

1 **Reactivity of cellulose during hydrothermal carbonization of lignocellulosic**  
2 **biomass**

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17 **Abstract**

18 Hydrothermal carbonization of pure cellulose and birchwood samples was carried out at  
19 temperatures between 160 and 280 °C, 0.5 h residence time and biomass-to-water ratio  
20 of 20 wt% dry basis, to investigate HTC reactivity of cellulose naturally occurring  
21 lignocellulosic biomass. Pure cellulose samples remained unaltered at temperatures up  
22 to 220 °C, but significantly decomposed at 230 °C producing a thermal recalcitrant  
23 aromatic, high energy-dense material, showing lignin-like behavior. Fourier Transform  
24 Infrared spectroscopy (FTIR) showed dehydration and aromatization reactions  
25 occurring at temperatures equal or higher than 230 °C for pure cellulose samples while  
26 similar increase in aromatization for birchwood hydrochars was evident only at  
27 temperatures equal or higher than 260 °C. Acid hydrolysis, Thermogravimetric analysis  
28 (TGA) and FTIR suggest that a higher thermal resistance of natural occurring cellulose  
29 in birchwood (when compared to pure cellulose sample) could be related to a  
30 ‘protecting shield’ offered by interlinked lignin in the plant matrix.

31 **Keywords:** hydrothermal carbonization, solid biofuel, cellulose reactivity, birchwood,  
32 acid hydrolysis

33 **Highlights:**

34 HTC induces decomposition of pure cellulose at temperature higher than 220 °C

35 HTC promotes aromatization of cellulose at temperature equal or higher than 230 °C

36 Cellulose decomposition in biomass is mitigated by lignin component during HTC

37 **1. Introduction**

38 The unrestrainable growth of global energy demand together with the increasing  
39 environmental concern of using fossil fuel sources for energy production has prompted  
40 government authorities to issue new regulations and renewable energy share targets.  
41 EU-wide targets and policy objectives planned for 2030, within the 2030 climate and  
42 energy framework, include: at least 40% cuts in greenhouse gas emissions (from 1990  
43 levels), 32% share for renewable energy and 32.5% improvement in energy efficiency  
44 [1]. To reach the EU 2030 targets and, in particular, the renewable energy ones, in the  
45 last years the scientific community has boosted the study for the development of new  
46 technologies for the production of thermal and electrical energy using alternative  
47 renewable sources. Between the different renewable energy sources, biomass and, in  
48 particular, residual organic materials offer several advantages. Residual biomass is  
49 widely available in large amount at low cost, its conversion into an energy dense bio-  
50 fuel and/or valuable carbon material represents an opportunity to decrease the amount of  
51 waste with beneficial impact for the environment and human health [2]. Waste biomass  
52 exploitation could be highly economically profitable for waste management companies  
53 and represents a carbon-neutral and programmable energy source [3]. The conversion of

54 residual biomass into energy or valuable carbon-rich feedstock is affected by its high  
55 moisture content, its perishability and low energy density. Anaerobic digestion (AD) of  
56 wet residual biomass to produce methane-rich biogas represents one of the possible and  
57 eco-sustainable route for increasing the renewable energy production while decreasing  
58 green-house gas emissions [4]. However, the need for high initial investment costs and  
59 the requirement for strict operating conditions has limited the widespread of AD, while  
60 most of the operating plants are nowadays surviving owing mainly to economic  
61 incentives [5]. Thermochemical technologies applied to conversion of residual  
62 biomass such as torrefaction [6–8], slow and fast pyrolysis [9–11], gasification [12] or a  
63 combination of these technologies [13] have been largely investigated in the last  
64 decades but their widespread diffusion mainly failed due to the low energy efficiency,  
65 especially for high moisture content biomass, and the low versatility due to the need of  
66 very specific operating conditions strictly related to the nature, morphology and  
67 physical and chemical composition of feedstock. In the more recent years, wet  
68 thermochemical conversion of biomass is attracting more and more interest among  
69 scientists and technology developers. Wet pyrolysis, also known as hydrothermal  
70 carbonization (HTC), is carried out in water at sub-critical conditions, typically between  
71 180 and 280 °C and at autogenous vapor pressure (10-60 bar). Water, at HTC reaction  
72 conditions, promotes dehydration and decarboxylation of biomass, converting wet  
73 residual organic material into a carbon-rich solid material, named hydrochar [14–16].  
74 Comparative studies between dry thermochemical conversion and HTC of waste  
75 biomass showed that the latter could be more energetically favorable, promoting higher  
76 degree of carbonization of feedstock at same reaction temperatures [17]. Moreover,  
77 hydrochars display significantly better energy and fuel qualities than the corresponding

78 pyrochars obtained at the same temperatures [18,19]. Since the rediscovery of wet  
79 thermochemical treatment of biomass as a valuable process for CO<sub>2</sub> sequestration and  
80 production of renewable solid biofuels [20–22], HTC has been used to convert many  
81 kinds of waste biomass: lignocellulosic material as olive mill industry wastes [23]  
82 loblolly pine [24]; agro-waste such as: tomato peel [25], orange waste [26], wheat straw  
83 [27], food waste [28], organic fraction of municipal solid waste [29], paper mill industry  
84 wastes [30–32], sewage sludge [33,34] and plastic wastes [35]. Despite the different  
85 nature of the treated feedstock, most of the HTC works focused on the influence of the  
86 operative variables like temperature, residence time and biomass to water ratio on the  
87 energy, chemical and morphological properties of the produced hydrochars for their  
88 possible applications as solid bio-fuels and valuable carbon materials (e.g. activated  
89 carbons). According to our current knowledge, very few works attempted at describing  
90 the evolution of biomass chemical structure during HTC [36], some studies reported the  
91 reactivity of biomass macro-components during HTC albeit starting from commercially  
92 available single components and none of them studied how biomass macro-constituents  
93 reacted and/or interacted when intermeshed in lignocellulosic matrix [37–39].  
94 Systematically larger char yields were observed from the pyrolysis of chemically  
95 isolated lignin, compared to expected yields from the pyrolysis of lignin embedded in  
96 plant material thus demonstrating that an entirely different reaction pathway is involved  
97 when the constituents are embedded in plant material [40]. Nevertheless, unveiling the  
98 evolution of reactivity of biomass macro-components during HTC is of great  
99 importance to predict the properties of produced hydrochars. This work represents the  
100 first study that uses TGA, FTIR and acid hydrolysis analysis, to prove that naturally  
101 occurring lignin component interconnected in the biomass matrix could play a role in

102 increasing cellulose thermal resistance during HTC thus improving energy properties of  
103 hydrochars.

## 104 **2. Materials and Methods**

### 105 *2.1 Material preparation and hydrothermal carbonization*

106 HTC of pure cellulose (Sigma Aldrich 50  $\mu\text{m}$ ) and dried birchwood, milled with a Retsch  
107 SM2000 (Retsch GmbH) and sieved to grain size  $\leq 1$  mm, was carried out in an unstirred  
108 50 ml batch reactor at 160, 180, 200, 220, 230, 240, 260 and 280  $^{\circ}\text{C}$ , fixed 0.5 h residence  
109 time and biomass-to-water ratio of 20 wt% on a dry basis. About  $6.00 \pm 0.05$  g of dry  
110 feedstock was loaded into the reactor and  $30 \text{ g} \pm 0.5$  g of deionized water was added to it.  
111 The mixture was carefully mixed, the reactor sealed, purged with pure nitrogen and heated  
112 up to the set temperature (temperature increment of about 8-10  $^{\circ}\text{C}/\text{min}$ ) and left for the  
113 set residence time. At the end of the reaction time the system was quenched by placing  
114 the reactor on a large stainless steel disk kept at  $-30$   $^{\circ}\text{C}$  and by blowing compressed air to  
115 the reactor's walls. Once the system had reached a temperature of 30  $^{\circ}\text{C}$ , the reactor's  
116 outlet valve was opened and reaction gas collected in a graduated cylinder, previously  
117 filled with water, to evaluate the produced gas volume. The gas mass yield was then  
118 calculated assuming that gas was composed only of  $\text{CO}_2$ . Hydrochar was then collected  
119 by filtration and the solid residue dried in a conventional ventilated oven for at least 12  
120 hours. Dried hydrochars were stored in sealed glass vials for further analysis and  
121 characterization.

### 122 *2.2 Analytical determination and characterization*

123 The raw materials and hydrochars were subjected to acid hydrolysis to determine the  
124 composition in macro-constituents. High Heating Value (HHV), elemental and proximate

125 composition in terms of Volatile Matter (VM), Fixed Carbon (FC) and Ash content and  
126 attenuated Total Reflectance (ATR)\_FTIR spectroscopy were also performed.

127 For acid hydrolysis the samples were first extracted with acetone according to the  
128 guidelines of SCAN-CM 49:03. 250 mL acetone was used in a Soxhlet apparatus with 1  
129 g of sample for 2 hours to guarantee removal of extractives and oils remaining on char  
130 particles. The monosaccharide and lignin contents of the extractive-free samples were  
131 then determined based on the National Renewable Energy Laboratory (NREL)  
132 procedure for determination of structural components in biomass [41]. Sugar recovery  
133 standards were prepared from analytical grade D-(+)-glucose, D-(+)-xylose, D-(+)-  
134 galactose, D-(+)-mannose, L-(+)-arabinose and L-(+)-rhamnose. Hydrolysed monomers  
135 were quantified after filtration based on respective peak areas using a Dionex ICS-3000  
136 ion chromatograph (Dionex Corp.) and corrected to respective polymeric forms on a  
137 dried, as-received basis [41]. Lignin contents were determined as the sum of  
138 gravimetrically determined acid-insoluble lignin and acid-soluble lignin, which was  
139 quantified with a Shimadzu UV-2550 spectrophotometer (Shimadzu Corp.) at 205 nm.  
140 All acetone extractions and subsequent sugar and lignin determinations were performed  
141 in duplicate with overall recoveries of 87-107% on a mass basis and a replicate root  
142 mean squared error of 1.4% within a range of 0-99% glucan in the samples.

143 FTIR was carried out on a series of samples using a Perkin Elmer Spectrum 400 FT-  
144 IR/NIR spectrometer (Perkin Elmer Inc., Tres Cantos, Madrid) in mid-IR mode,  
145 equipped with a Universal ATR sampling device containing diamond/ZnSe crystal. The  
146 spectra were recorded in the range from 650 to 4000  $\text{cm}^{-1}$ , with a resolution of 4  $\text{cm}^{-1}$ ,  
147 by averaging 16 scans, spectra were baseline corrected and normalized.

148 The spectra were interpreted using a bilinear principal component analysis (PCA) model  
149 based on absorbance units. The normalized transmittance IR spectra (Fig. S1) were first  
150 converted into absorbance, further corrected for baseline offsets using the signals within  
151 3600-4000  $\text{cm}^{-1}$  and then mean centered for PCA (Fig. S2). The PCA results were given  
152 through sample scores on the orthogonal principal components (PCs) and the respective  
153 changes in the IR spectra illustrated through orthonormal PC loadings. The interested  
154 reader is referred to the published literature on further details on the PCA method  
155 [42,43].

156 Ultimate analyses were performed using a LECO 628 analyser equipped with Sulphur  
157 module for CHN (ASTM D-5373 standard method) and S (ASTM D-1552 standard  
158 method) content determination.

159 The HHV of solid samples was evaluated according to the CEN/TS 14918 standard by  
160 means of an IKA C 200 calorimeter. Hydrochar mass yield ( $S_y$ ) was calculated  
161 according to Eq. (1):

$$162 \quad S_y = M_{HCdb}/M_{Rdb} \quad (1)$$

163 where  $M_{HCdb}$  represents the mass (dry basis) of the solid remaining after thermal  
164 treatment (*i.e.* hydrochar), and  $M_{Rdb}$  represents the mass (dry basis) of the raw sample  
165 before thermal treatment. Similarly, gas mass yield ( $G_y$ ) was defined as the mass of gas  
166 produced per unit mass of dry raw biomass sample. Liquid mass yield ( $L_y$ ) was  
167 calculated as the complement to 1 of the sum of  $S_y$  and  $G_y$ . The energy densification  
168 ratio (EDR) was calculated according to Eq. (2), where  $HHV_{HCdb}$  represents the higher  
169 heating value of hydrochar on a dry basis and  $HHV_{Rdb}$  represents the higher heating  
170 value of the raw material. Energy yield (EY) was calculated according to Eq. (3).

171  $EDR = HHV_{HCdb}/HHV_{Rdb}$  (2)

172  $Ey = EDR * Sy$  (3)

173 Proximate analysis was carried out by means of TGA using a TA Instruments Q500  
174 TGA. Between 2 and 20 mg of sample were placed in a 100  $\mu$ L platinum sample pan,  
175 held at room temperature under high purity nitrogen at 60 mL/min (with an additional  
176 40 mL/min balance protective gas) for 15 minutes, and then heated at 15  $^{\circ}$ C/min to 105  
177  $^{\circ}$ C and held for 20 minutes at this temperature to remove moisture. Samples were  
178 further heated at 15  $^{\circ}$ C/min to 900  $^{\circ}$ C and held for 7 minutes, and then cooled at 15  
179  $^{\circ}$ C/min to 450  $^{\circ}$ C; the total mass loss from 105  $^{\circ}$ C through to this point as a portion of  
180 the total sample mass on a dry ash-free basis is taken as VM. Samples were then heated  
181 up at 15  $^{\circ}$ C/min to 750  $^{\circ}$ C in an equivalent flow of air, and held at this temperature for  
182 15 min; the remaining mass as a portion of sample mass on a dry basis is taken as ash  
183 content. FC on a dry ash-free basis was then calculated by subtracting Ash and VM  
184 from the dry mass.

185 Sixteen different experiments (eight for cellulose and eight for birchwood) were carried  
186 out at least in triplicate and average results are reported in tables and figures. cellulose  
187 and birchwood hydrochars produced at different HTC temperature were coded CE\_T  
188 and BW\_T respectively being “T” the value of process temperature expressed in  
189 degrees Celsius.

### 190 **3. Results**

#### 191 *3.1 Mass yields and hydrochar energy properties*

192 The summary of the results of cellulose and corresponding hydrochars in terms of mass  
193 yields, energy properties (HHV, EY and EDR) and elemental, proximate analysis, H/C



194 and O/C atomic ratios are reported in tables 1 and 2 respectively. The data show on the  
 195 one hand that cellulose did not appreciably degrade during HTC at temperatures lower  
 196 than 220 °C, on the other hand, that at 230 °C cellulose underwent significant changes  
 197 in mass yields and energy properties. When HTC temperature was increased from 220  
 198 to 230 °C, hydrochar mass yield dropped from 82.4 to 55.4 wt%, HHV increased from  
 199 17.2 MJ/kg to 21.8 MJ/kg, fixed carbon from 3.7 to 42.0 wt%, carbon content increased  
 200 from 45.8 to 58.8 wt%. More evident changes in cellulose hydrochar composition  
 201 occurred when temperature was raised to 240 °C, whereby HHV increased to 26.8  
 202 MJ/kg, however changes were reduced when further rising HTC temperature to 280 °C  
 203 (HHV = 27.5 MJ/kg). The high changes in energy content and composition could be  
 204 related to the breaking of the beta-glucosidic covalent bond that reportedly occurs at  
 205 around 230 °C [36].

206 Table 1 – Cellulose HTC mass yields (wt%, d.b.), energy yields EY and EDR (%) d.b,  
 207 HHV (MJ/kg) (Mass yields standard deviations < 1.5, HHV standard deviations < 0.05).

| <b>Sample</b> | <b>SY</b> | <b>GY</b> | <b>LY*</b> | <b>HHV<sub>HC</sub></b> | <b>EY</b> | <b>EDR</b> |
|---------------|-----------|-----------|------------|-------------------------|-----------|------------|
| <b>CE_raw</b> | -         | -         | -          | 16.90                   | 100.0     | 100.0      |
| <b>CE_160</b> | 98.3      | 0.1       | 1.6        | 17.03                   | 99.0      | 100.8      |
| <b>CE_180</b> | 97.5      | 0.2       | 2.3        | 16.96                   | 97.8      | 100.4      |
| <b>CE_200</b> | 95.5      | 0.2       | 4.2        | 16.95                   | 95.8      | 100.3      |
| <b>CE_220</b> | 82.4      | 0.5       | 17.1       | 17.23                   | 84.0      | 101.9      |
| <b>CE_230</b> | 55.4      | 2.0       | 42.6       | 21.76                   | 71.4      | 128.8      |
| <b>CE_240</b> | 51.2      | 5.4       | 43.4       | 26.76                   | 81.1      | 158.3      |
| <b>CE_260</b> | 50.9      | 7.8       | 41.3       | 27.03                   | 81.4      | 159.9      |
| <b>CE_280</b> | 49.2      | 9.0       | 41.9       | 27.46                   | 79.9      | 162.5      |

208 \*Calculated by difference (LY=100-SY-GY)

209 As shown in table 2, VM, FC, ash content, carbon and hydrogen composition of  
 210 cellulose hydrochars remained almost unaltered up to 220 °C. This corroborates the

211 mass yield and energy properties results. The sharp decrease of VM and corresponding  
 212 increase of FC and C content at 230 °C confirm that a significant change in cellulose  
 213 chemical composition occurred at that process temperature. Data also demonstrate that  
 214 when further increasing HTC temperature, FC and C contents slightly increase while H  
 215 content remains approximately constant or slowly decreases as commonly reported in  
 216 HTC literature [25,29].

217 Table 2 – Cellulose raw and hydrochar proximate and elemental analyses. All values  
 218 except from H/C and O/C atomic ratios in wt% d.b. (standard deviations for proximate  
 219 and ultimate analysis < 1.2 and 0.2, respectively).

| Sample | VM   | FC   | ASH  | C    | H   | N   | O*   | H/C  | O/C  |
|--------|------|------|------|------|-----|-----|------|------|------|
| CE_raw | 97.7 | 2.4  | 0.0  | 45.1 | 5.5 | 0.0 | 49.4 | 1.44 | 0.82 |
| CE_160 | 97.9 | 2.1  | 0.0  | 45.1 | 5.5 | 0.0 | 49.4 | 1.45 | 0.82 |
| CE_180 | 97.8 | 2.2  | 0.0  | 45.2 | 5.5 | 0.0 | 49.3 | 1.45 | 0.82 |
| CE_200 | 97.6 | 2.4  | 0.0  | 45.2 | 5.5 | 0.0 | 49.3 | 1.45 | 0.82 |
| CE_220 | 96.3 | 3.7  | 0.0  | 45.8 | 5.5 | 0.0 | 48.7 | 1.42 | 0.80 |
| CE_230 | 58.0 | 42.0 | 0.0  | 58.8 | 4.6 | 0.0 | 36.6 | 0.94 | 0.47 |
| CE_240 | 58.0 | 42.0 | 0.0  | 70.5 | 3.9 | 0.1 | 25.5 | 0.67 | 0.27 |
| CE_260 | 52.8 | 47.2 | 0.0  | 72.2 | 3.9 | 0.1 | 23.8 | 0.65 | 0.25 |
| CE_280 | 52.6 | 47.4 | 0.3% | 73.2 | 4.0 | 0.1 | 22.4 | 0.65 | 0.23 |

220 \*Calculated by difference O =100-Ash-C-H-N

221 HTC birchwood mass yields and energy properties results are shown in table 3 and  
 222 proximate and elemental analysis are shown in table 4. Unlike pure cellulose  
 223 hydrochars, birchwood hydrochars do not show sharp changes at any specific  
 224 temperature. Mass yields, HHV, C and FC contents changed gradually with increasing  
 225 reaction temperature. Significant body of evidence exists in the literature to demonstrate  
 226 that, during hydrothermal reaction of lignocellulosic biomass, hemicellulose is more  
 227 reactive than cellulose while lignin component is quite recalcitrant to degradation  
 228 [37,44].

229 Table 3 – Birchwood HTC mass yields and energy properties (all values in wt% d.b.,  
 230 mass yields standard deviations < 0.9, HHV standard deviations < 0.3).

| Sample        | SY   | GY  | LY*  | HHV <sub>HC</sub> | EY    | EDR   |
|---------------|------|-----|------|-------------------|-------|-------|
| <b>BW_raw</b> | -    | -   | -    | 18.98             | 100.0 | 100.0 |
| <b>BW_160</b> | 92.0 | 0.3 | 7.7  | 18.73             | 90.8  | 98.7  |
| <b>BW_180</b> | 85.4 | 0.9 | 13.7 | 19.06             | 85.8  | 100.4 |
| <b>BW_200</b> | 70.9 | 1.6 | 27.5 | 20.18             | 75.4  | 106.3 |
| <b>BW_220</b> | 66.0 | 2.3 | 31.7 | 21.35             | 74.3  | 112.5 |
| <b>BW_230</b> | 62.3 | 3.6 | 34.2 | 22.69             | 74.4  | 119.5 |
| <b>BW_240</b> | 57.7 | 4.1 | 38.2 | 24.40             | 74.2  | 128.6 |
| <b>BW_260</b> | 53.3 | 7.4 | 39.3 | 27.05             | 75.9  | 142.5 |
| <b>BW_280</b> | 52.1 | 8.6 | 39.2 | 27.89             | 76.6  | 146.9 |

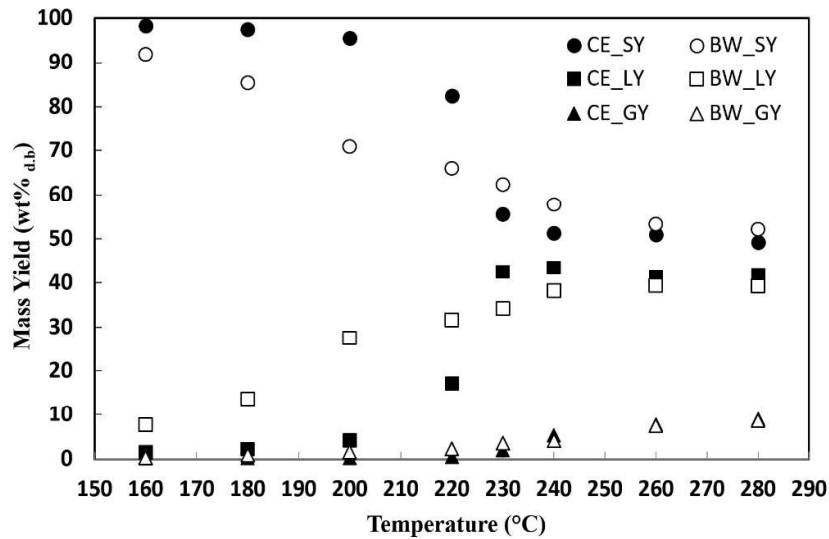
231 \*Calculated by difference (LY=100-SY-GY)

232 Table 4 – Birchwood raw and hydrochar proximate and elemental analyses. All values  
 233 apart from H/C and O/C atomic ratios in wt% d.b. (standard deviations for proximate  
 234 and ultimate analysis < 1.6 and 0.2, respectively).

| Sample        | VM   | FC   | ASH | C    | H   | N   | O*   | H/C  | O/C  |
|---------------|------|------|-----|------|-----|-----|------|------|------|
| <b>BW_raw</b> | 87.8 | 12.2 | 0.3 | 50.4 | 5.4 | 0.1 | 43.8 | 1.28 | 0.65 |
| <b>BW_160</b> | 88.5 | 11.5 | 0.2 | 50.8 | 5.5 | 0.0 | 43.5 | 1.29 | 0.64 |
| <b>BW_180</b> | 86.5 | 13.5 | 0.2 | 51.9 | 5.5 | 0.1 | 42.3 | 1.25 | 0.61 |
| <b>BW_200</b> | 82.2 | 17.8 | 0.2 | 54.7 | 5.4 | 0.1 | 39.6 | 1.17 | 0.54 |
| <b>BW_220</b> | 77.4 | 22.6 | 0.2 | 57.7 | 5.2 | 0.1 | 37.0 | 1.08 | 0.48 |
| <b>BW_230</b> | 72.0 | 28.0 | 0.2 | 60.8 | 5.1 | 0.1 | 33.8 | 0.99 | 0.42 |
| <b>BW_240</b> | 62.4 | 37.6 | 0.2 | 66.5 | 4.8 | 0.2 | 28.4 | 0.86 | 0.32 |
| <b>BW_260</b> | 54.4 | 45.6 | 0.1 | 71.7 | 4.6 | 0.2 | 23.4 | 0.77 | 0.24 |
| <b>BW_280</b> | 50.8 | 49.2 | 0.1 | 73.2 | 4.6 | 0.2 | 21.8 | 0.75 | 0.22 |

235 \*Calculated by difference O =100-Ash-C-H-N

236 The initial decrease of HTC solid mass yield, faster than that observed for pure  
 237 cellulose, can be ascribed to the decomposition of extractives and hemicellulose  
 238 occurring at low HTC temperature (160-200 °C), figure 1.



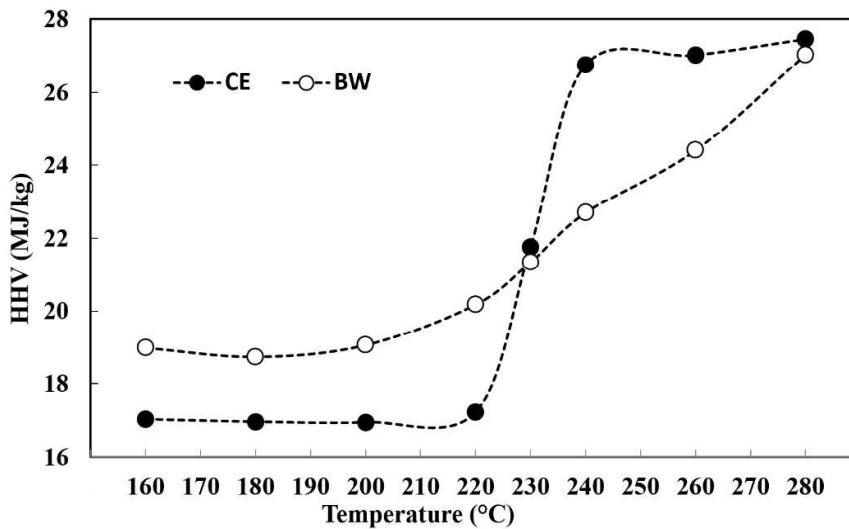
239

240 Fig. 1 Cellulose and birchwood HTC mass yields vs. HTC temperature (SY= solid  
 241 yield, LY= liquid yield, GY= gas yield)

242 Figure 1 shows the comparison of mass yields changes of pure cellulose and birchwood  
 243 hydrochars with HTC temperature. Notably, while cellulose solid yield remains  
 244 approximately constant up to HTC temperature of 220 °C, birchwood hydrochar yield  
 245 starts decreasing already at 160 °C, due to hydrolysis and/or removal of extractives first  
 246 and hemicellulose degradation starting at about 180 °C [37]. The decrease of birchwood  
 247 solid yield results in a liquid yield increase, while gas yields are negligible up to 230 °C  
 248 and increase slowly up to approximately 9 wt% at 280 °C for both feedstock. After a  
 249 sharp drop at 230 °C, cellulose hydrochar yields remained constant at approximately 50  
 250 wt% between 240 and 280 °C, thereby demonstrating that no further significant  
 251 degradation occurred at temperatures equal to or higher than 240 °C. Conversely,  
 252 birchwood hydrochar samples showed mass yield progressively decreasing down to 52  
 253 wt% at a HTC temperature of 280 °C. Mass yields results for both samples are  
 254 consistent with the production and/or concentration of thermal resistant aromatic  
 255 material under HTC condition [36,37]. Recent works demonstrated that HTC of residual

256 biomass occurring at higher temperatures can also improve the rate of back  
257 polymerization of organics from the liquid to the solid phase, thus resulting in a  
258 decrease of liquid yield and the production of secondary char [45,46].

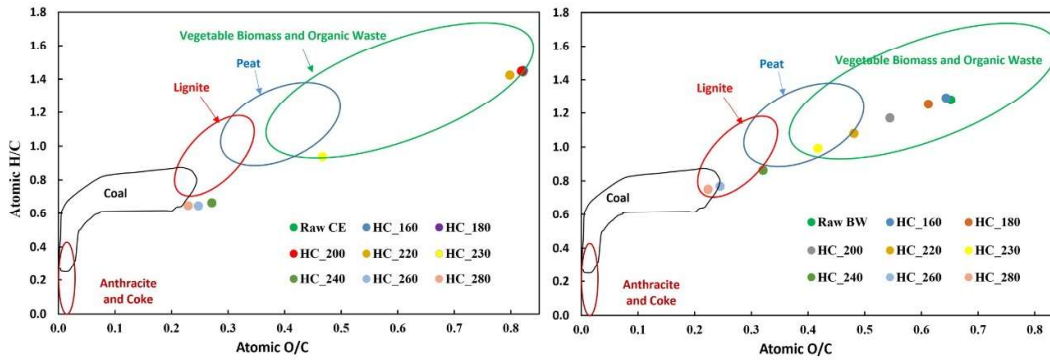
259 Figure 2 shows trend in HHV changes vs. HTC temperature for cellulose when  
260 compared to birchwood. cellulose hydrochar showed a 56% increase in HHV between  
261 220 and 240 °C, while in the same range of HTC temperature HHV of the  
262 corresponding birchwood hydrochars showed an increase of only 14%.



263

264 Fig. 2 Cellulose (CE) and birchwood (BW) HHV changes with HTC temperature

265 Van Krevelen plots for cellulose and birchwood samples reported in figure 3, show the  
266 sharp differences in carbonization degree of cellulose hydrochars between the samples  
267 produced at temperatures lower than 230 °C and those resulting from HTC performed  
268 above 230 °C. Conversely, birchwood hydrochar samples show a progressive  
269 carbonization with HTC temperature.



270  
271 Fig. 3 Van Krevelen plots of Cellulose (left) and Birchwood (right) samples

272 *3.2 Acid hydrolysis*

273 Acid hydrolysis results, reported in table 5, show that the glucose fraction in pure  
 274 cellulose sample, initially equal to approximately 99 wt%, decreased to 40 wt% at 230  
 275 °C and almost disappeared at 240 °C confirming the complete rupture of beta-glycosidic  
 276 linkage of cellulose. The sharp decrease of the glucose fraction in cellulose hydrochar  
 277 samples was accompanied by the increase of a lignin-like fraction with the production  
 278 of more resistant aromatic structured compounds. Acid hydrolysis of birchwood show  
 279 no evidence of such a sharp drop at 230 °C; conversely, the glucose fraction decreased  
 280 more gradually with increasing HTC temperature confirming what recently reported in  
 281 literature [44]. For example, at 240 °C cellulose and birchwood hydrochars showed a  
 282 glucose fraction of 0.8 and 20.3 wt%, respectively, which suggests a sort of protective  
 283 effect of the lignin content in the birchwood towards the cellulose content of the same  
 284 feedstock .

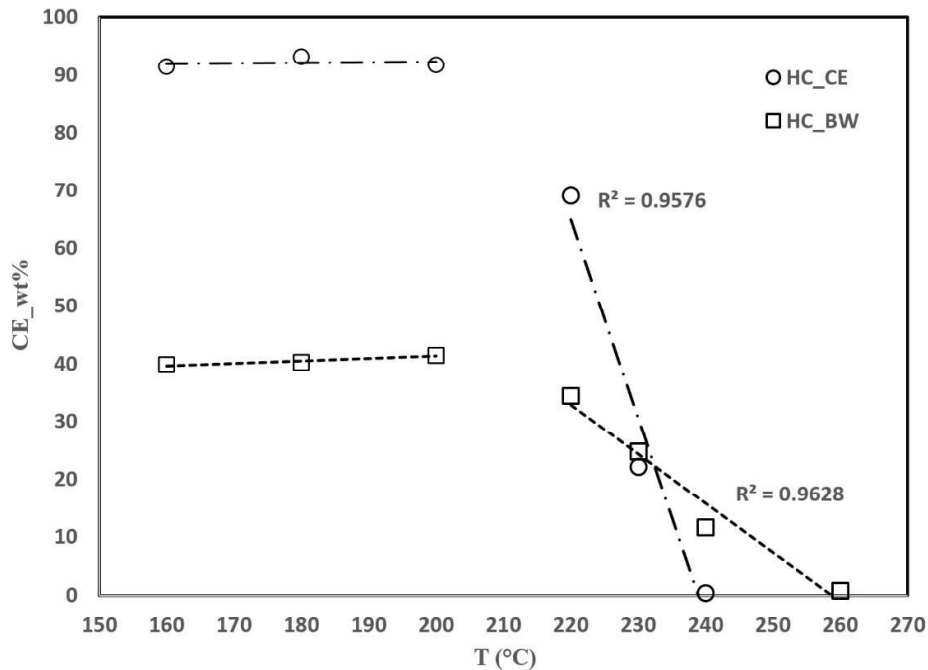
285 Table 5 – Acid hydrolysis analysis of raw cellulose and birchwood and corresponding  
 286 hydrochars. All values in wt% d.b. Standard deviation of the data < 2.2.

| Sample | Extractives | Lignin | Galactose | Glucose | Xylose | Mannose |
|--------|-------------|--------|-----------|---------|--------|---------|
| CE_raw | 0.0         | 0.0    | b.d.l.    | 98.9    | 3.0    | 2.6     |
| CE_160 | 0.0         | 0.0    | b.d.l.    | 93.1    | 2.8    | 2.5     |

|               |      |      |        |      |      |      |
|---------------|------|------|--------|------|------|------|
| <b>CE_180</b> | 0.0  | 0.0  | b.d.l. | 95.7 | 2.7  | 2.5  |
| <b>CE_200</b> | 0.0  | 0.0  | b.d.l. | 96.2 | 1.8  | 1.9  |
| <b>CE_220</b> | 3.0  | 9.0  | b.d.l. | 84.0 | 0.5  | 0.9  |
| <b>CE_230</b> | 24.0 | 40.0 | b.d.l. | 40.2 | 0.1  | 0.2  |
| <b>CE_240</b> | 21.0 | 74.0 | b.d.l. | 0.8  | n.a. | n.a. |
| <b>CE_260</b> | 21.0 | 79.0 | b.d.l. | 0.1  | n.a. | n.a. |
| <b>CE_280</b> | 16.0 | 79.0 | b.d.l. | 0.2  | n.a. | n.a. |
| <b>BW_raw</b> | 2.0  | 28.8 | 1.0    | 39.4 | 21.9 | 1.6  |
| <b>BW_160</b> | 4.0  | 20.5 | 0.6    | 43.4 | 18.3 | 1.5  |
| <b>BW_180</b> | 16.0 | 17.0 | 0.3    | 47.3 | 7.5  | 1.0  |
| <b>BW_200</b> | 26.0 | 15.0 | 0.1    | 58.7 | 2.0  | 0.3  |
| <b>BW_220</b> | 26.0 | 24.0 | b.d.l. | 52.4 | 0.2  | 0.1  |
| <b>BW_230</b> | 28.0 | 31.0 | b.d.l. | 39.9 | 0.1  | 0.1  |
| <b>BW_240</b> | 28.0 | 49.0 | b.d.l. | 20.3 | 0.1  | 0.0  |
| <b>BW_260</b> | 34.0 | 60.0 | b.d.l. | 1.6  | 0.1  | n.a. |
| <b>BW_280</b> | 34.0 | 61.0 | b.d.l. | 0.1  | 0.0  | n.a. |

287 b.d.l.: below detection limits

288 Figure 4 shows the change of cellulose percentage (normalized by the corresponding  
289 hydrochar mass yield) vs. HTC temperature. Cellulose weight percentage in pure  
290 cellulose and birchwood hydrochars shows a bi-modal trend depending on the HTC  
291 reaction temperature. On the one hand, between 160 and 200 °C, cellulose in both series  
292 of hydrochars is approximately constant; on the other hand, between 220 and 260 °C, it  
293 decreased quickly, thereby showing linear trends. Notably, cellulose rate of  
294 decomposition in birchwood was more than four times slower than pure cellulose  
295 hydrochars. The lower rate of cellulose decomposition in birchwood could be associated  
296 to the presence of interwoven lignin acting as a protecting shield in the lignocellulosic  
297 matrix.



298

299 Fig. 4 Normalized cellulose weight component changes with HTC temperature in pure  
 300 cellulose and birchwood hydrochars

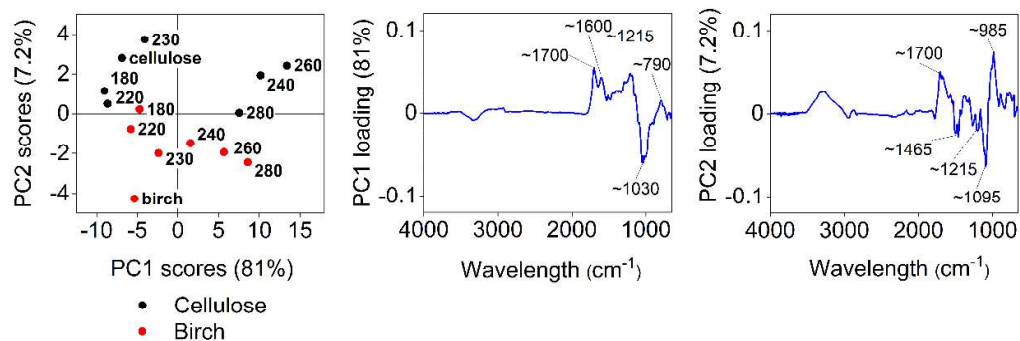
### 301 3.3 FTIR analysis

302 FTIR analysis has been widely used to investigate chemical and structural changes  
 303 occurring in biomass during HTC [36,47]. The PCA results based on the first two PCs  
 304 are illustrated in figure 5. In general, positive sample scores in PCA correspond to  
 305 increased absorbance on positive loadings. The same also applies with negative scores  
 306 and negative loadings. As illustrated in figure 5, the first PC explained 81% of the  
 307 variation in the absorbance spectra and mainly separated the samples based on HTC  
 308 temperature. Higher HTC temperatures hence lead to a decreased adsorbance at around  
 309  $3300\text{ cm}^{-1}$  attributable to dehydration reactions. Increased absorbance at around 2950-  
 310  $3100$  and  $790\text{ cm}^{-1}$  can be assigned to sp C-H stretching and bending out of plane modes  
 311 respectively, testifying a progressive aromatization of hydrochars with HTC  
 312 temperature. Moreover, the sharp increase of absorbance especially at approximately



313 1700 and 1215  $\text{cm}^{-1}$  can be attributed to the free C=O stretching and bending modes due  
314 to cellulose beta-glucosidic linkage breaking [48]. The decrease in absorbance observed  
315 at around 1030  $\text{cm}^{-1}$  provides evidence that pyranose structure of glucose is lost with  
316 increasing HTC temperature [49].

317 The second PC, which explained approximately 7% of the variation in the spectra,  
318 provided mainly a separation between the cellulose and birch wood samples. The  
319 cellulose samples showed consistently higher IR absorbance at 1700 and 985  $\text{cm}^{-1}$ ,  
320 probably due to higher concentration of free carbonyl (C=O) and aliphatic (C-O-C  
321 ether and alcohol C-O) respectively, due to pyranose structure breaking in cellulose  
322 hydrochar samples.

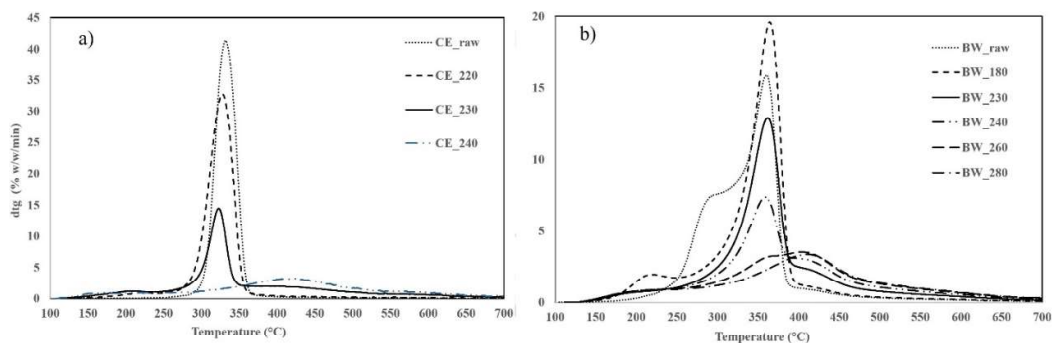


323

324 Fig. 5 Sample scores based on the first two PCs and the respective first (middle) and  
325 second (right) PC loadings.

### 326 3.4 Derivative Thermogravimetric (DTG) analysis and cellulose reactivity

327 Figure 6a and b show DTG curves of cellulose and birchwood, respectively.



328

329 Fig. 6a, b DTG curves of: raw cellulose and selected cellulose hydrochars (a); raw  
 330 birchwood and selected birchwood hydrochars (b).

331 Graphs show a sharp decrease in reactivity of cellulose hydrochars above 230 °C and of  
 332 birchwood hydrochars above 240 °C. It may be noted that all birchwood samples are  
 333 significantly less reactive than cellulose samples as it is expected owing to the content  
 334 of lignin in birchwood (cellulose samples obviously do not contain any lignin). In  
 335 addition, reactivity decreases more sharply and faster with HTC temperature in cellulose  
 336 hydrochars when compared to birchwood hydrochar. However, the 240 °C birchwood  
 337 hydrochar sample shows some residual marginally higher reactivity than the pure  
 338 cellulose hydrochar. Again, probably this is related to the presence of lignin in  
 339 birchwood and the higher ‘resilience’ of the compounds formed during carbonisation.

#### 340 4. Conclusions

341 This study sheds light on the role of lignin in cellulose decomposition in lignocellulosic  
 342 biomass. Cellulose component present in naturally occurring biomass matrix is less  
 343 reactive than free cellulose due to a ‘protecting shield’ offered by lignin. Mass yields,  
 344 energy properties, FTIR, DTG and acid hydrolysis analysis demonstrate that pure  
 345 cellulose is barely affected at a HTC temperature lower than 230 °C (residence time: 0.5  
 346 h) but severely degrades at a HTC temperature equal to or higher than 230 °C.

347 Conversely, natural occurring cellulose in birchwood degrades progressively at  
348 increasing HTC temperature, decomposing completely at 280 °C.

349 **Declaration of competing interest**

350 The authors declared that they have no conflicts of interest to this work. We declare that  
351 we do not have any commercial or associative interest that represents a conflict of  
352 interest in connection with the work submitted.

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357

358 **References**

- 359 [1] European Commission, A policy framework for climate and energy in the period,  
360 Communication From the Commission To the European Parliament, the Council,  
361 the European Economic and Social Committee and the Committee of the  
362 Regions. (2014) 18. [http://eur-lex.europa.eu/legal-](http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52014DC0015&from=EN)  
363 [content/EN/TXT/PDF/?uri=CELEX:52014DC0015&from=EN](http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52014DC0015&from=EN).
- 364 [2] C. Gopu, L. Gao, M. Volpe, L. Fiori, J.L. Goldfarb, Valorizing municipal solid  
365 waste: Waste to energy and activated carbons for water treatment via pyrolysis,  
366 Journal of Analytical and Applied Pyrolysis. 133 (2018) 48–58.  
367 <https://doi.org/10.1016/j.jaap.2018.05.002>.
- 368 [3] N.S. Bentsen, C. Felby, Biomass for energy in the European Union - A review of  
369 bioenergy resource assessments, Biotechnology for Biofuels. 5 (2012) 1–10.  
370 <https://doi.org/10.1186/1754-6834-5-25>.
- 371 [4] C. Mao, Y. Feng, X. Wang, G. Ren, Review on research achievements of biogas  
372 from anaerobic digestion, Renewable and Sustainable Energy Reviews. 45 (2015)  
373 540–555. <https://doi.org/10.1016/j.rser.2015.02.032>.
- 374 [5] L. Yang, F. Xu, X. Ge, Y. Li, Challenges and strategies for solid-state anaerobic  
375 digestion of lignocellulosic biomass, Renewable and Sustainable Energy  
376 Reviews. 44 (2015) 824–834. <https://doi.org/10.1016/j.rser.2015.01.002>.
- 377 [6] M.J. Prins, K.J. Ptasiński, F.J.J.G. Janssen, Torrefaction of wood Part 2. Analysis  
378 of products, Journal of Analytical and Applied Pyrolysis. 77 (2006) 35–40.  
379 <https://doi.org/10.1016/j.jaap.2006.01.001>.
- 380 [7] R. Volpe, A. Messineo, M. Millan, M. Volpe, R. Kandiyoti, Assessment of olive  
381 wastes as energy source: pyrolysis, torrefaction and the key role of H loss in  
382 thermal breakdown, Energy. 82 (2015) 119–127.  
383 <https://doi.org/10.1016/j.energy.2015.01.011>.
- 384 [8] V. Benavente, A. Fullana, Torrefaction of olive mill waste, Biomass and  
385 Bioenergy. 73 (2015) 186–194. <https://doi.org/10.1016/j.biombioe.2014.12.020>.
- 386 [9] Z. Li, N. Li, W. Yi, P. Fu, Y. Li, X. Bai, Design and operation of a down-tube  
387 reactor demonstration plant for biomass fast pyrolysis, Fuel Processing  
388 Technology. 161 (2017) 182–192. <https://doi.org/10.1016/j.fuproc.2016.12.014>.
- 389 [10] D. Chiaramonti, M. Prussi, M. Buffi, A.M. Rizzo, L. Pari, Review and  
390 experimental study on pyrolysis and hydrothermal liquefaction of microalgae for  
391 biofuel production, Applied Energy. 185 (2017) 963–972.  
392 <https://doi.org/10.1016/j.apenergy.2015.12.001>.
- 393 [11] R. Volpe, J.M.B. Menendez, T.R. Reina, A. Messineo, M. Millan, Evolution of  
394 chars during slow pyrolysis of citrus waste, Fuel Processing Technology. 158  
395 (2017) 255–263. <https://doi.org/10.1016/j.fuproc.2017.01.015>.
- 396 [12] S. You, Y.S. Ok, S.S. Chen, D.C.W. Tsang, E.E. Kwon, J. Lee, C.H. Wang, A  
397 critical review on sustainable biochar system through gasification: Energy and  
398 environmental applications, Bioresource Technology. 246 (2017) 242–253.

- 399 <https://doi.org/10.1016/j.biortech.2017.06.177>.
- 400 [13] M.J. Prins, K.J. Ptasiński, F.J.J.G. Janssen, More efficient biomass gasification  
401 via torrefaction, *Energy*. 31 (2006) 3458–3470.  
402 <https://doi.org/10.1016/j.energy.2006.03.008>.
- 403 [14] A. Funke, F. Ziegler, Hydrothermal carbonization of biomass: A summary and  
404 discussion of chemical mechanisms for process engineering, *Biofuels Bioproduct  
405 & Biorefinery*. 4 (2010) 160–177. <https://doi.org/10.1002/bbb>.
- 406 [15] A. Kruse, A. Funke, M.-M. Titirici, Hydrothermal conversion of biomass to fuels  
407 and energetic materials, *Current Opinion in Chemical Biology*. 17 (2013) 515–  
408 521. <https://doi.org/10.1016/j.cbpa.2013.05.004>.
- 409 [16] S. Román, B. Ledesma, A. Álvarez, C. Coronella, S. V. Qaramaleki, Suitability  
410 of hydrothermal carbonization to convert water hyacinth to added-value products,  
411 *Renewable Energy*. 146 (2020) 1649–1658.  
412 <https://doi.org/10.1016/j.renene.2019.07.157>.
- 413 [17] M. Volpe, L. Fiori, R. Volpe, A. Messineo, Upgrading of Olive Tree Trimmings  
414 Residue as Biofuel by Hydrothermal Carbonization and Torrefaction: a  
415 Comparative Study, *Chemical Engineering Transaction*. 50 (2016) 13–18.  
416 <https://doi.org/10.3303/CET1650003>.
- 417 [18] Z. Liu, R. Balasubramanian, Upgrading of waste biomass by hydrothermal  
418 carbonization (HTC) and low temperature pyrolysis (LTP): A comparative  
419 evaluation, *Applied Energy*. 114 (2014) 857–864.  
420 <https://doi.org/10.1016/j.apenergy.2013.06.027>.
- 421 [19] M. Pecchi, F. Patuzzi, V. Benedetti, R. Di Maggio, M. Baratieri,  
422 Thermodynamics of hydrothermal carbonization: Assessment of the heat release  
423 profile and process enthalpy change, *Fuel Processing Technology*. 197 (2020)  
424 106206. <https://doi.org/10.1016/j.fuproc.2019.106206>.
- 425 [20] M.-M. Titirici, A. Thomas, M. Antonietti, Back in the black: hydrothermal  
426 carbonization of plant material as an efficient chemical process to treat the CO<sub>2</sub>  
427 problem?, *New Journal of Chemistry*. 31 (2007) 787–789.  
428 <https://doi.org/10.1039/b616045j>.
- 429 [21] A. Saha, P. Saha, M.T. Reza, Co-Hydrothermal Carbonization of coal-biomass  
430 blend: Influence of temperature on solid fuel properties, *Fuel Processing  
431 Technology*. 167 (2017) 711–720. <https://doi.org/10.1016/j.fuproc.2017.08.016>.
- 432 [22] N. Saha, A. Saha, M.T. Reza, Effect of hydrothermal carbonization temperature  
433 on pH, dissociation constants, and acidic functional groups on hydrochar from  
434 cellulose and wood, *Journal of Analytical and Applied Pyrolysis*. 137 (2019)  
435 138–145. <https://doi.org/10.1016/j.jaap.2018.11.018>.
- 436 [23] M. Volpe, D. Wüst, F. Merzari, M. Lucian, G. Andreottola, A. Kruse, L. Fiori,  
437 One stage olive mill waste streams valorisation via hydrothermal carbonisation,  
438 *Waste Management*. 80 (2018) 224–234.  
439 <https://doi.org/10.1016/j.wasman.2018.09.021>.

- 440 [24] M.T. Reza, M.H. Uddin, J.G. Lynam, S.K. Hoekman, C.J. Coronella,  
441 Hydrothermal carbonization of loblolly pine: reaction chemistry and water  
442 balance, *Biomass Conversion and Biorefinery*. 4 (2014) 311–321.  
443 <https://doi.org/10.1007/s13399-014-0115-9>.
- 444 [25] E. Sabio, A. Álvarez-Murillo, S. Román, B. Ledesma, Conversion of tomato-peel  
445 waste into solid fuel by hydrothermal carbonization: Influence of the processing  
446 variables, *Waste Management*. 47 (2016) 122–132.  
447 <https://doi.org/10.1016/j.wasman.2015.04.016>.
- 448 [26] E. Erdogan, B. Atila, J. Mumme, M.T. Reza, A. Toptas, M. Elibol, J. Yanik,  
449 Characterization of products from hydrothermal carbonization of orange pomace  
450 including anaerobic digestibility of process liquor, *Bioresource Technology*. 196  
451 (2015) 35–42. <https://doi.org/doi.org/10.1016/j.biortech.2015.06.115>.
- 452 [27] M.T. Reza, E. Rottler, L. Herklotz, B. Wirth, Hydrothermal carbonization (HTC)  
453 of wheat straw: Influence of feedwater pH prepared by acetic acid and potassium  
454 hydroxide, *Bioresource Technology*. 182 (2015) 336–344.  
455 <https://doi.org/10.1016/j.biortech.2015.02.024>.
- 456 [28] T. Wang, Y. Zhai, H. Li, Y. Zhu, S. Li, C. Peng, B. Wang, Z. Wang, Y. Xi, S.  
457 Wang, C. Li, Co-hydrothermal carbonization of food waste-woody biomass  
458 blend towards biofuel pellets production, *Bioresource Technology*. 267 (2018)  
459 371–377. <https://doi.org/10.1016/j.biortech.2018.07.059>.
- 460 [29] M. Lucian, M. Volpe, L. Gao, G. Piro, J.L. Goldfarb, L. Fiori, Impact of  
461 hydrothermal carbonization conditions on the formation of hydrochars and  
462 secondary chars from the organic fraction of municipal solid waste, *Fuel*. 233  
463 (2018) 257–268. <https://doi.org/doi.org/10.1016/j.fuel.2018.06.060>.
- 464 [30] M. Mäkelä, V. Benavente, A. Fullana, Hydrothermal carbonization of industrial  
465 mixed sludge from a pulp and paper mill, *Bioresource Technology*. 200 (2016)  
466 444–450. <https://doi.org/10.1016/j.biortech.2015.10.062>.
- 467 [31] H. Wikberg, T. Ohra-aho, M. Honkanen, H. Kanerva, A. Harlin, M. Vippola, C.  
468 Laine, Hydrothermal carbonization of pulp mill streams, *Bioresource*  
469 *Technology*. 212 (2016) 236–244. <https://doi.org/10.1016/j.biortech.2016.04.061>.
- 470 [32] Z. Al-Kaabi, R. Pradhan, N. Thevathasan, A. Gordon, Y.W. Chiang, A. Dutta,  
471 Bio-carbon production by oxidation and hydrothermal carbonization of paper  
472 recycling black liquor, *Journal of Cleaner Production*. 213 (2019) 332–341.  
473 <https://doi.org/10.1016/j.jclepro.2018.12.175>.
- 474 [33] Y. Zhai, X. Liu, Y. Zhu, C. Peng, T. Wang, L. Zhu, C. Li, G. Zeng,  
475 Hydrothermal carbonization of sewage sludge: The effect of feed-water pH on  
476 fate and risk of heavy metals in hydrochars, *Bioresource Technology*. 218 (2016)  
477 183–188. <https://doi.org/10.1016/j.biortech.2016.06.085>.
- 478 [34] F. Merzari, M. Langone, G. Andreottola, L. Fiori, Methane production from  
479 process water of sewage sludge hydrothermal carbonization. A review.  
480 Valorising sludge through hydrothermal carbonization, *Critical Reviews in*  
481 *Environmental Science and Technology*. 49 (2019) 947–988.

- 482 <https://doi.org/10.1080/10643389.2018.1561104>.
- 483 [35] M.E. Iñiguez, J.A. Conesa, A. Fullana, Hydrothermal carbonization (HTC) of  
484 marine plastic debris, *Fuel*. 257 (2019) 116033.  
485 <https://doi.org/10.1016/j.fuel.2019.116033>.
- 486 [36] X. Zhuang, H. Zhan, Y. Song, C. He, Y. Huang, X. Yin, Insights into the  
487 evolution of chemical structures in lignocellulose and non lignocellulose  
488 biowastes during hydrothermal carbonization (HTC), *Fuel*. 236 (2019) 960–974.  
489 <https://doi.org/10.1016/j.fuel.2018.09.019>.
- 490 [37] A.M. Borrero-López, E. Masson, A. Celzard, V. Fierro, Modelling the reactions  
491 of cellulose, hemicellulose and lignin submitted to hydrothermal treatment,  
492 *Industrial Crops & Products*. 124 (2018) 919–930.  
493 <https://doi.org/10.1016/j.indcrop.2018.08.045>.
- 494 [38] X. Lu, P.J. Pellechia, J.R.V. Flora, N.D. Berge, Influence of reaction time and  
495 temperature on product formation and characteristics associated with the  
496 hydrothermal carbonization of cellulose, *Bioresource Technology*. 138 (2013)  
497 180–190. <https://doi.org/10.1016/j.biortech.2013.03.163>.
- 498 [39] Y. Gao, X.H. Wang, H.P. Yang, H.P. Chen, Characterization of products from  
499 hydrothermal treatments of cellulose, *Energy*. 42 (2012) 457–465.  
500 <https://doi.org/10.1016/j.energy.2012.03.023>.
- 501 [40] A. George, T.J. Morgan, R. Kandiyoti, Pyrolytic Reactions of Lignin within  
502 Naturally Occurring Plant Matrices: Challenges in Biomass Pyrolysis Modeling  
503 Due to Synergistic Effects, *Energy & Fuels*. 28 (2014) 6918–6927.  
504 <https://doi.org/10.1021/ef501459c>.
- 505 [41] a. Sluiter, B. Hames, R. Ruiz, C. Scarlata, J. Sluiter, D. Templeton, D. Crocker,  
506 NREL/TP-510-42618 analytical procedure - Determination of structural  
507 carbohydrates and lignin in Biomass, Laboratory Analytical Procedure (LAP).  
508 (2012) 17. <https://doi.org/NREL/TP-510-42618>.
- 509 [42] P. Geladi, Chemometrics in spectroscopy. Part 1. Classical chemometrics,  
510 *Spectrochimica Acta - Part B Atomic Spectroscopy*. 58 (2003) 767–782.  
511 [https://doi.org/10.1016/S0584-8547\(03\)00037-5](https://doi.org/10.1016/S0584-8547(03)00037-5).
- 512 [43] R. Bro, A.K. Smilde, Principal component analysis, *Analytical Methods*. 6  
513 (2014) 2812–2831. <https://doi.org/10.1039/c3ay41907j>.
- 514 [44] M. Mäkelä, M. Volpe, R. Volpe, L. Fiori, O. Dahl, Spatially resolved spectral  
515 determination of polysaccharides in hydrothermally carbonized biomass, *Green  
516 Chemistry*. 20 (2018) 1114–1120. <https://doi.org/10.1039/C7GC03676K>.
- 517 [45] M. Lucian, M. Volpe, L. Fiori, Hydrothermal Carbonization Kinetics of  
518 Lignocellulosic Agro-Wastes: Experimental Data and Modeling, *Energies*. 516  
519 (2019). <https://doi.org/10.3390/en12030516>.
- 520 [46] M. Volpe, L. Fiori, From olive waste to solid biofuel through hydrothermal  
521 carbonisation: The role of temperature and solid load on secondary char  
522 formation and hydrochar energy properties, *Journal of Analytical and Applied*

- 523 Pyrolysis. 124 (2017) 63–72. <https://doi.org/doi.org/10.1016/j.jaap.2017.02.022>.
- 524 [47] M.T. Reza, W. Becker, K. Sachsenheimer, J. Mumme, Hydrothermal  
525 carbonization (HTC): Near infrared spectroscopy and partial least-squares  
526 regression for determination of selective components in HTC solid and liquid  
527 products derived from maize silage, *Bioresource Technology*. 161 (2014) 91–  
528 101. <https://doi.org/doi.org/10.1016/j.biortech.2014.03.008>.
- 529 [48] W. Peng, L. Wang, M. Ohkoshi, M. Zhang, Separation of hemicelluloses from  
530 Eucalyptus species: Investigating the residue after alkaline treatment, *Cellulose  
531 Chemistry and Technology*. 49 (2015) 757–764.
- 532 [49] A. Álvarez-Murillo, E. Sabio, B. Ledesma, S. Rom, Generation of biofuel from  
533 hydrothermal carbonization of cellulose. *Kinetics modelling, Energy*. 94 (2016)  
534 600–608. <https://doi.org/10.1016/j.energy.2015.11.024>.
- 535



## Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

## Supplementary Materials

### Reactivity of cellulose during hydrothermal carbonization of lignocellulosic biomass

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Fig. S1a, b report FTIR spectra of Cellulose (a) and Birchwood (b) of raw and hydrochars samples at different HTC temperatures.

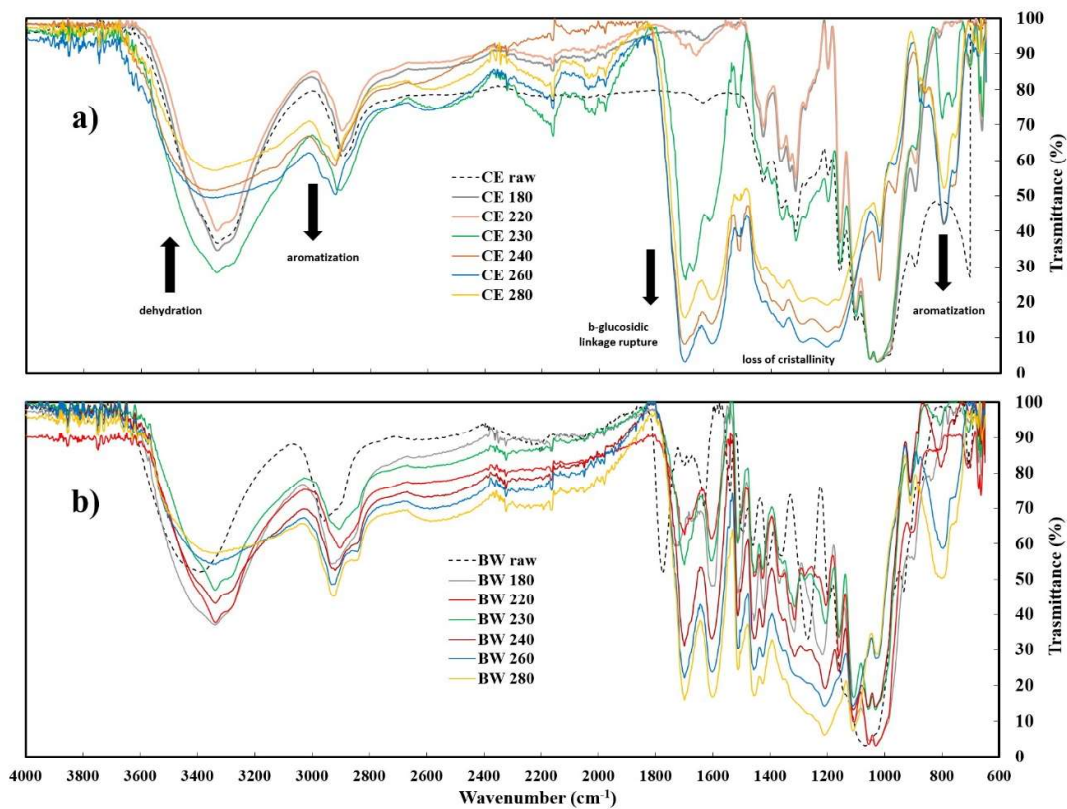


Fig. S1a, b FTIR spectra of Cellulose (a) and Birchwood (b) raw and corresponding hydrochars

Figure S2 reports the normalized FTIR spectra transmittance and their conversion into absorbance spectra for Principal Component Analysis, (PCA) of Cellulose and Birchwood raw and hydrochar samples at different HTC temperature.

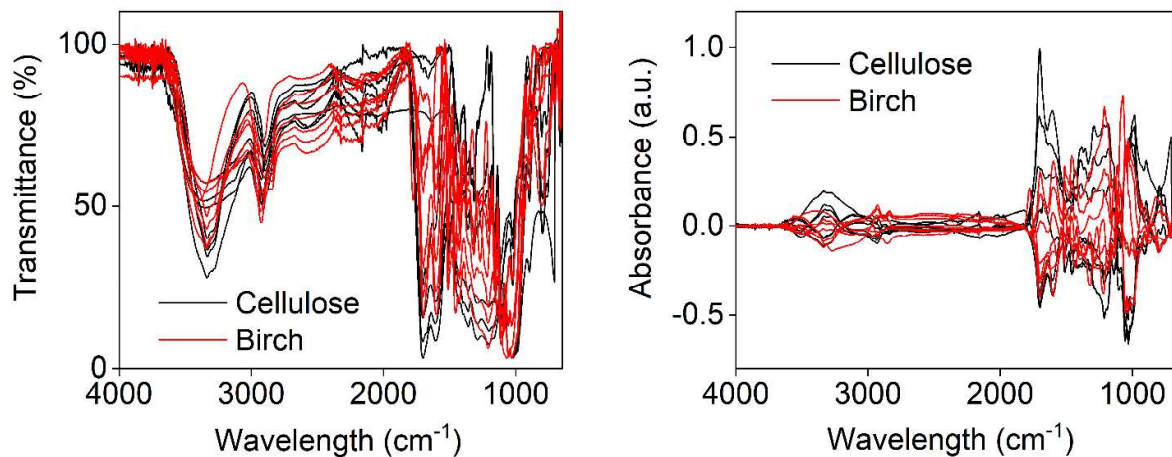


Fig. S2: The FTIR spectra after baseline correction and normalization in transmittance (%), (left) and the preprocessed spectra in absorbance units before PCA (right).