1	Recovering ammonium by treated and untreated zeolitic mixtures: a comprehensive
2	experimental and modelling study
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# 13 Highlights:

14	•	Two zeolitic mixtures with different mineralogy were used for $NH_4^+$ recovery
15	•	Zeolitic mixtures showed good performance for $NH_4^+$ removal from solution
16	•	The mineralogical composition of zeolitic mixtures affected $NH_4^+$ adsorption
17	•	The bimodal-pseudo-first order model well described NH4 <sup>+</sup> adsorption kinetics
18	•	The Langmuir-Sips model provides the best fit for the $NH_4^+$ adsorption isotherm
19		

## 20 ABSTRACT

The recovery of ammonium  $(NH_4^+)$  from aqueous solutions by zeolite is attractive. In this study, the physical-chemistry of  $NH_4^+$  adsorption process from aqueous solution by two zeolitic mixtures,

23 either treated or not treated with NaCl, was assessed.

Results suggested that the zeolitic mixture richer in mordenite and with high specific surface area 24 adsorbed more NH<sub>4</sub><sup>+</sup> than the one richer in clinoptilolite and heulandite showing a lower specific 25 26 surface area. NaCl treatment increased the amount of NH<sub>4</sub><sup>+</sup> adsorbed by the zeolitic mixtures. The higher amount of NH<sub>4</sub><sup>+</sup> adsorbed by the zeolitic mixtures treated with NaCl was explained by the 27 low/high density water model accounting for cation exchange among the two kosmotropic systems: 28 29 Na-enriched zeolitic mixtures and NH<sub>4</sub><sup>+</sup>-enriched aqueous solution. The adsorption kinetics were best approximated by the bimodal pseudo-first-order model. The two sorption kinetic constants, k1 30 and  $k_2$  were related to the adsorption (mediated by  $k_1$ ) and the ion exchange (mediated by  $k_2$ ) 31 32 processes. The fitting of NH<sub>4</sub><sup>+</sup> data to Langmuir-Sips model suggested that the NaCl treatment increased the number of active sites only of the zeolitic mixture with the large amount of mordenite. 33 Thus, it is conceivable that modulation of NaCl treatment of zeolitic mixtures can be applied to 34 obtain new materials for water remediation from NH<sub>4</sub><sup>+</sup> contamination. 35

36

## 37 Keywords:

38 Ammonium removal; Zeolites mineralogy; NaCl treatment; Kinetics models; Isotherm models.

39

#### 40 1. INTRODUCTION

41 Porous materials with an appreciable absorptive capacity are considered an attractive solution for recovering nutrients from treated and not treated wastewaters [1]. Natural zeolites for 42 environmental applications are gaining renewed interest owing to their abundance, ease of 43 extraction, and unique properties, such as cation exchange (CEC) and adsorption (AC) capacities 44 [1–3]. Having high affinity for the adsorption of cations, zeolite is commonly used for the removal 45 46 of NH<sub>4</sub><sup>+</sup> from aqueous solutions such as treated and untreated wastewaters [1]. Then, in view of a circular economy approach, N enriched zeolite can be reused as slow-release fertilizers in 47 agriculture [4–6]. 48

The amount of  $NH_4^+$  adsorbed by zeolite depends, among the others, on mineralogical structure, chemical composition, and chemical treatment [7]. According to the process of formation, it is possible to distinguish among more than 70 types of natural zeolites. The most common zeolite forms are clinoptilolite and mordenite with a  $NH_4^+$  adsorption capacity ranging between 41 and 72 mg g<sup>-1</sup> [1].

By studying 8 different zeolites, Langwaldt et al. [2] revealed that the adsorption capacity depends 54 mainly on the mineralogical composition of zeolites. They found that zeolites containing the largest 55 amount of chabazite (82%) adsorbed 48 mg  $NH_4^+$  g<sup>-1</sup>, against zeolites with 93% of clinoptilolite that 56 adsorbed up to 25 mg NH<sub>4</sub><sup>+</sup> g<sup>-1</sup>. However, such a large variability in NH<sub>4</sub><sup>+</sup>adsorbtion has been also 57 found among zeolites with the same mineralogy. Wang and Peng [8] reported that the amount of 58 NH<sub>4</sub><sup>+</sup> adsorbed by clinoptilolite from different origins may range from 3 to 23 mg g<sup>-1</sup>. The latter 59 variability depends on a variety of precursor materials, including volcanic and impact glass, 60 aluminosilicate gels and aluminosilicate minerals, such as other zeolites, smectite, kaolinite, 61 62 feldspars and feldspatoids. Indeed, such factors may affect surface area and total pore volume as well as the Si/Al ratio. The latter properties are directly correlated with the negative charge of 63 zeolite. In fact, the lower the Si/Al ratio, the larger the amount of the negative charge is [9]. 64

Also, the treatment to which zeolites undergo prior to any application may affect adsorption capacity. Thushari et al. [10] found that, following the treatment with NaCl, NH<sub>4</sub><sup>+</sup> adsorption by zeolite increased by 25% compared to the untreated one. Recently, also Muscarella et al. [11] revealed that 1 M NaOH treatment of zeolite resulted in a NH<sub>4</sub><sup>+</sup> adsorbing capacity of 18.3 mg g<sup>-1</sup>, whereas an acid treatment with 0.1 M HCl produced zeolite with NH<sub>4</sub><sup>+</sup> adsorption capacity of 22.5 mg g<sup>-1</sup>.

In addition, the interactions between the extra framework cations and the framework of zeolite have
been studied by using atomistic simulation techniques [12] and no conceptual model of the cation
exchange between aqueous solution and zeolite have been proposed.

The high variability in adsorbing NH<sub>4</sub><sup>+</sup> showed by zeolite with different mineralogy as related to acid, alkaline or salt treatment stresses the need for further investigation. Moreover, at our best knowledge, no mechanism has been proposed up to now about the exchange between NH<sub>4</sub><sup>+</sup> in aqueous solution and cations adsorbed onto zeolite either treated or not with NaCl. Thus, the aim of the present study is to provide more insights about the ability of zeolite with different mineralogy, either treated or not with NaCl, in exchanging and adsorbing NH<sub>4</sub><sup>+</sup> from aqueous solution. Furthermore, a new conceptual model about the exchange and adsorption processes between NH<sub>4</sub><sup>+</sup>

81 in solution and cations adsorbed onto zeolite is proposed.

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## 83 2. MATERIAL AND METHODS

## 84 2.1 Tested zeolite

Two mixtures of natural zeolites (Ø 0.5-1 mm) were used in this study. They are indicated as ZNS and ZNC. Before their chemical and mineralogical characterization, and use, the two zeolitic mixtures were washed trice with distilled water to remove particulate impurities on the surfaces and dried at 105°C for 2 hours [11]. ZSS and ZSC were, respectively, two mixtures obtained by the previous ones after treatment with 1M NaCl for 24 h with a zeolite-to-solution ratio of 1:50 (w/v). After NaCl treatment, and prior to all the experiments, ZSS and ZSC have been washed with

distilled water in order to remove the excess of sodium. The specific surface area of the two 91 92 untreated zeolitic mixtures was evaluated by the nitrogen gas adsorption method, using an automated equipment (Nova touch LX1, Quantachrome Instrument, FL, USA), employing 93 multipoint BET isotherm adsorption data fitting. 94 Morphology and elemental composition of treated and untreated zeolite mixtures were examined by 95 scanning electron microscope (SEM) coupled with energy-dispersive X-ray (EDX) analyser 96 97 (Phenom Pharos and Phenom XL, Phenom-World Thermo Scientific, Waltham, Massachusetts, USA). Powder X-ray diffraction (XRD) data for structure and phase identification of zeolite 98 mixtures were collected by X'PERT PRO, X-ray Diffractometer (Pan Analytical, Malvern, UK). 99 100 The main functional groups of zeolite mixtures were assessed by ATR-FTIR Spectrometry (Spectrum Two, PerkinElmer). Spectra were collected in the wavenumber range 4000-400 cm<sup>-1</sup> on 101 samples dried for 2 h at 105 °C and finely ground. The spectra have been elaborate by using Origin 102 103 (Version 7.5) software program.

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# 105 **2.2** Determination of the point of zero charge

The point of zero charge (pH<sub>pzc</sub>) of the zeolite mixtures was evaluated using the pH drift method 106 according to Nasiruddin et al. [13] as described in Vaičiukynienė et al. [14]. Sodium chloride 107 108 (0.01M) was used as a background electrolyte. Eight solutions with pH values in a range from 2 to 9 were prepared by adjusting the pH adding small amounts of 0.5M HCl or 0.5M NaOH solutions. 109 Then, 1.0 g of zeolitic mixtures were soaked with 40 mL of each solution and left to settle down for 110 111 24 h at room temperature. The final pH of each solution was measured. The  $pH_{pzc}$  of zeolitic mixtures was evaluated; if the initial pH of the solution was equal to the final pH of the solution, 112 then that was considered the  $pH_{pzc}$ , and the charge on the surface was zero [14]. 113

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### 115 2.3 Theory: kinetic and isotherm models

116 The  $NH_4^+$  sorption efficiency of zeolitic mixtures can be evaluated by studying both adsorption 117 kinetics and isotherms. When a kinetic curve is obtained, different models can be applied for the 118 fitting. In particular, the monomodal pseudo-first order model given in eq. (1) is based on the 119 assumption that the adsorption rate depends only on the amount of  $NH_4^+$  placed in contact with the 120 solid porous surface:

121 
$$q_t = q_e (1 - e^{-k_1 t})$$
 (1)

Here,  $q_t$  is the amount of NH<sub>4</sub><sup>+</sup> adsorbed on the surface of the porous system at time *t*,  $q_e$  is the equilibrium adsorption capacity (in mg g<sup>-1</sup>) and  $k_1$  is the pseudo-first order monomodal constant. The latter is directly related to the adsorption rate of the NH<sub>4</sub><sup>+</sup> on the solid surface. The larger is the  $k_1$  value, the faster is the adsorption rate.

Otherwise, the bimodal pseudo-first order model is based on the assumption that there is polymodality in the NH<sub>4</sub><sup>+</sup> adsorption, where two pseudo-first order monomodal adsorption patterns exist simultaneously, thereby resulting from different adsorption mechanisms:

129 
$$q_t = q_{e1}(1 - e^{-k_1 t}) + q_{e2}(1 - e^{-k_2 t})$$
 (2)

Here,  $q_t$  is the amount of NH<sub>4</sub><sup>+</sup> adsorbed on the surface of the porous system at time *t*,  $q_{e1}$  and  $q_{e2}$ are the equilibrium adsorption capacities (in mg g<sup>-1</sup>) at t<sub>1</sub> and t<sub>2</sub>, respectively, while  $k_1$  and  $k_2$  are the respective pseudo-first order monomodal kinetic constants. The larger the kinetic constant values, the faster are the adsorption rates mediated by k<sub>1</sub> and k<sub>2</sub>.

134 The adsorption isotherms allow to explain solid surface properties such as affinity for the adsorbent 135 and relationship between the amounts of ions exchanged by the porous system and ion

- 136 concentration at the equilibrium in solution [15]. Moreover, the adsorption isotherms can highlight
- 137 substantial differences in solid characteristics such as purity, mineral content, and chemical

138 composition.

Freundlich and Langmuir isotherm models are usually applied to evaluate NH4<sup>+</sup> removal from
wastewaters by using zeolites [9,16], although some criticisms have been found in the literature
[17].

142 Freundlich equation is in the form:

143 
$$q_e = K_F C^{1/n}$$
 (3)

where  $q_e$  is the amount of adsorbate per unit mass of adsorbent, *C* is NH<sub>4</sub><sup>+</sup> concentration at the equilibrium in solution,  $K_F$  is named Freundlich equilibrium constant or distribution factor, and 1/nis a correction factor related to the number (n) of adsorbing sites on the surface of the porous material [18].

- 148 According to Moshoeshoe et al. [16], eq. (3) fits experimental data with  $R^2$  values usually > 0.98.
- However, Canellas [9] revealed that the thermodynamics of  $NH_4^+$  removal from an aqueous

solution using natural zeolites of various origins is better described by the Langmuir isotherm givenin eq. (4).

$$152 \qquad q_e \propto \frac{K_L C}{K_L C + 1} \tag{4}$$

Here,  $q_e$  and *C* have the same meaning as in eq. (3), while  $K_L$  is the Langmuir equilibrium constant. To account for the non-ideal conditions, eq. (4) can be transformed as in eq. (5):

$$155 \quad q_e \propto \frac{K_L c^n}{K_L c^n + 1} \tag{5}$$

where n is an empirical parameter related to the effects of the adsorbate concentration [19].

157 Eq. (3) assumes a heterogeneous surface with a non-uniform distribution of heat of adsorption over

the surface and binding sites are not equivalent and/or independent [20]. Eq. (4) is valid for

- 159 monolayer adsorptions on a surface with a finite number of identical sites. However, equation (5) is
- a better model, it can be written as in equation (6) (also referred to as Langmuir-SIPS), where three
- adjustable parameters are accounted for [19,21,22].
- 162 They are  $q_{max}$ , i.e., the maximum amount of adsorbate that is a function of the number of active
- sites on the zeolite surface,  $K_L$ , that is the Langmuir's constant referring to the equilibrium

164  $Z + NH_4^+ \rightleftharpoons ZNH_4$ 

with Z = zeolite. In eq. (6) the meaning of the parameter n is the same as reported for eq. (5).

$$166 \qquad q_e = \frac{q_{max}K_L C^n}{K_L C^n + 1} \tag{6}$$

167 The advantage of eq. (6) lays in the fact that it contains both Freundlich and Langmuir isotherms. In 168 fact, when n = 1 it becomes the Langmuir isotherm reported in eq. (4), while as  $K_L$  approaches to 169 0, eq. (6) becomes the Freundlich eq. (3).

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## 171 **2.4** NH<sub>4</sub><sup>+</sup> adsorption by zeolitic mixtures

To evaluate NH<sub>4</sub><sup>+</sup> adsorption ability, 1 g of each sample (ZNC, ZSC, ZNS, ZSS) was shaken with 172 100 mL of 20 g  $NH_4^+ L^{-1}$  solution on an orbital shaker for 24 h at 80 rpm at 25°C. The 173 concentration of 20 g NH<sub>4</sub><sup>+</sup> L<sup>-1</sup> was used on the basis of the isotherm study that showed NH<sub>4</sub><sup>+</sup> 174 175 saturation of the samples after 24 h at this concentration. After 24 h, samples were washed three times with 200 mL of distilled water to remove the excess of NH<sub>4</sub><sup>+</sup> and dried in an oven for 2 h at 176 105 °C. NH4<sup>+</sup> adsorbed by zeolitic mixtures was determined by Kjeldahl distillation with 30 mL of 177 33% (w/v) NaOH solution for six minutes [11]. Amounts of  $NH_4^+$  adsorbed by zeolitic mixtures 178 179 were analysed by two-way ANOVA (mineralogy and NaCl treatment as factors) followed by Tukey test to assess significant differences at P < 0.05 among zeolitic mixtures. The experiment was 180 carried out in triplicate. 181

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# 183 **2.5** Adsorption kinetics

Adsorption kinetics, during 48 h, were assessed by contacting 1 g of each sample (ZNC, ZSC, ZNS,
ZSS) with 100 mL of a 50 mg NH<sub>4</sub><sup>+</sup> L<sup>-1</sup> solution on a horizontal shaker at 80 rpm at 25°C. NH<sub>4</sub><sup>+</sup>
concentration in solution was determined after 15, 30, 45 minutes, and 1, 2, 4, 8, 24, 48 hours by
Berthelot colorimetric method [23]. The NH<sub>4</sub><sup>+</sup> kinetics adsorption data were mathematically

analysed according to eqs. 1 and 2, by using Origin (Version 7.5) software program. Theexperiment was carried out in triplicate.

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## 191 **2.6 Equilibrium studies**

- 192 Data for  $NH_4^+$  adsorption isotherms were obtained by shaking 2 g of each sample (ZNC, ZSC, ZNS,
- 193 ZSS) with 200 mL of NH<sub>4</sub>Cl solution at different concentration (5, 10, 25, 50, 100, 250, 500, 1000,

194 5000, 10000, 20000 mg  $NH_4^+L^{-1}$ ) on a horizontal shaker at 80 rpm, 25°C for 24 h.

- 195 The NH<sub>4</sub><sup>+</sup> adsorbed on zeolitic mixtures, determined by Berthelot colorimetric method, was fitted to
- the Freundlich, Langmuir and Langmuir-Sips equation. All the isotherms were calculated by using
- 197 non-linear regression with Origin (Version 7.5) software program. The experiment was carried out
- 199

198

#### 200 3. RESULTS AND DISCUSSION

in triplicate.

## 201 3.1 XRD, SEM-EDX and FT-IR characterization of the zeolitic mixtures

- 202 The specific surface area of the two untreated zeolitic mixtures was  $40\pm2$  g m<sup>-2</sup> for ZNS and  $47\pm2$  g
- $m^{-2}$  for ZNC. XRD analyses revealed that ZNS was a mixture of heulandite, mordenite,
- clinoptilolite and stellerite, whereas ZNC was made only by heulandite and mordenite (Table 1).
- 205 Moreover, the XRD spectra (not reported here) showed also that the NaCl treatment had no impact
- 206 on the mineralogical composition of the zeolitic mixtures.
- 207 The SEM images suggested that NaCl treatment significantly affected surface properties of treated
- 208 zeolitic mixtures (Fig. 1). Indeed, the latter were more regular, close-grained and with well-defined
- 209 edges compared to the untreated ones which showed disordered crystal orientation (Fig. 1). Such
- changes were similar to those reported by Thushari et al. [10] for NaCl treated zeolite.
- 211 Elemental analysis performed by SEM-EDX revealed that, following the treatment with NaCl,
- oxygen, silicon, and aluminium decreased, while Na increased (Fig. S1). The increase of NaCl was

also confirmed by the presence of non-homogeneously distributed NaCl on the surfaces of treated 213 214 zeolitic mixtures. However, the Si/Al, Si/O and Al/O ratios were not affected by NaCl treatment, thus suggesting no alteration in the lattice structure of the two zeolitic mixtures (Table 2). 215 All the zeolitic mixtures showed similar FTIR spectra (Fig. 2). In particular, in the range 500-420 216 cm<sup>-1</sup>, bending of the bonds between tetrahedral (Si, Al) and octahedral species (Al, Fe, Mg) have 217 been observed; the intervals 720-650 cm<sup>-1</sup>, and 1250-950 cm<sup>-1</sup> were due to symmetrical and 218 asymmetrical stretching typical of the internal tetrahedra linkages [24]. Other bands were in the 219 range 650-500 cm<sup>-1</sup> (double ring), 420-300 cm<sup>-1</sup> (pore opening), 820-750 cm<sup>-1</sup> (symmetrical 220 stretching) and 1150-1050 cm<sup>-1</sup> (asymmetrical stretching), typical of external tetrahedra linkages 221 222 (Fig. 2). Indeed, according to Byrappa and Kumar [24] the infrared spectrum useful for the 223 characterization of the structural features of zeolite frameworks is in the region of 1500-400 cm<sup>-1</sup>.

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## 225 **3.2** pH<sub>pzc</sub> of zeolitic mixtures

The pH<sub>pzc</sub> determined according to the pH drift method, ranged from 5.5 to 6.5. It was lower in ZNC and ZSC as compared to ZNS and ZSS. Moreover, it was not affected by NaCl treatment (Fig. 3). Therefore, ZNC and ZSC surfaces show negative charges at pH values lower than that of ZNS and ZSS. Based on such results, it can be argued that ZNC and ZSC should have greater potential for removing ammonium ions from aqueous solution than ZNS and ZSS. However, considering that the pHs of all tested zeolitic mixtures were lower than 5.5, we can conclude that there is no effect of pH<sub>pzc</sub> on zeolite NH<sub>4</sub><sup>+</sup> ion removal from the solution.

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#### 234 **3.3** NH<sub>4</sub><sup>+</sup> adsorption by zeolitic mixtures

The amount of NH<sub>4</sub><sup>+</sup> adsorbed by the two zeolitic mixtures ranged from  $29.3 \pm 0.2$  to  $35.0 \pm 0.4$  mg

 $g^{-1}$  (Fig. 4). Regardless of the treatment for their activation, ZNC and ZSC adsorbed more  $NH_4^+$ 

than ZNS and ZSS. The higher amount of NH<sub>4</sub><sup>+</sup> adsorbed by ZNC and ZSC compared to ZNS and

238 ZSS can be ascribed to concomitant factors such as the higher specific surface area of ZNC

compared to ZNS and to the higher mordenite content, that has been demonstrated to have higher
cation exchange capacity [2]. Indeed, Chen et al. [25] have found that mordenite shows higher
specific surface area as compared to clinoptilolite and heulandite. Also, modernite crystal structure
may contribute to higher NH4<sup>+</sup> adsorption compared to clinoptilolite and heulandite due to the
presence of larger pores and cavities [26].

The amount of  $NH_{4^+}$  adsorbed by the tested zeolitic mixtures was larger than that reported by Lin et al. [27], i.e., 17 mg  $NH_{4^+}$  g<sup>-1</sup>. Although these authors used clinoptilolite with a diameter range like that of the zeolites used in this study (0.8-1.43 mm), they applied a  $NH_{4^+}$  solution for zeolite saturation with a concentration of 990 mg  $NH_{4^+}$  L<sup>-1</sup> which was much lower than that used here (i.e., 20 g L<sup>-1</sup>). This difference can explain the discrepancy between our data and those from Lin et al. [27].

After NaCl treatment, larger Na<sup>+</sup> but lower K<sup>+</sup>, Ca<sup>2+</sup> and Mg<sup>2+</sup> content was observed in the treated zeolitic mixtures rather than in the untreated ones. This indicates that the K<sup>+</sup>, Ca<sup>2+</sup> and Mg<sup>2+</sup> ions on the exchange surfaces of the untreated zeolitic mixtures were replaced by Na<sup>+</sup> after sodium chloride treatment.

The amount of NH<sub>4</sub><sup>+</sup> adsorbed by NaCl treated zeolitic mixtures increased as compared to the untreated zeolitic mixtures. However, such an increase was similar among zeolitic mixtures, thus suggesting no mineralogy effect following NaCl treatment.

These results agreed with previous research on zeolites [10,28,29]. Recently, Thushari et al. [10] 257 suggested that one of the reasons for increased NH<sub>4</sub><sup>+</sup> adsorption by NaCl treated zeolites could be 258 the differences in exchange efficiency of cations associated with their charge. For example, 259 mordenite revealed a selectivity sequence of  $K^+ > NH_4^+ > Ba^{2+} > Na^+ > Ca^{2+} > Mg^{2+}$  [15,30], thus 260 suggesting that monovalent cations are preferentially adsorbed than the divalent ones. Thushari et 261 al. [10] explained the high NH<sub>4</sub><sup>+</sup> by NaCl treated zeolite as it follows: 1) following the NaCl 262 treatment, monovalent  $Na^+$  replace  $Mg^{2+}$  and  $Ca^{2+}$  transforming zeolite into a unified valency 263 structure and facilitating the exchange; 2) the replacement of  $Ca^{2+}$  and  $Mg^{2+}$  with  $Na^{+}$  increases 264

surface area and pore volume of NaCl treated zeolites, thereby increasing their adsorption capacity. 265 266 The large  $NH_4^+$  amount adsorbed by the NaCl treated zeolites could be explained, speculatively, by using the low-density/high-density water model set up by Conte and Schmdit [31] based on what 267 observed by Dhopatkar et al. [32]. This model was elaborated to explain the dynamics of nutrients 268 269 in soils. Namely, on the one hand, a pseudo-ice thin layer is formed on the Na<sup>+</sup>-functionalised zeolitic mixtures surface (that is, zeolitic mixtures surface is a kosmotrope). On the other hand, due 270 to the high charge density, the 20 g  $NH_4^+$  L<sup>-1</sup> solution acts as a kosmotrope, thus leading water 271 molecules to the formation of another pseudo-ice structure (Fig. 5). Due to the pseudo-ice 272 behaviour, we can argue that low-density water (LDW) micro-domains are present in both 273 274 aforementioned systems. As water molecules move away from the kosmotropic environment, high-275 density water (HDW) microdomains can be accounted for. In other words, chaotropic environments are detectable when the transition from LDW to HDW is achieved. 276 277 Due to the density gradient, water molecules included in the HDW domain move towards the LDW one in order to obtain water density homogeneity. However, water molecules do not move alone. 278 They all take part to the hydration shell of the ions present in the system. Consequently, both the 279 sodium ions on the zeolitic mixtures surface and the NH<sub>4</sub><sup>+</sup> ions in the high concentrated solution 280 281 move towards the HDW part of the system. Therefore, negative charged surface zeolite sites are 282 available for NH<sub>4</sub><sup>+</sup> to be adsorbed.

The hypothesised mechanism cannot be applied to the not NaCl treated zeolitic mixtures where the hydrogens of the hydroxyl groups of the zeolite are covalently bound to the oxygens. Therefore, a larger amount of  $NH_4^+$  can be adsorbed on the NaCl treated zeolitic mixtures compared to the untreated ones.

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## **3.4** NH<sub>4</sub><sup>+</sup> monomodal and bimodal pseudo-first order adsorption kinetic models

Adsorption of  $NH_4^+$  by zeolitic mixtures was studied at various time intervals (Figs. 6A and 6B).

290 The amount of  $NH_4^+$  adsorbed by the different zeolitic mixtures after 48 h of incubation was in the

order ZSS > ZNS, and ZSC > ZNC (Figs. 6A and 6B). NH<sub>4</sub><sup>+</sup> adsorption rate by zeolitic mixtures was fast at the beginning of the incubation and then it slowed down (Figs. 6A and 6B). After 2 hours, zeolitic mixtures adsorbed about 80% of NH<sub>4</sub><sup>+</sup> from the solution. The aforementioned results fall within the range observed by Alshameri et al. [33] and Kotoulas et al. [34]. Those authors analysed clinoptilolite with a grain size similar to that used in the present study (i.e., 0.71-1.0 mm), and found that around 80% of NH<sub>4</sub><sup>+</sup> was removed from mono-component solution between one and six hours of incubation.

NaCl treatment had no effect on NH<sub>4</sub><sup>+</sup> adsorption rate by ZSC, whereas it improved that of the ZSS 298 (Figs. 6A and 6B). Indeed, ZSS adsorbed more NH4<sup>+</sup> and saturated faster than ZNS. Such a 299 300 behaviour remains unsolved. It might be ascribed to the different mineralogy of the tested zeolitic mixtures. In fact, ZNS contains less mordenite and more clinoptilolite and stellerite as compared to 301 ZNC (Table 1). Indeed, it could either be related to a different binding strength of the -OH hydroxyl 302 303 groups, which is closely linked to the different framework, or to the different exposure in the framework of the -OH groups on the surface, which may or may not facilitate NH<sub>4</sub><sup>+</sup> binding. 304 The kinetics of NH<sub>4</sub><sup>+</sup> adsorption by the zeolitic mixtures was mathematically analysed by the 305 application of non-linear model reported in the eqs. 1 and 2. 306

The bimodal first-order kinetic model returned the highest R<sup>2</sup> and the lowest  $\chi^2$  values than the 307 monomodal first-order kinetic model, thus suggesting a better fitting of the experimental data than 308 the former model (Table 3). Indeed, although the monomodal first order kinetic model returned also 309 significant high R<sup>2</sup> values, the fitting curve was not able to include some of the experimental points 310 (see sharp-bend in Fig. 6A). Conversely, the bimodal model well fitted all the experimental points 311 312 (Fig. 6B). Based on such results, the sorption process can be described as two simultaneous pseudo first-order reactions where the two sorption constants ( $k_1$  and  $k_2$ ), calculated according to eqs. 1 and 313 2, suggest two different NH<sub>4</sub><sup>+</sup> sorption mechanisms [35] both based on the number of active sites 314 315 present in the solid systems. Based on such findings, we can argument that k<sub>1</sub> is related to the adsorption process of NH4<sup>+</sup> on the zeolitic mixtures surface, whereas k<sub>2</sub> to the ion exchange process 316

between NH<sub>4</sub><sup>+</sup> in solution and cations placed on the zeolitic mixture surface [36]. In fact, both 317 318 untreated zeolitic mixtures revealed the similar  $k_1$  and  $k_2$  values, being in all cases  $k_1 > k_2$ . After the NaCl treatment the kinetic constant remained similar among the treated zeolitic mixtures, but  $k_2 > 1$ 319 k<sub>1</sub>. Such speculation derives from the following deductions: before the NaCl treatment, both 320 monovalent and divalent cations are adsorbed on the zeolitic mixtures, with divalent cations 321 stronger adsorbed than the monovalent ones due to the high charge per unit of surface. In such a 322 323 case, the adsorption process of NH<sub>4</sub><sup>+</sup> is favoured over the ion exchange process. After the NaCl treatment, both monovalent and divalent cations adsorbed on the zeolitic mixtures are substituted by 324 Na<sup>+</sup>. In the latter case, NH<sub>4</sub><sup>+</sup> easily may substitute Na<sup>+</sup> by ion exchange process. Indeed, it is 325 326 conceivable that a faster replacement mechanism of Na<sup>+</sup> with NH<sub>4</sub><sup>+</sup> occurs ( $k_2 > k_1$ ), compared to a slower one  $(k_1 > k_2)$  when NH<sub>4</sub><sup>+</sup> exchanges with all the other cations among which the divalent ones 327 (Ca<sup>2+</sup> and Mg<sup>2+</sup>; Fig. 7). Considering that  $k_1$  is greater than  $k_2$  for zeolitic mixtures not treated with 328 329 NaCl and that the opposite occurs following the treatment with NaCl, we may hypothesize that k<sub>1</sub> is an expression of the adsorption process whereas k<sub>2</sub> of the ion exchange one. Thus, such results also 330 suggested that the treatment with NaCl improved the rate of NH<sub>4</sub><sup>+</sup> sorption and that such 331 improvement does not depend on zeolitic mixtures mineralogy. 332

### 333 **3.5** The isotherms evaluation

Isotherms models that describe the physical-chemical sorption process are important for the 334 interpretation and prediction of sorption data [15,22]. Thus, for the effective use of a natural zeolite 335 as an ion exchanger, different isotherm models have been used to accurately describe the NH<sub>4</sub><sup>+</sup> 336 sorption equilibrium. In addition, isotherms models are needed to describe the equilibrium 337 relationship between the amounts of ions adsorbed by zeolite and their equilibrium concentration in 338 339 the solution. The fitting goodness of the different isotherm models tested in this study (eqs. 3, 4, and 6) was evaluated on the basis of R<sup>2</sup> and  $\chi^2$  values. Based on such values, sorption of NH<sub>4</sub><sup>+</sup> by 340 zeolitic mixtures was significantly consistent with all models, with a preference for the model 341

described by eq. 6 that showed the highest  $R^2$  and the lowest  $\chi^2$  values, respectively (Table 4; Fig. 8A, 8B).

The parameters calculated through the non-linear method by applying eq. 6 to  $NH_4^+$  adsorbed on ZNS and ZSS suggested that the treatment with NaCl increased the number of active sites (lower value of the "n" parameter) leading to higher  $NH_4^+$  adsorption (higher value of  $q_{max}$ ). Such results agree with data obtained from static experiment where  $NH_4^+$  adsorbed by ZSS was higher than that adsorbed by ZNS (Fig. 4).

Also, the equilibrium constant,  $K_L$ , calculated by eq. 6, suggested higher affinity of ZSS for NH<sub>4</sub><sup>+</sup>. Indeed, ZSS showed higher  $K_L$  value compared to ZNS suggesting the shift of the equilibrium Z + NH<sub>4</sub><sup>+</sup>  $\rightleftharpoons$  (Z-NH<sub>4</sub>) towards the associated phase (Z-NH<sub>4</sub>).

The higher affinity of ZSS for  $NH_4^+$  can be attributed to the greater ability of ZSS to form H-bonds or ionic linkages with  $NH_4^+$ . The formation of H-bonds between zeolite and  $NH_4^+$  can be accepted by assuming that the isomorphic substitution of  $AI^{3+}$  with  $Si^{4+}$  results in insufficient positive charges in the crystal lattice [37]. On the one hand, the existence of excessive net negative charges means that some O atoms in the structure exist in unbound form. Therefore, the bound and unbound O atoms are considered as independent sites on which the -H in  $NH_4^+$  can bind [37,38]. On the other

hand, ionic bonding is conceivable since the zeolite and the adsorbate have different charge

359 densities.

With regard to the ZNC and ZSC, the parameters calculated by applying the eq. 6 to NH<sub>4</sub><sup>+</sup> adsorbed data confirmed the absence of NaCl treatment effect on adsorption capacity of zeolitic mixtures.

362 Indeed, the n and K<sub>L</sub> parameters did not show significant differences among ZNC and ZSC; only

 $q_{max}$  was of 3.6 mg of NH<sub>4</sub><sup>+</sup> per g of zeolite higher in ZSC compared to ZNC (Fig. 8B).

364

## 365 4. CONCLUSION

The specific surface area as well as the mineralogical composition of zeolitic mixtures affected their
ability in adsorbing NH<sub>4</sub><sup>+</sup> from a mono-component solution: high presence of mordenite enhanced

such an ability. The treatment with NaCl increased the amount of NH<sub>4</sub><sup>+</sup> adsorbed by the two zeolitic 368 369 mixtures similarly in static adsorption experiment, thus suggesting no dependence of NaCl treatment from the mineralogical composition. The high amount of NH4<sup>+</sup> adsorbed on NaCl treated 370 zeolitic mixtures can be explained using the low density/high density water model firstly proposed 371 for nutrient adsorption by soil. The bimodal pseudo-first-order model explained better than the other 372 tested models the adsorption kinetics of NH<sub>4</sub><sup>+</sup> on zeolitic mixtures. Furthermore, it suggested two 373 374 different mechanisms of NH<sub>4</sub><sup>+</sup> adsorption depending on the size of the cations and on the number of positive charges. Indeed, the two sorption constants,  $k_1$  and  $k_2$ , calculated by the bimodal pseudo-first-375 order model, may be linked to the adsorption (the former) and ion exchange (the latter) processes. The NaCl 376 377 treatment affected the rate of NH<sub>4</sub><sup>+</sup> adsorption of zeolitic mixtures. The Langmuir-Sips model provided the best fit to the equilibrium data. The parameters obtained by applying such model 378 suggested that treatment with NaCl increased the number of active sites of zeolitic mixtures with 379 380 the low amount of mordenite.

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# 389 **References**

- 390 [1] S. Chowdhury, N. Khan, G.H. Kim, J. Harris, P. Longhurst, N.S. Bolan, in: Prasad, M.N.V.
- and Shih, K., Eds., Environmental Materials and Waste, Academic Press, 2016, pp. 569-589.
- 392 [2] J. Langwaldt, Sep. Sci. Technol. 43 (2008) 2166–2182.
- 393 https://doi.org/10.1080/01496390802063937
- J. Volavšek, O. Pliekhov, O. Pliekhova, G. Mali, N.Z. Logar, Nanomat. 12 (2022) 1–13.
   https://doi.org/10.3390/nano12081352
- 396 [4] G. Mannina, L. Badalucco, L. Barbara, A. Cosenza, D. Di Trapani, V.A. Laudicina, S.M.
- 397 Muscarella, D. Presti, Water (Switzerland). 14 (2022) 156.
- 398 https://doi.org/10.3390/w14020156
- 399 [5] T.D. Saliu, N.A. Oladoja, Environ. Chem. Lett. 19 (2021) 2299-2316.
- 400 https://doi.org/10.1007/s10311-020-01159-7
- 401 [6] X. You, C. Valderrama, J.L. Cortina, Sci. Total Environ. 656 (2019) 902–909.
- 402 https://doi.org/10.1016/j.scitotenv.2018.11.420
- 403 [7] M.L. Nguyen, C.C. Tanner, New Zeal. J. Agric. Res. 41 (1998) 427–446.
- 404 https://doi.org/10.1080/00288233.1998.9513328
- 405 [8] S. Wang, Y. Peng, Chem. Eng. J. 156 (2010) 11–24.
- 406 https://doi.org/10.1016/j.cej.2009.10.029
- 407 [9] J. Canellas, Sch. Water, Environ. Energy Res. Degree, Cranf. Univ. (2018).
- 408 [10] D.Thushari, N. Wijesinghe, K.B. Dassanayake, S.G. Sommer, G.Y. Jayasinghe, P. J Scales,
- 409 D. Chen, J. Environ. Sci. Health. 51 (2016) 614–25.
- 410 https://doi.org/10.1080/10934529.2016.1159861
- 411 [11] S.M. Muscarella, L. Badalucco, B. Cano, V.A. Laudicina, G. Mannina, Bioresour. Technol.
- 412 341 (2021) 125812. https://doi.org/10.1016/j.biortech.2021.125812
- 413 [12] Y.M. Channon, C.R.A. Catlow, R.A. Jackson, S.L. Owens, Microporous Mesoporous Mater.
- 414 24 (1998) 153–161. https://doi.org/10.1016/S1387-1811(98)00155-3

- M.N. Khan, A. Sarwar, Surf. Rev. Lett. 14 (2007) 461-469. 415 [13]
- https://doi.org/10.1142/S0218625X07009517 416
- D. Vaičiukynienė, A. Mikelionienė, A. Baltušnikas, A. Kantautas, A. Radzevičius, Sci. Rep. 417 [14]
- 10 (2020) 1–11. https://doi.org/10.1038/s41598-019-55906-0 418
- [15] M. Sarioglu, Sep. Purif. Technol. 41 (2005) 1–11. 419
- https://doi.org/10.1016/j.seppur.2004.03.008 420
- 421 [16] M. Moshoeshoe, M. Silas Nadiye-Tabbiruka, V. Obuseng, Am. J. Mater. Sci. 7 (2017) 196-221. https://doi.org/10.5923/j.materials.20170705.12 422
- N.Z. Misak, React. Funct. Polym. 43 (2000) 153-164. https://doi.org/10.1016/S1381-423 [17]
- 424 5148(99)00046-2

434

- D. Sparks, Environmental Soil Chemistry. second ed., Academic Press, Oxford, 2003. 425 [18]
- P. Conte, R. Bertani, P. Sgarbossa, P. Bambina, H.-P. Schmidt, R. Raga, G. Lo Papa, D.F. 426 [19]
- Chillura Martino, P. Lo Meo, Agronomy. 11 (2021) 615. 427
- https://doi.org/10.3390/agronomy11040615 428
- N. Widiastuti, H. Wu, H.M. Ang, D. Zhang, Desalination. 277 (2011) 15-23. 429 [20]
- https://doi.org/10.1016/j.desal.2011.03.030 430
- M. Belhachemi, F. Addoun, Appl. Water Sci. 1 (2011) 111–117. 431 [21]
- https://doi.org/10.1007/s13201-011-0014-1 432
- R. Malekian, J. Abedi-Koupai, S.S. Eslamian, S.F. Mousavi, K.C. Abbaspour, M. Afyuni, [22] 433 Appl. Clay Sci. 51 (2011) 323–329. https://doi.org/10.1016/j.clay.2010.12.020
- 435 [23] R.L. Mulvaney, in: D.L. Sparks (ed.), Methods of soil analysis. Part 3-Chemical methods, Madison, 1996, pp. 1123-1184. 436
- K. Byrappa, B.V.S. Kumar, Asian J. Chem. 19 (2007) 4933–4935. ISSN 0970-7077 [24] 437
- H.F. Chen, Y.J. Lin, B.H. Chen, I. Yoshiyuki, S.Y.H. Liou, R.T. Minerals. 8 (2018) 499. 438 [25]
- https://doi.org/10.3390/min8110499 439
- 440 [26] O.Y. Korkuna, T.Y. Vrublevska, Stud. Surf. Sci. Catal. 158 (2005) 1035–1042.

- 441 https://doi.org/10.1016/s0167-2991(05)80445-8
- 442 [27] L. Lin, Z. Lei, L. Wang, X. Liu, Y. Zhang, C. Wan, D.J. Lee, J.H. Tay, Sep. Purif. Technol.
  443 103 (2013) 15–20. https://doi.org/10.1016/j.seppur.2012.10.005
- 444 [28] N.A. Booker, E.L. Cooney, A.J. Priestley, Water Sci. Technol. 34 (1996) 17–24.
- 445 https://doi.org/10.2166/wst.1996.0167
- 446 [29] S. Komarowski, Q. Yu, Environ. Technol. 18 (1997) 1085–1097.
- 447 https://doi.org/10.1080/09593331808616628
- 448 [30] D.L. Bish, D.W. Ming, Natural Zeolites : Occurrences, Properties, Applications, Rev.
- 449 Mineral. Geochemistry. 45 (2001) 207–210. https://doi.org/10.2138/rmg.2001.45.0
- 450 [31] P. Conte, H.-P. Schmidt, EMagRes, John Wiley & Sons, Ltd, 6 (2017) 453–464.
- 451 <u>https://doi.org/10.1002/9780470034590.emrstm1535</u>
- 452 [32] N. Dhopatkar, A.P. Defante, A. Dhinojwala, Sci. Adv. 2 (2016) e1600763.
  453 https://doi.org/10.1126/sciadv.1600763
- 454 [33] A. Alshameri, C. Yan, Y. Al-Ani, A.S. Dawood, A. Ibrahim, C. Zhou, H. Wang, J. Taiwan
- 455 Inst. Chem. Eng. 45 (2014) 554–564. https://doi.org/10.1016/j.jtice.2013.05.008.
- 456 [34] A. Kotoulas, D. Agathou, I.E. Triantaphyllidou, T.I. Tatoulis, C.S. Akratos, A.G.
- 457 Tekerlekopoulou, D. V. Vayenas, Water (Switzerland) 11 (2019) 136.
- 458 https://doi.org/10.3390/w11010136
- 459 [35] R.A. Griffin, J.J. Jurinak, Soil Sci. Soc. Am. J. 38 (1974) 75–79.
- 460 https://doi.org/10.2136/sssaj1974.03615995003800010026x
- 461 [36] S.E. Jørgensen, in: S.E. Jørgensen, M.J. Gromiec (Eds.), Adsorption and Ion Exchange, ,
  462 Math. Submodels Water Qual. Syst., Elsevier, 1989, pp. 65–81.
- 463 [37] P. Liu, Y. Liu, A. Zhang, Z. Liu, X. Liu, L. Yang, Water (Switzerland) 14 (2022) 2413.
  464 https://doi.org/10.3390/w14152413
- 465 [38] G. Gilli, P. Gilli, M. Khosravi, V. Murthy, I.D.R. Mackinnon, J. Mol. Struct. 552 (2019) 1-
- 466 17. https://doi.org/10.3390/molecules24203652.