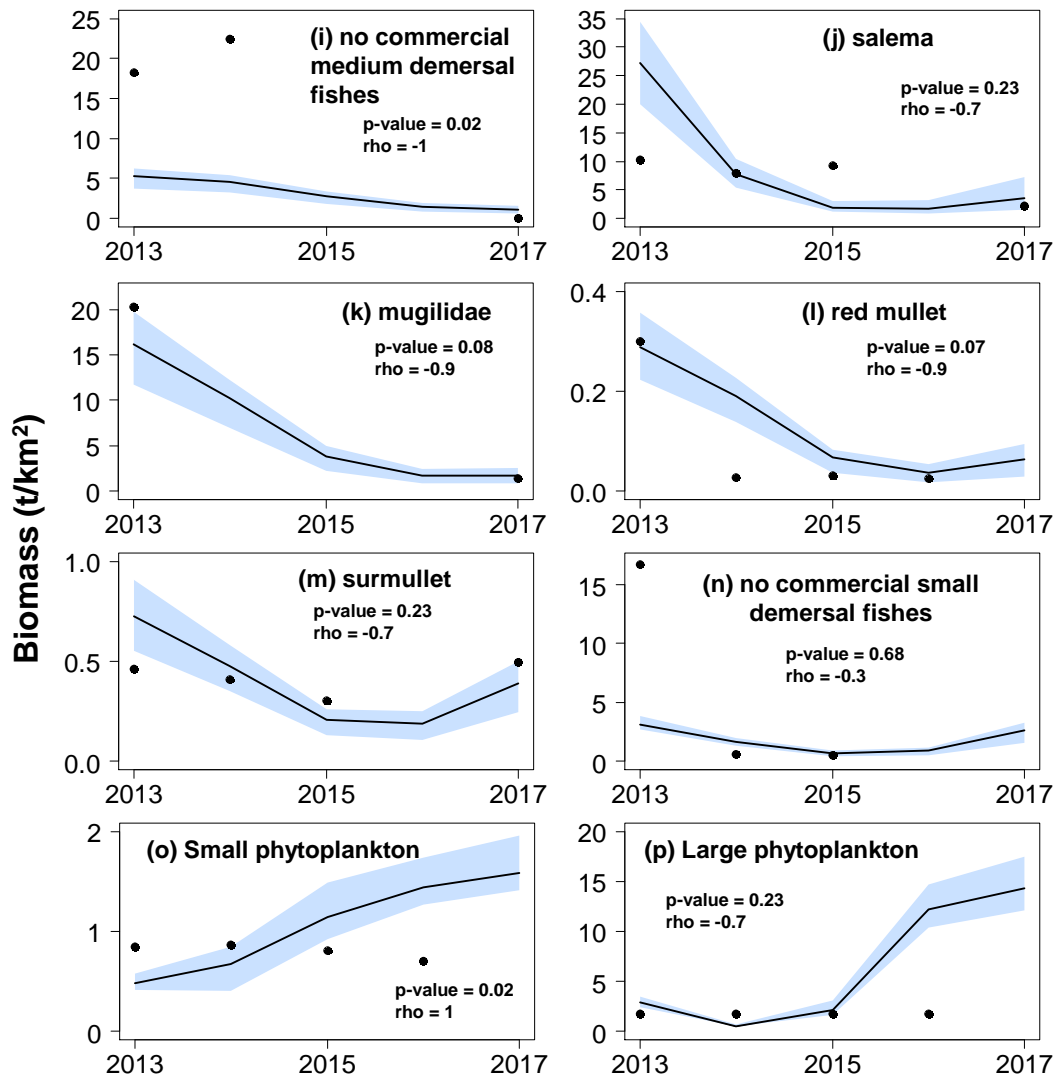
<b>40. Torpedos</b>	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)			X
	Absolut biomass	-		2008, 2009 and 2014	MEDITS (Bertrand et al., 2002)	-		X		X
<b>41. Coastal benthic cephalopods</b>	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	-	ICM database and Sacanell (2012)			X
	Absolut biomass	-		-		-		X		X
<b>42. Benthopelagic cephalopods</b>	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)			X
	Absolut biomass	-		2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	2004-2006, 2008, 2009 and 2011-2014	MEDITS (Bertrand et al., 2002)	X		X
<b>43. Other benthic cephalopods</b>	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2017	ICM database and Sacanell (2012)			X
	Absolut biomass	-		2008 and 2012-2014	MEDITS (Bertrand et al., 2002)	-		X		X
<b>46. European lobster</b>	Absolut catch	-		2008-2013 and 2015	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004-2008 and 2011	ICM database and Sacanell (2012)			X
	Absolut biomass	-		-		2004-2017	ICM database and Sacanell (2012)			X
<b>47. Other commercial decapods</b>	Absolut catch	-		2008-2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004, 2006, 2008-2011, 2013, 2014 and 2017	ICM database and Sacanell (2012)			X
	Absolut biomass	-		-		-		X		X
<b>49. Purple sea urchin</b>	Relative biomass	-		-		2004, 2005, 2009-2012	UVC (Hereu et al., 2017)	X		X
	Absolut catch	-		2008, 2009, 2011, 2014 and 2017	ICM database; Lloret et al., 2008a and Lloret et al., 2008b;	2004 and 2016	ICM database and Sacanell (2012)			X
<b>50. Other sea urchins</b>	Relative biomass	-		-		2004, 2005, 2008-2011	UVC (Hereu et al., 2017)	X		X
<b>64. Small phytoplankton</b>	Relative biomass	2013-2017	Joint Research Center 2017a	2008-2017	Joint Research Center 2017a	2004-2017	Joint Research Center 2017a	X		X
<b>65. Large phytoplankton</b>	Relative biomass	2013-2017	Joint Research Center 2017a	2008-2017	Joint Research Center 2017a	2004-2017	Joint Research Center 2017a	X		X
<b>All fleets</b>	Relative effort	2013-2017	IFREMER (2015) data from WP2	2008-2017	Ministry of Agriculture, Livestock, Fisheries and Food of the Catalan Government and data from WP2	2004-2017	Ministry of Agriculture, Livestock, Fisheries and Food of the Catalan Government and data from 2	X		



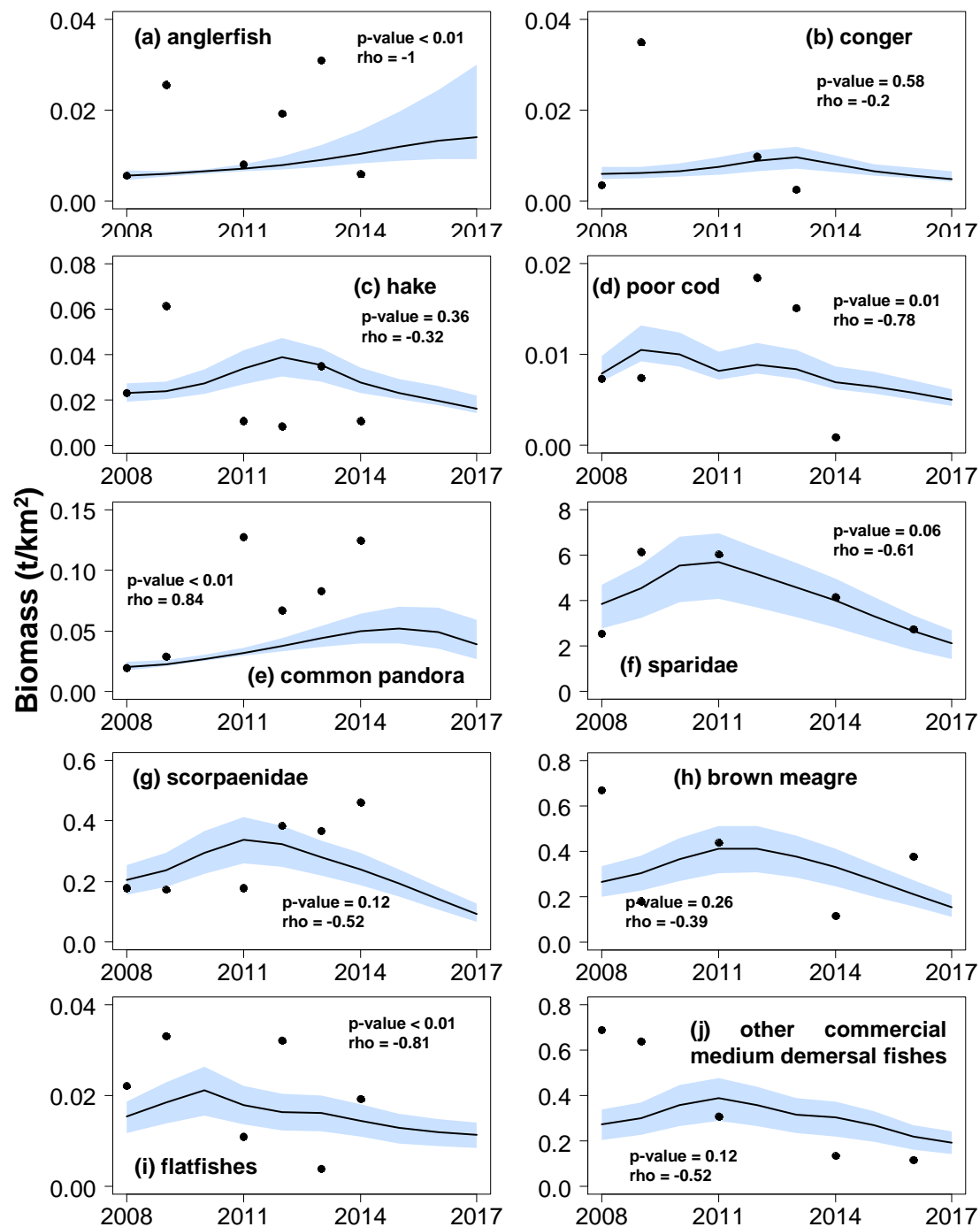
**Figure 1.** Predicted (solid lines) versus observed (dots) biomass ( $t \cdot km^{-2}$ ) values for the groups with available data for the Cerbère-Banyuls MPA model for the period 2013-2017. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.

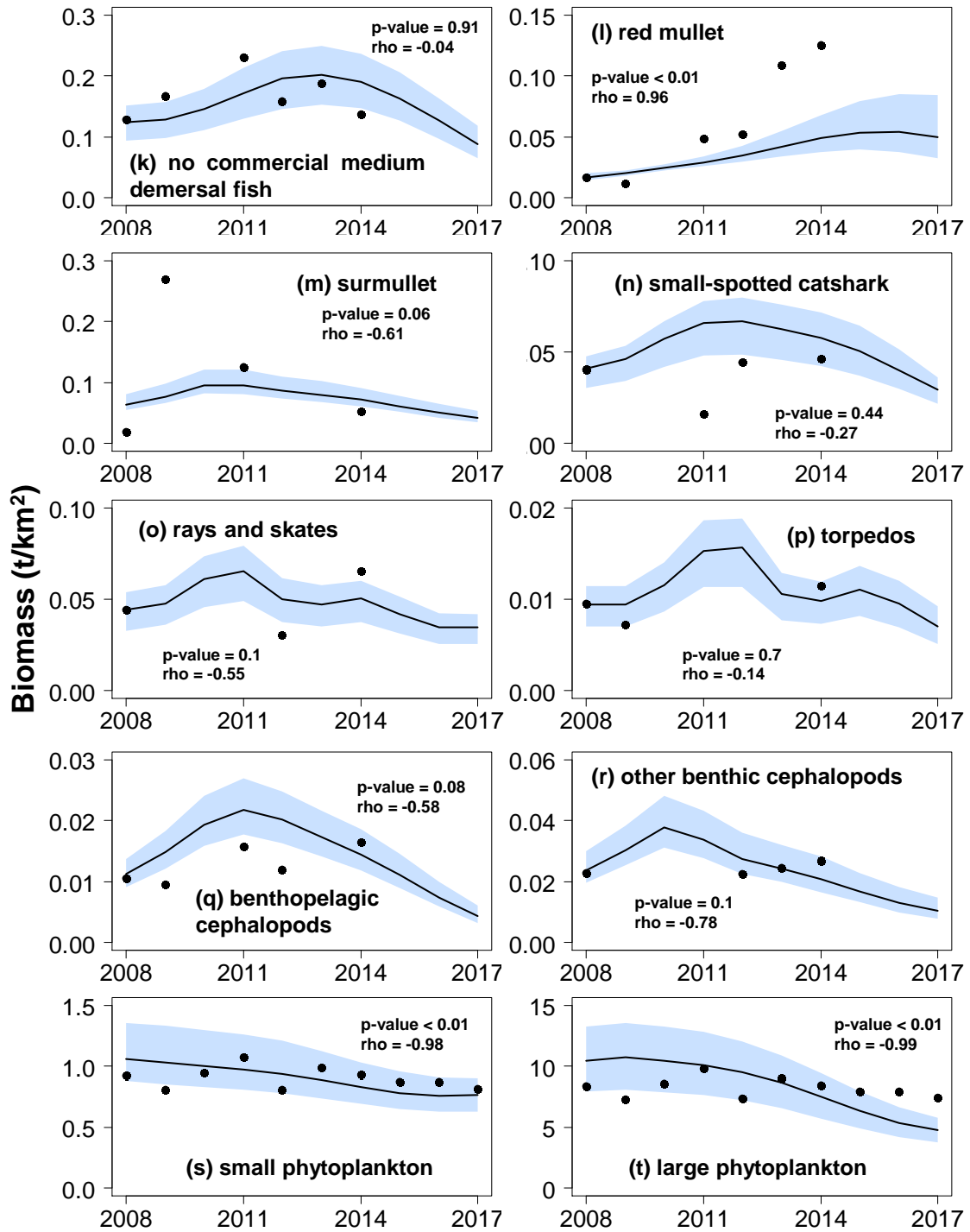




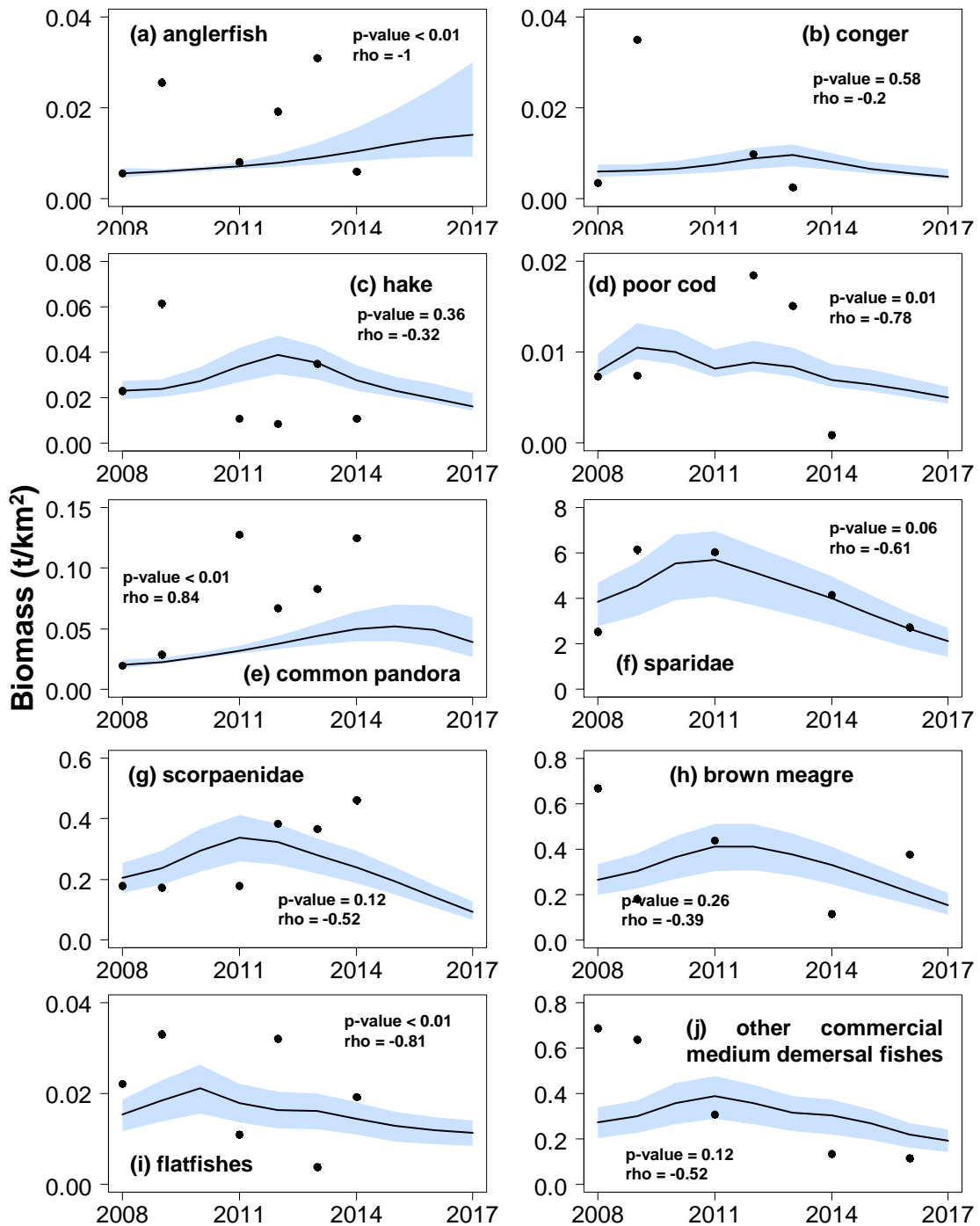


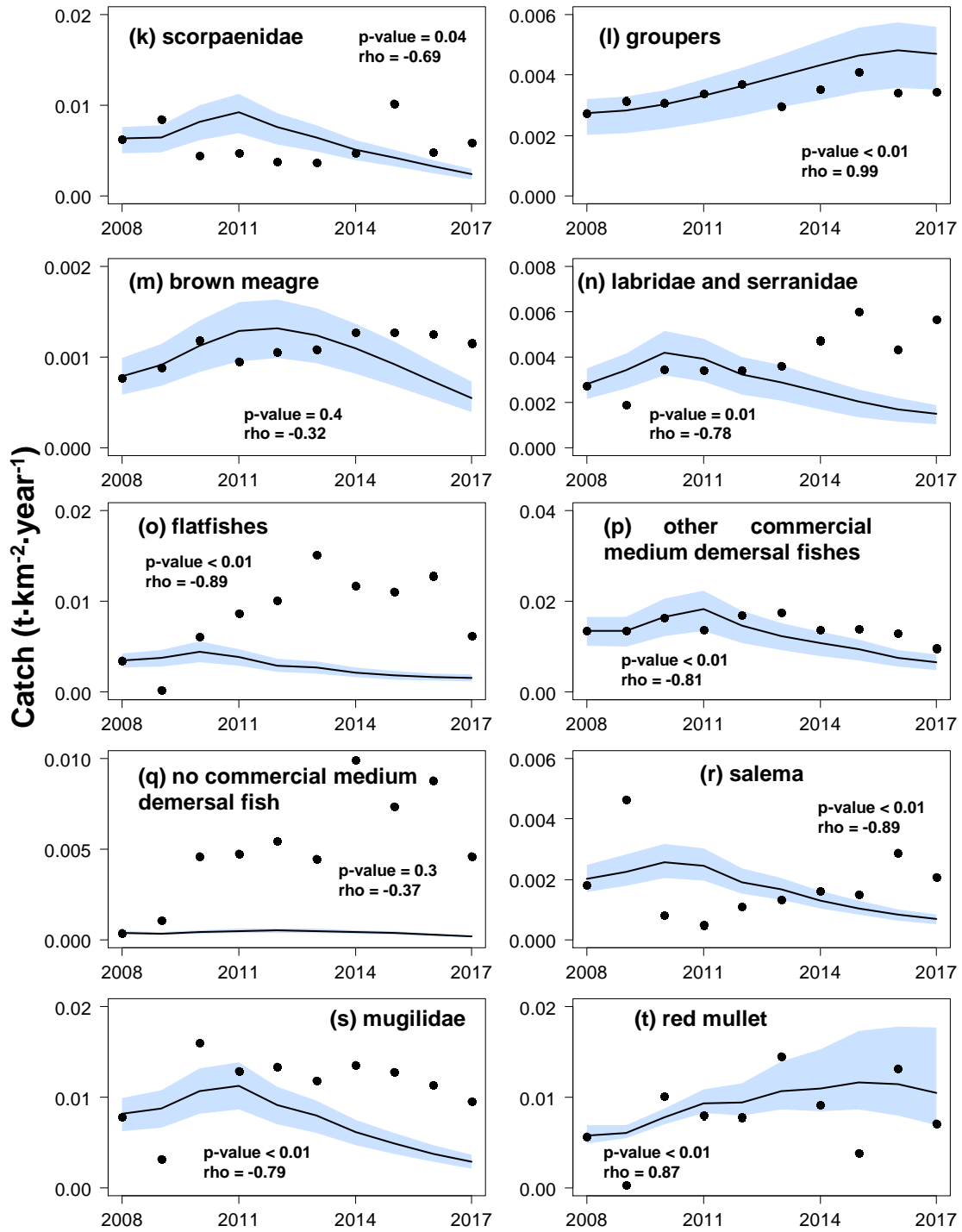
**Figure 1.** Predicted (solid lines) versus observed (dots) biomass ( $t \cdot km^{-2}$ ) values for the groups with available data for the Cap de Creus MPA model for the period 2008-2017. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.

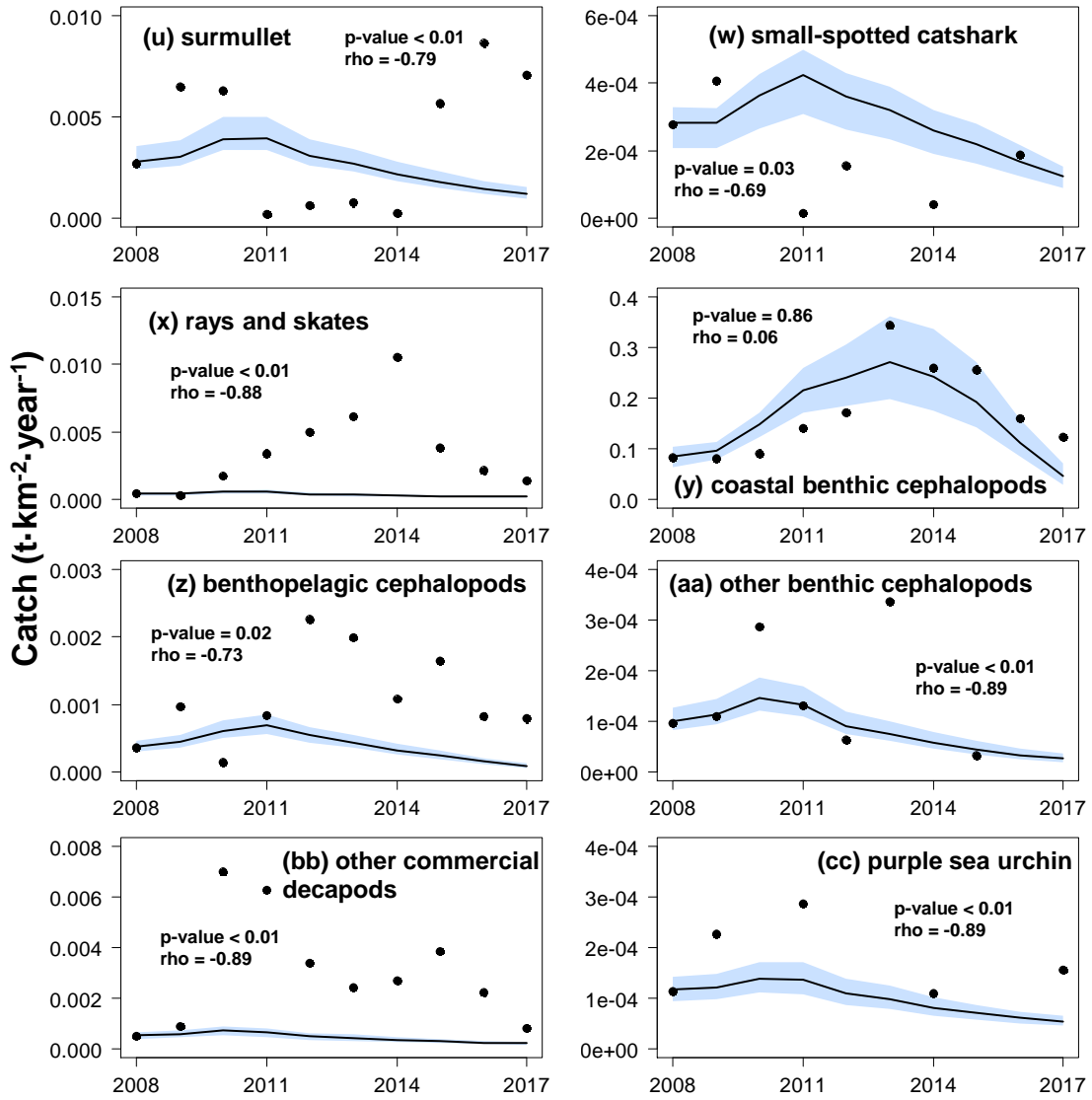




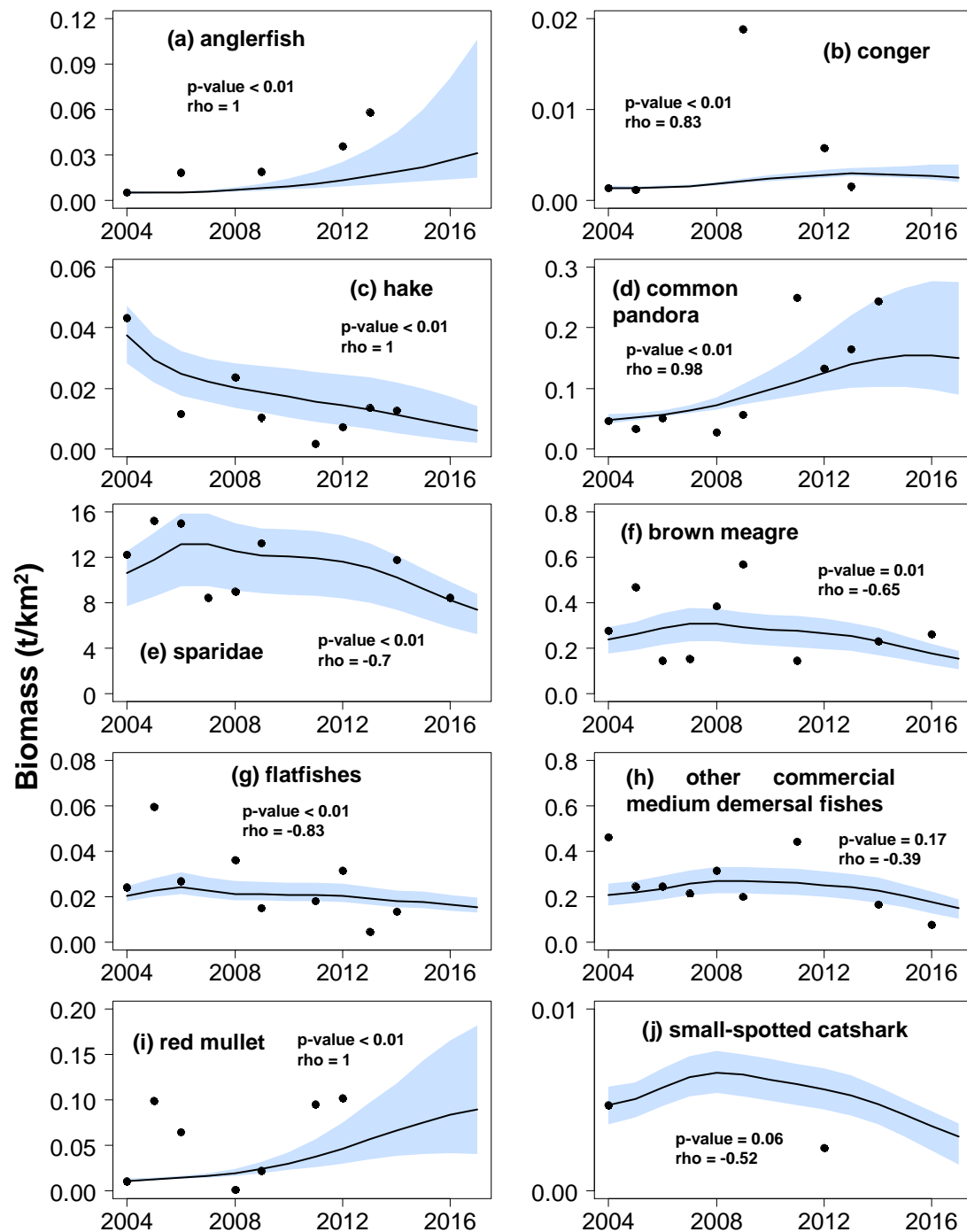
**Figure 2.** Predicted (solid lines) versus observed (dots) catch ( $t \cdot km^{-2} \cdot year^{-1}$ ) values for the groups with available data for the Cap de Creus MPA model for the period 2008-2017. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.

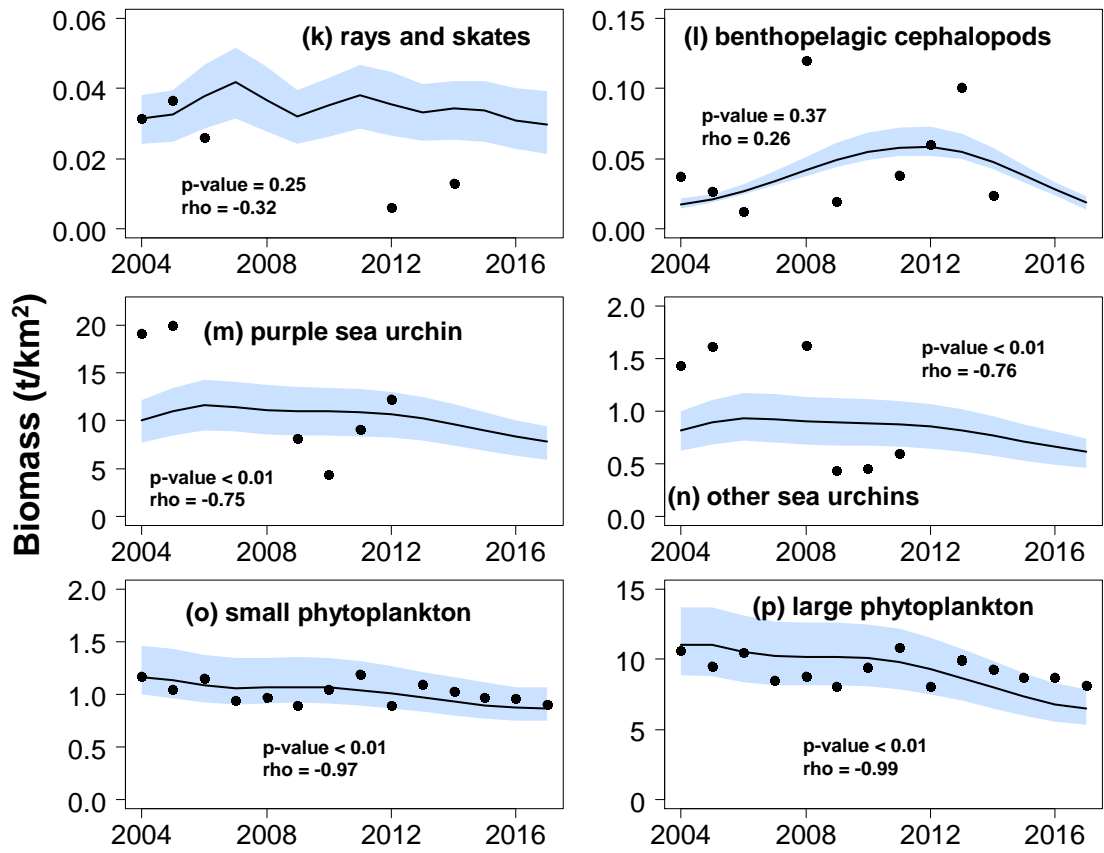






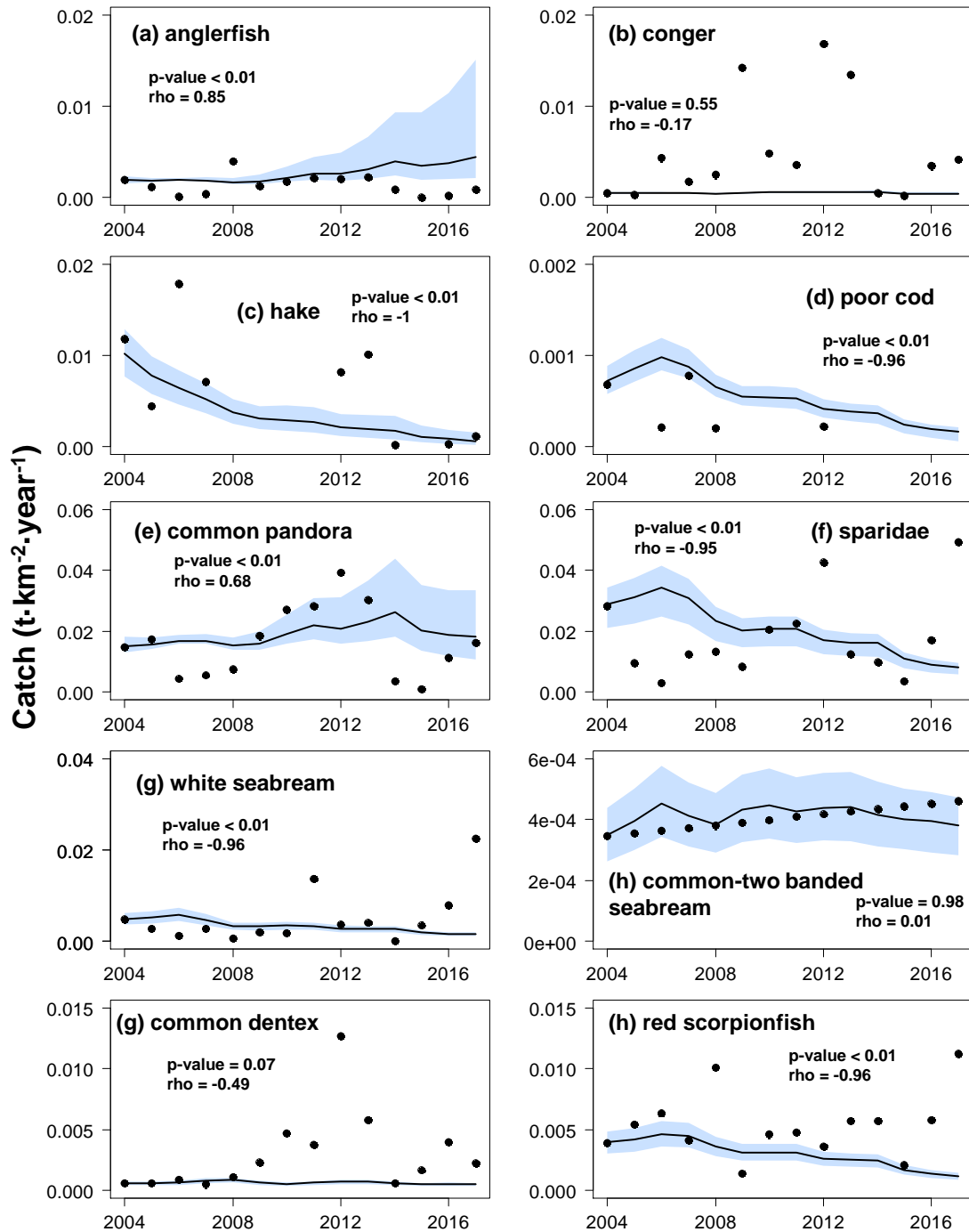
**Figure 1.** Predicted (solid lines) versus observed (dots) biomass ( $t \cdot km^{-2}$ ) values for the groups with available data for the Medes MPA model for the period 2004-2017. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.

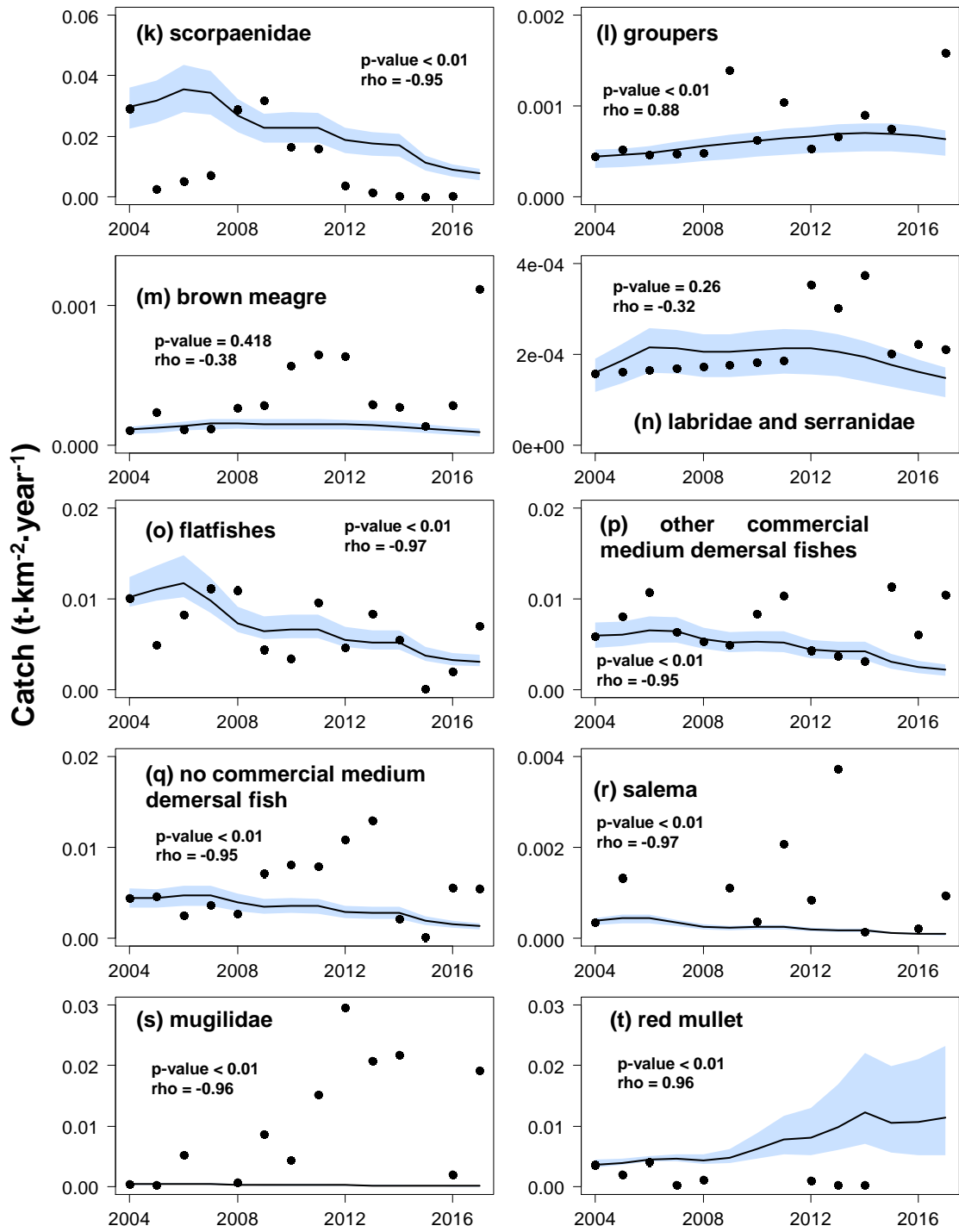


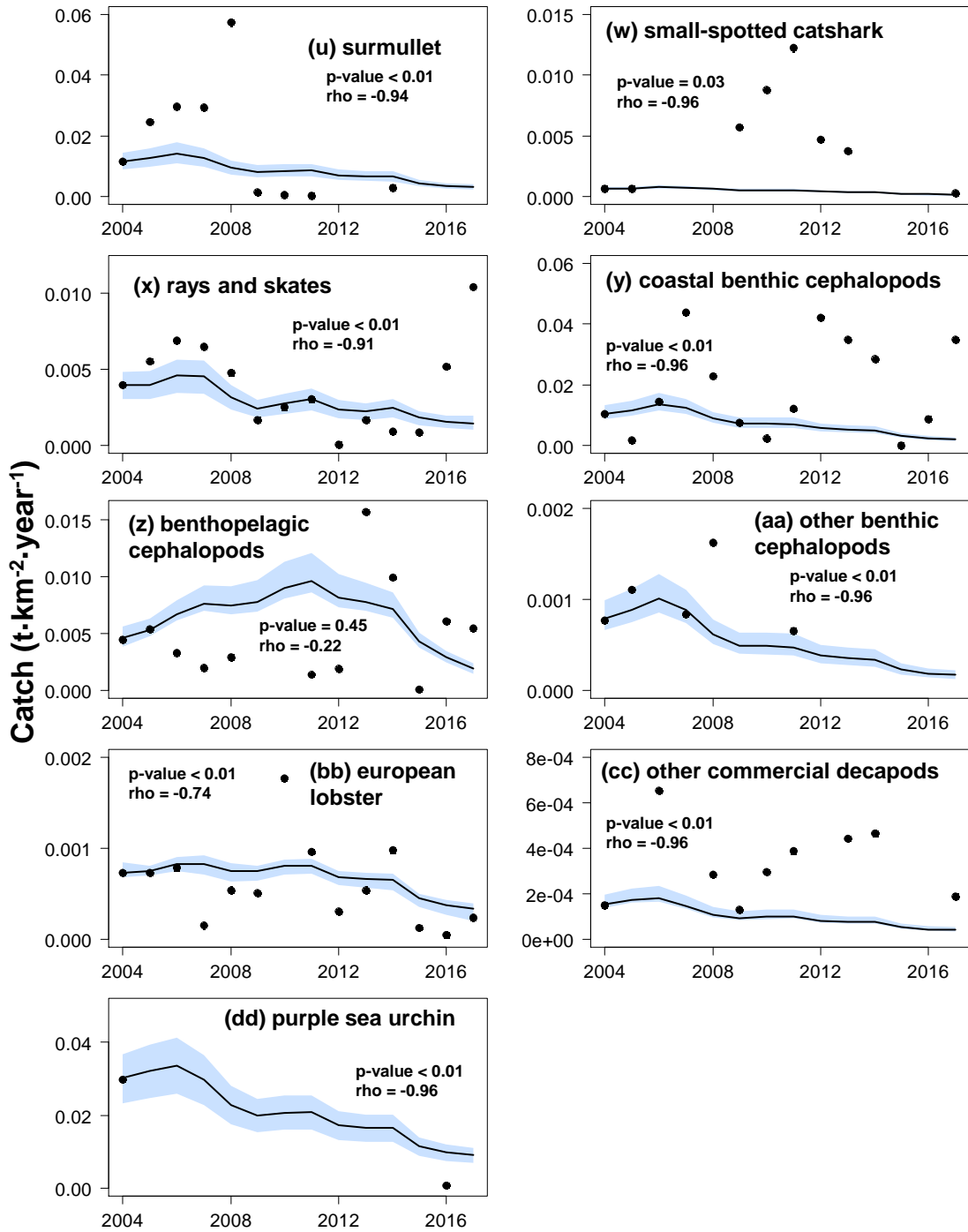




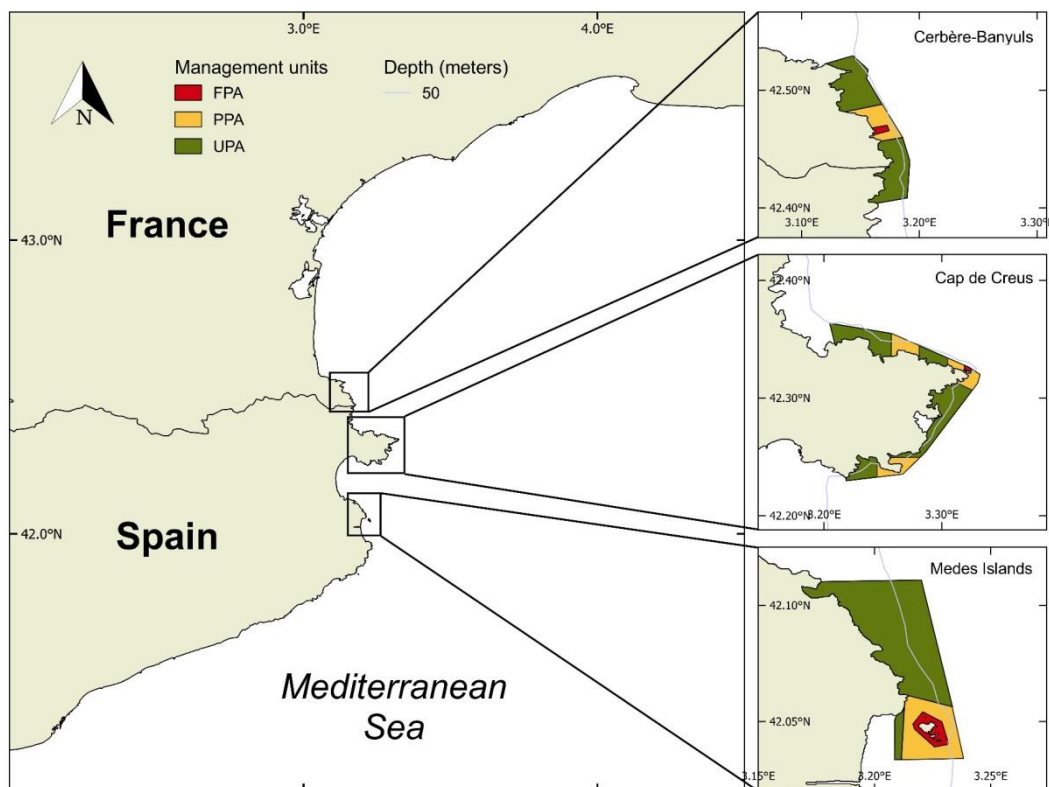
**Figure. 2.** Predicted (solid lines) versus observed (dots) catch ( $t \cdot km^{-2}$ ) values for the groups with available data for the Medes MPA model for the period 2004-2017. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.



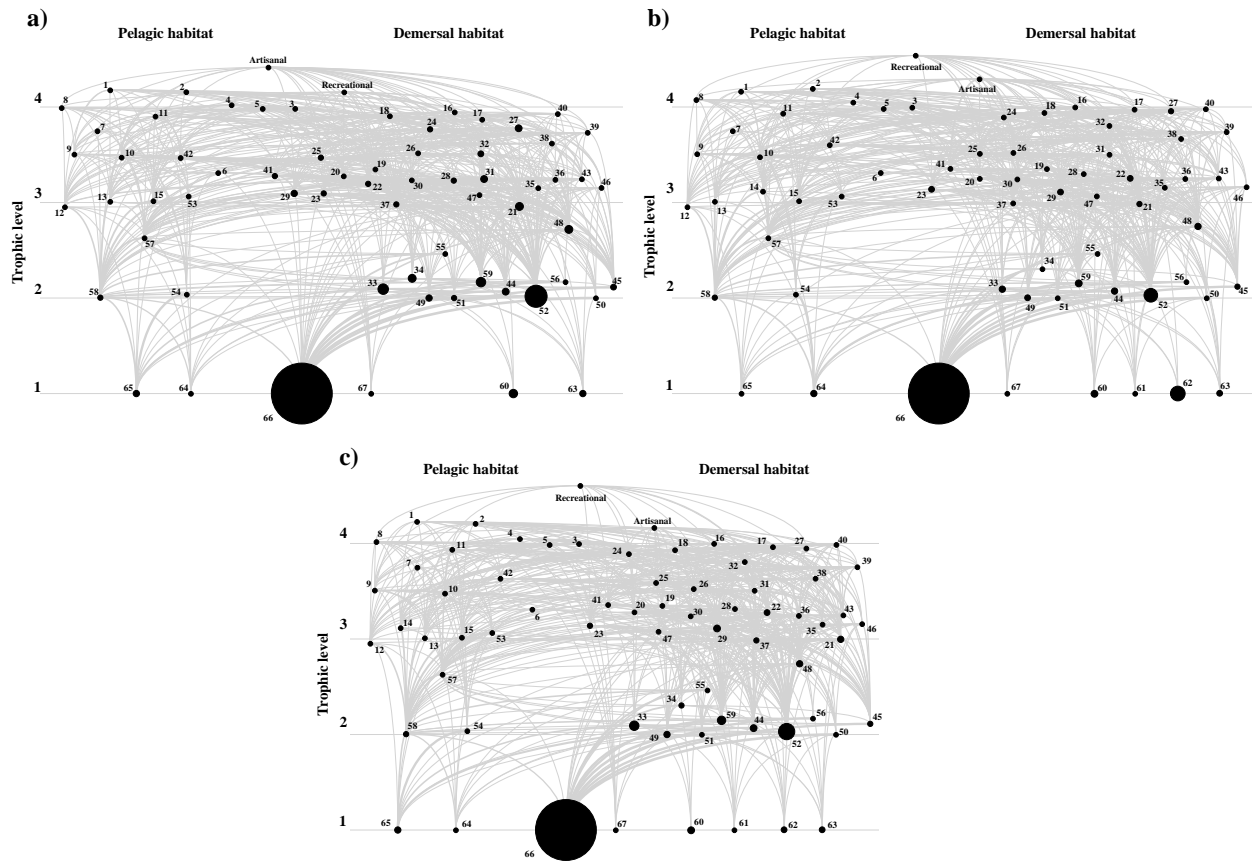




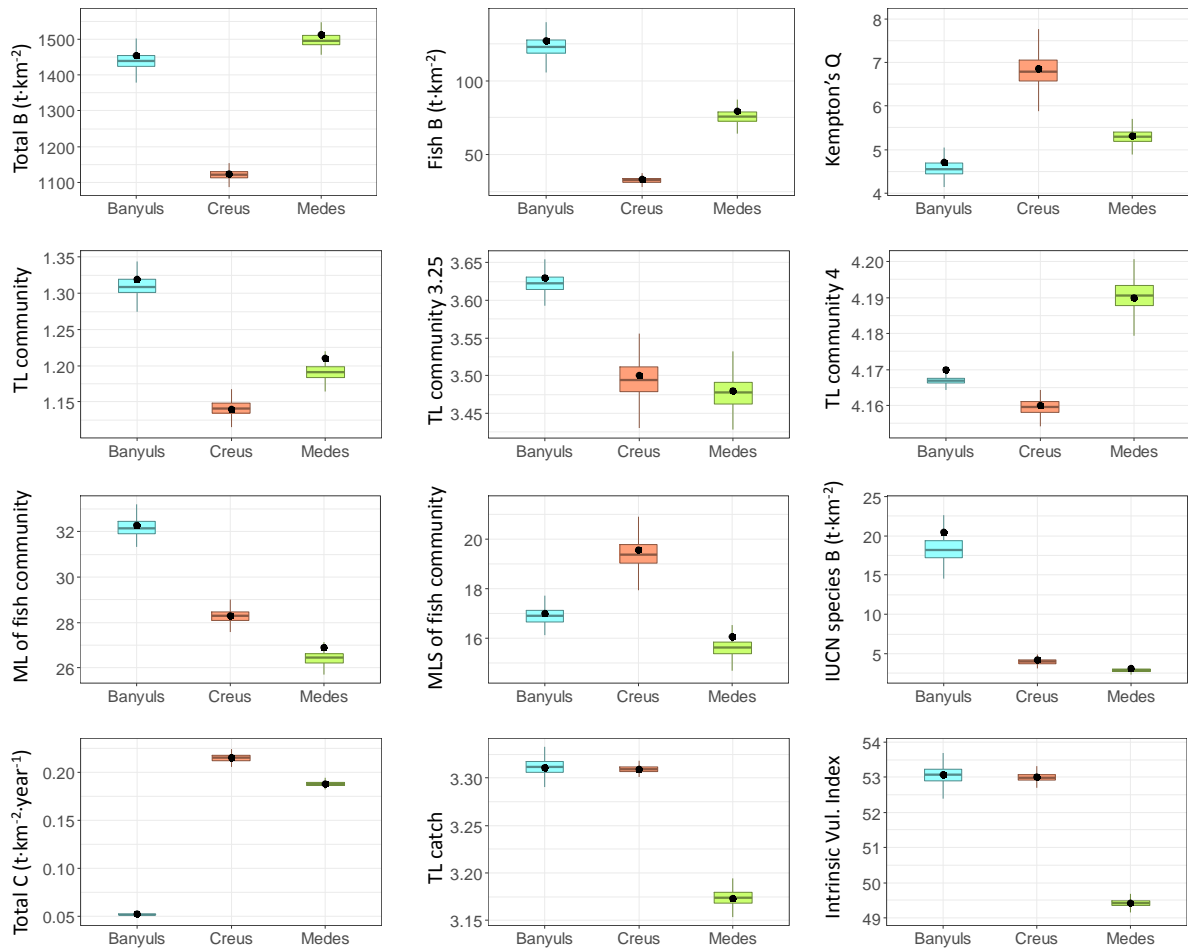
**Figure. 1.** The study area encompassing the three multi-zone MPAs in the Northwestern Mediterranean Sea with the fully protected areas (FPAs), partially protected areas (PPAs) and unprotected neighboring areas (UPAs).



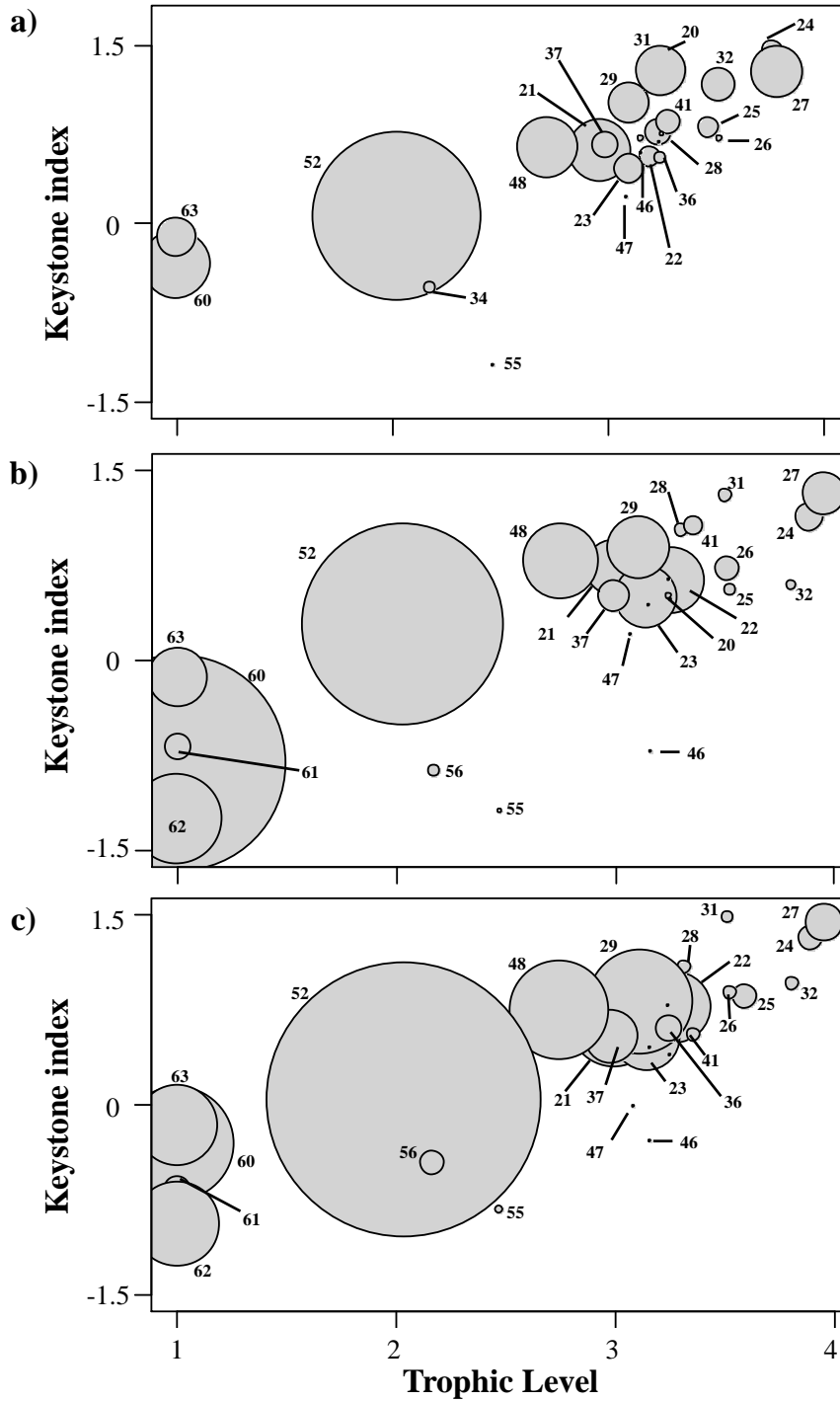
**Figure 2.** Flow diagram of Cerbère-Banyuls (a), Cap de Creus (b) and Medes Island (c) MPA models. The numbers identify the functional groups of the model (listed in SOM 2). The size of each circle is proportional to the biomass of the functional group. The thickness of the connecting lines is proportional to the magnitude of their trophic flows.



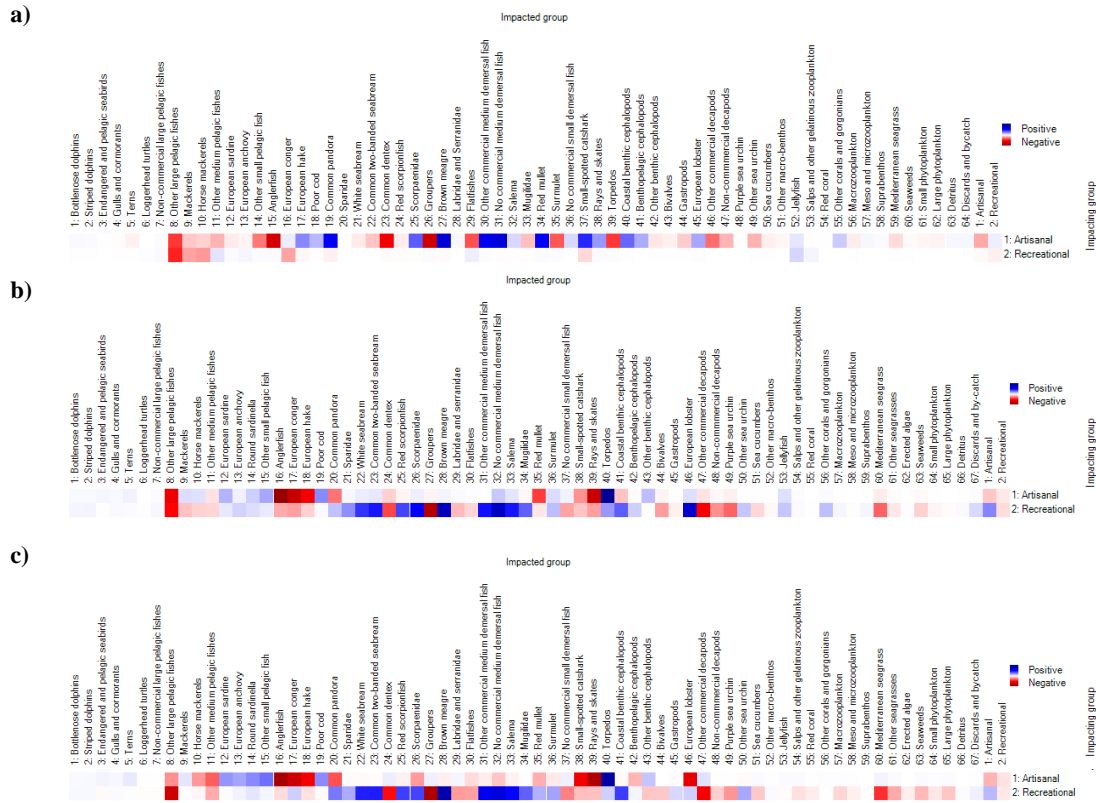
**Figure 3.** Ecological indicators estimated for the three multi-zone MPA models of Cerbère-Banyuls, Cap de Creus and Medes Islands. Boxplot shows the distribution of values for an ecological indicator derived from the Monte Carlo routine while the dot represents the value of the indicator in the baseline Ecopath balanced model.



**Figure 4.** Functional groups plotted against Keystone Index (KS) (Valls et al., 2015) and trophic level for Cerbère-Banyuls (a), Cap de Creus (b) and Medes (c) multi-zone MPA models. The numbers identify the functional groups of the model (listed in SOM 2). The size of the circles is proportional to the biomass of the functional group.

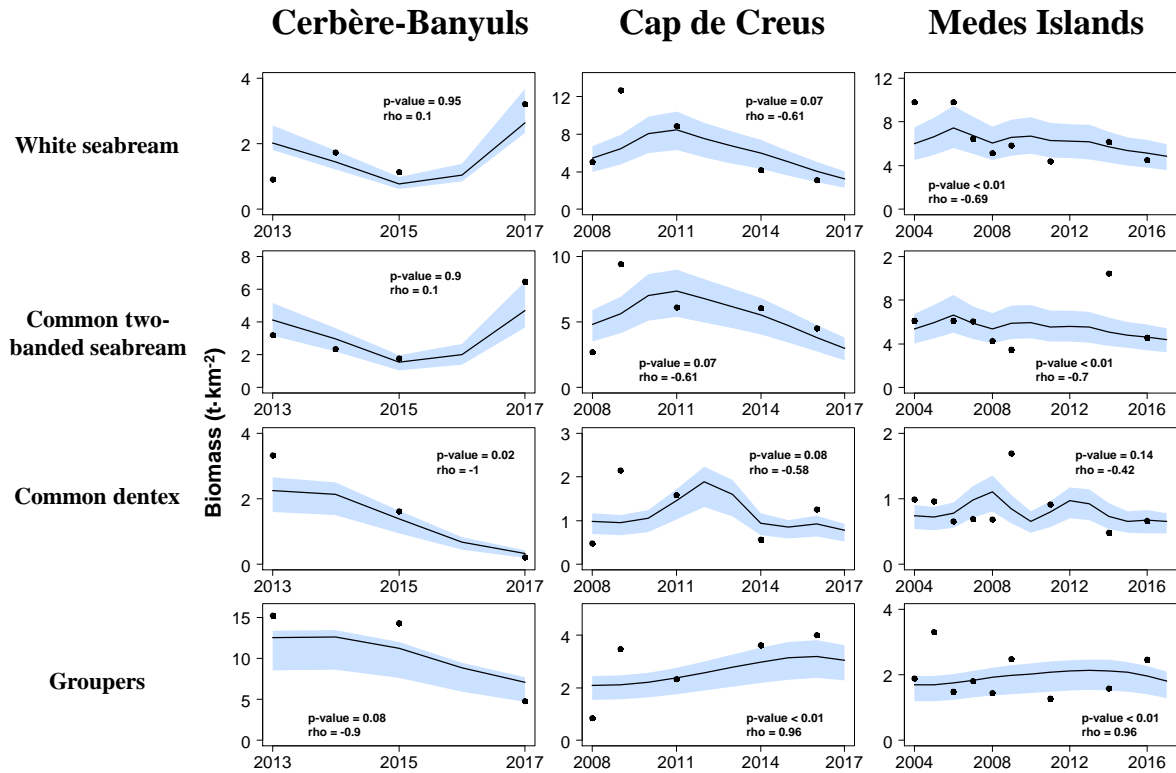


**Figure 5.** Mixed Trophic Impact (MTI) analysis of the three MPA applied to the fisheries in a) Cerbère-Banyuls, b) Cap de Creus, and c) Medes Islands multi-zone MPA models.

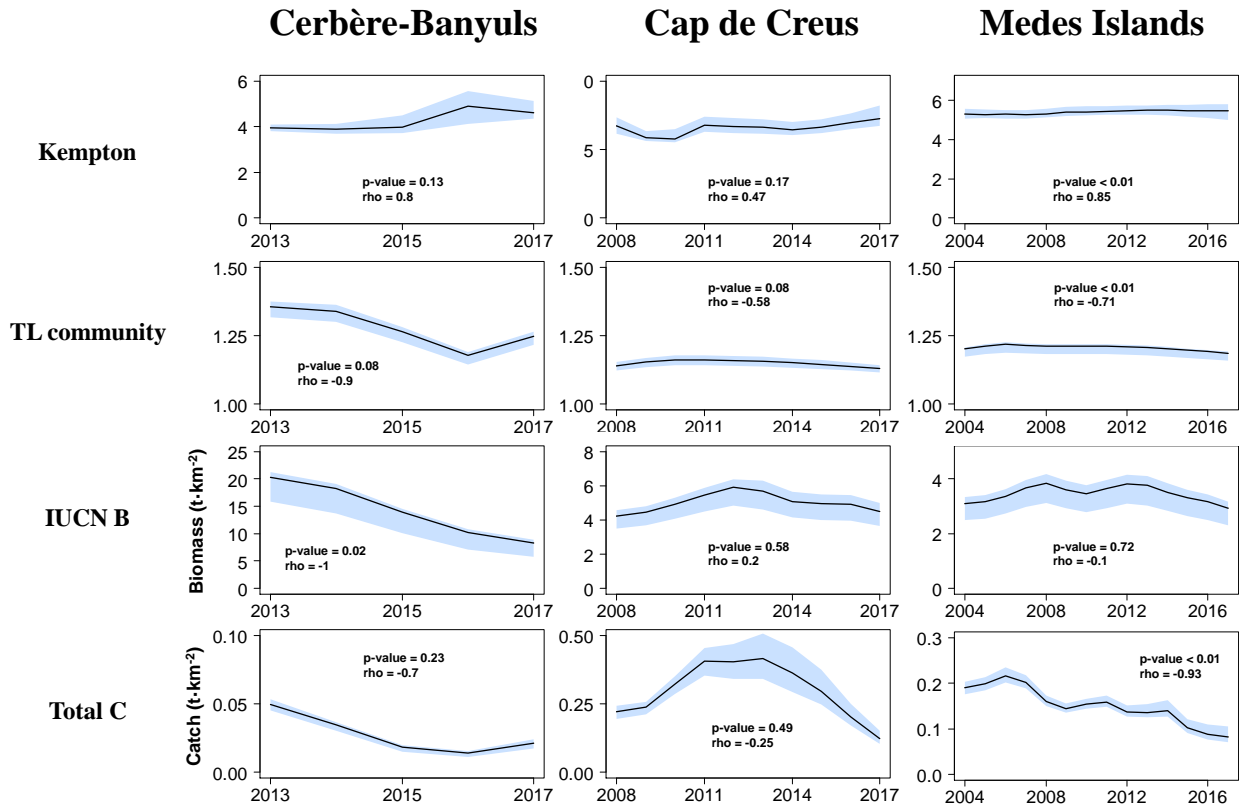




**Figure. 6.** Predicted (solid lines) versus observed (dots) biomass ( $t \cdot km^{-2}$ ) values for target species of Cerbère-Banyuls, Cap de Creus and Medes Islands multi-zone MPAs models. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.



**Figure. 7.** Temporal trends of ecological indicators of Cerbère-Banyuls, Cap de Creus and Medes Islands multi-zone MPAs models. Blue shadows represent the 5% and 95% percentiles obtained using the Monte Carlo routine. Rho and p-values result from Spearman's rank correlation test.





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**Table**

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