

1 **Ingestion of plastic and non-plastic microfibers by farmed gilthead sea bream (*Sparus aurata*)**  
2 **and common carp (*Cyprinus carpio*) at different life stage**

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34 **Graphical Abstract....(in progress)**

35 **Abstract**

36 Environmental pollution from plastic particles is a major global concern, being a potential threat to  
37 aquatic organisms and ecosystems. The accumulation of microplastics in the freshwater and marine  
38 environments has strong ecological implications due to their long persistence, their potential toxicity,  
39 and their ability to adsorb other pollutants and act as vectors of pathogens. Nevertheless, whereas the  
40 number of studies on the presence of microplastics in wild fish has increased, less attention has been  
41 paid to farmed fish species. Here, we investigated the occurrence of microplastics in the digestive  
42 tracts of gilthead sea bream (*Sparus aurata*) and common carp (*Cyprinus carpio*) at different life  
43 stage and reared by an intensive and semi-intensive production system, respectively. Our results  
44 showed the presence of natural microfibrils and microplastics including fibers and fragments in both  
45 species, with microfibrils (~ 90 %) being the dominant type. In both fish species, the presence of  
46 microparticles was not revealed at larval stage. Fry and adult gilthead sea bream specimens showed  
47 microfibril abundances of 0.21 and 1.3 items/individual, respectively. A lower load of microparticles  
48 ( $p < 0.05$ ) occurred in fry (0.06 items/individual) and adult common carp specimens (0.25  
49 items/individual). As to the chemical composition of the microitems, natural (cotton 16%, linen 4%),  
50 semi-synthetic (rayon 24%, lyocell 4%), and single or blended synthetic fibers (cotton:polyamide  
51 12%, cotton:polyester 4%, wool:polyester 4%, nylon 8%, polyester 8%, polyacrylic 4% and PTFE  
52 12%) were identified in gilthead sea bream. Linen, rayon, lyocell, cotton:polyester and polyester  
53 (12.5% concentration for each polymer) fibers were identified in common carp, while PTFE (37.5%)  
54 was present as fragments. Rayon was the most frequent chemical type (21.2%), followed by PTFE  
55 (18.18%). Polymer composition of extracted microparticles showed significant differences between  
56 the fish species analysed in this paper ( $p < 0.05$ ). Notably, a considerably lower contamination level  
57 of synthetic polymers (average 0.11 items/individual) was detected in farmed fishes compared with  
58 that found in wild specimens. To the best of our knowledge, this is the first study reporting plastic  
59 and non-plastic microfibril contamination in farmed gilthead sea bream and common carp at different  
60 life stage.

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62 **Keywords:** microparticles, microplastics, *Sparus aurata*, *Cyprinus carpio*, aquaculture

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67 **Introduction**

68 Microplastics (MPs), defined as plastic particles ranged between 100 nm - 5 mm in size (Cole et al.,  
69 2011), have become a constitutive part of environmental litter in aquatic and terrestrial ecosystems  
70 around the world (Alimba & Faggio, 2019; Barnes et al., 2009; Halpern et al., 2008). In particular, in  
71 the marine environment, plastic waste represents the most abundant litter category, which can amount  
72 to more than 80% of the debris reported. Both maritime and terrestrial anthropogenic activities are  
73 responsible for the continued input of plastic in aquatic environment, making it a ubiquitous pollutant  
74 (ref).

75 The main problem associated with MPs is their bioavailability to a variety of aquatic animals. Many  
76 studies have demonstrated the occurrence of MPs in commercially wild-caught fish and shellfish  
77 species (Giani et al., 2019; Mancuso et al., 2019; Bottari et al., 2019; Fang et al., 2019; Su et al.,  
78 2019; Romeo et al., 2015). In contrast, our knowledge of MPs ingested by rearing aquatic animals  
79 remains scarce (Hanachi et al., 2019; Ma et al., 2020; Wu et al., 2020). MPs tend to be found more  
80 frequently in the gastrointestinal tract of aquatic organisms (Savoca et al., 2020; Capillo et al., 2020),  
81 and their ingestion can not only cause mechanical damage (Jin et al., 2018; Lei, Liu, et al., 2018; Qiao  
82 et al., 2019), but also induce metabolic disturbances such as oxidative stress, suppression of  
83 detoxification in other vital tissues, as well as alteration of the immune system (Lei, Wu, et al., 2018;  
84 Yu et al., 2018). Moreover, MPs can act as vector of toxic compounds such as heavy metals, POPs  
85 and PCBs (Miranda and Freire de Carvalho-Souza 2016; Guo and Wang 2019; Wang and Chen 2019;  
86 Rochman et al. 2019), enhancing their bio-toxicity (Rodríguez-Seijo, Santos, da Silva, Cachada and  
87 Pereira, 2019). Some studies have also highlighted how the reproductive process of aquatic animals  
88 can be compromised by their exposure to MPs (Pitt et al., 2018; Sussarellu et al., 2016). Furthermore,  
89 these toxic contaminants can be transferred along the food chain through bioaccumulation and  
90 biomagnification (Van Cauwenberghe and Janssen, 2014). All these adverse effects caused by MPs  
91 undoubtedly represent a serious threat to the aquaculture industry and its sustainability. In this context  
92 research is needed to assess the risk of ingestion of environment-derived and farming- derived MPs  
93 on commercially species, since the consumption of aquatic products is considered the main key  
94 pathway for the potential human microplastic ingestion. Only recently, in the Persian Gulf, the  
95 average intake of MPs from fish muscle consumption was estimated between 169 and 555 elements  
96 per 300g of muscle (Akhbarizadeh, Moore, and Keshavarzi, 2018). However, whereas the number of  
97 investigations documenting the presence of MPs in wild fish has increased, few studies have  
98 addressed the presence of MPs in farmed fish species. For instance, Ma et al., 2020 provides evidence  
99 of MPs occurrence in aquaculture water in Pearl River Estuary of Guangzhou (China) showing how  
100 MPs abundance was higher compared to other areas worldwide. Wu et al., 2020 investigated the

101 accumulation of MPs in commercial aquatic species collected from the aquaculture sites at Xiangshan  
102 Bay (China) showing how farmed species are not exempt from the risk of exposure to plastic litter.

103 Recently, other contaminants of emerging concern have been identified in non-synthetic (also named  
104 as natural) and semi-synthetic "microfibers". The terms refer respectively to anthropogenic fibers  
105 from textiles of natural plant or animal origin (i.e. cotton, wool), and derived cellulosic sources (i.e.  
106 viscose/ rayon) worldwide distributed (Savoca et al., 2019; Suaria et al., 2020). Natural and semi-  
107 synthetic microfibers are rarely documented and not counted in assessing marine environment impact,  
108 resulting in underestimation of their potential threat. Despite the attention of research is today highly  
109 focused on plastics pollution, recent studies (Gago et al., 2018) discovered that anthropic fibers are  
110 also very common (Almroth et al., 2018; Barrows et al., 2018; Gago et al., 2018; Henry et al., 2019;  
111 Remy et al., 2015; Sanchez-Vidal et al., 2018). Natural and semi-synthetic fibers have been mostly  
112 observed in ingestion studies (Lusher et al., 2013; Remy et al., 2015; Rochman et al., 2015; Zhao et  
113 al., 2016), and, although they may not represent in essence an environmental issue, the artificial  
114 colorants, additives or flame retardants (commonly used during textile production)(R. R. Mather,  
115 R.H. Wardman, *The Chemistry of Textile Fibres* (The Royal Society of Chemistry, 2015),, and the  
116 chemicals they can accumulate from the aquatic environment raise concerns about their role as  
117 vectors of dangerous substances for marine ecosystems .(F. S. Cesa, A. Turra, J. Baruque-Ramos,  
118 *Synthetic fibers as microplastics in the marine environment: A review from textile perspective with  
119 a focus on domestic washings*. *Sci. Total Environ.* 598, 1116–1129 (2017)

120 In this work we examined the load of microfibers and microplastics present in the gastrointestinal  
121 tracts (GITs) of *Sparus aurata* (gilthead sea bream) and *Cyprinus carpio* (common carp) at different  
122 life stages, reared in Italy using an intensive and in Croatia by means of a semi-intensive production  
123 system respectively.

124 Gilthead sea bream (GSB) is a carnivorous sparid that inhabits the Atlantic European coast from  
125 Portugal to the United Kingdom, including the Mediterranean and the Black Seas (Froese and Pauly,  
126 2020). Commercial farming started in the 1980s, spreading from Italy and France to the rest of the  
127 Mediterranean countries. GSB culture has increased considerably in the last few years, reaching a  
128 high production and a high commercial value. Today, GBS is the most important finfish aquaculture  
129 product in the Mediterranean with a total production of 136,000 t in 2010  
130 (<http://www.feap.info>). Common carp is an omnivorous fish, and is the most widely distributed  
131 freshwater fish species across the globe (Froese and Pauly, 2020) and the third most important  
132 aquaculture species in the world. As one of the dominant cyprinid species, common carp is cultured

133 in over 100 countries with a total production of 3 million metric tons of global annual freshwater  
134 aquaculture production (FAO, 2006; Bostock et al., 2010).

135 The aims of this study were: i) to quantify the load of MPs in the digestive tract of GSB and common  
136 carp, starting from the larval stages until reaching the adult size ii) identify any differences between  
137 the characteristic of microparticles (size and color) extracted from the two species at different life  
138 stages and possibly ii) to identify the polymer composition of particles isolated and highlight any  
139 differences in particles composition between all fish groups investigated.

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## 142 **Materials and Methods**

143

### 144 **Fish samples**

145 Reared specimens of larvae, fry, and adults of GSB and common carp, were collected from two fish  
146 farms, located in Italy and Croatia, during May 2018 and 2019 respectively.

147 Seven hundred GSB larvae and seven hundred ninety- five common carp larvae were collected,  
148 placed in sterile glass containers and examined for microplastic content.

149 For both species, a total of 60 fry (26 days-old GSB and 6 days-old common carp) and 20 adult  
150 samples specimens were weighed and measured directly in the field, and, subsequently, wrapped with  
151 aluminium foil, and transported at 4°C to the laboratory. GSB specimens were analysed at the  
152 Department of Chemical, Biological, Pharmaceutical and Environmental Sciences, University of  
153 Messina, while common carp samples were processed at the Department for Biology and Pathology  
154 of Fish and Bees, Faculty of Veterinary Medicine, University of Zagreb. The assays were performed  
155 by the same operator and using the same methodology.

156

### 157 **Microplastic extraction protocol**

158 Once in the laboratory, all the samples were washed with deionized water to eliminate any external  
159 contamination. Larvae, fry and adult specimens of GSB and common carp were counted, measured  
160 and undertaken to chemical digestion. Larval samples of both species were digested in two separated  
161 pools (Table 1).

162 GITs of fry individuals, were digested in pools of 5 samples, while adult GITs were digested  
163 individually for both species. The intestine and hepatopancreas from adult common carp specimens  
164 were separated and treated individually.

165 All samples were processed adopting a modified version of the chemical digestion protocol  
166 previously suggested by Savoca et al. (2020).

167 Briefly, samples were placed in a conical glass flask. After adding a calculated quantity of 10% KOH  
168 solution (minimum ratio 1:5 w/v), the flask was covered with aluminium foil. To remove the organic  
169 matter, the flask was placed in an oscillation incubator to be continuously stirred at 50°C for 48 h.

170 Each sample was then put into a graduated glass cylinder adding hypersaline NaCl solution (15%) to  
171 obtain separation of the two phases by density. The supernatant was collected in a glass beaker, and  
172 doubly filtered through a glass fibre membrane with 1.5 mm and 0.7 mm pore size and 47 mm  
173 diameter (Whatman GF/F, UK) using a vacuum system (Millipore). After filtration procedures, the  
174 membranes were placed in sterile Petri glass dishes for subsequent observations under the  
175 stereomicroscope (Leica M205C) to isolate plastic debris. The isolated suspected microplastic were  
176 recorded and categorized based on their shape (fibres and fragment), size and colour.

177 Then a subsample was assayed for the chemical characterization.

178

#### 179 Contamination prevention

180 Workspaces and tools were cleaned from any particles according to Bottari et al. (2019). All materials  
181 used for the dissection, the extraction steps and the analysis were rigorously cleaned with ethanol and  
182 filtered deionized water. The same preventive measures used for sample contamination were adopted  
183 during the digestion procedures. In addition, deionized water, potassium peroxide, and hypersaline  
184 solution were always pre-filtered (0.45 mm filter). Only sterilized glass items were used for all the  
185 assays. Fish dissection and digestion protocols were performed in a clean air flow cabinet to exclude  
186 external contamination from fibres, which might represent a major source of contamination.

187 A paper filter put in Petri dishes was exposed to the laboratory air and used as control (blank) during  
188 the entire laboratory procedure.

189

#### 190 Microplastic characterization

191 The chemical composition of isolated microfibers and micro-fragments were identified by micro-  
192 infrared spectroscopy ( $\mu$ -FT-IR). Prior to each measurement, a microscopic image of each sample  
193 was taken.  $\mu$ -FT-IR spectra were recorded using a Bruker FTIR LUMOS microscope equipped with  
194 a liquid nitrogen cooled  $64 \times 64$  detector. Infrared spectra were recorded in transmission method in  
195 the range  $4000\text{--}900\text{ cm}^{-1}$  with a resolution of  $4\text{ cm}^{-1}$ . Background and baselines of recorded spectra  
196 were calculated and, if necessary, subtracted to the spectra, with Origin 9.0 software. To identify the  
197 polymers, the obtained spectra were compared with the multiple libraries provided by the Bio Rad  
198 KnowItAll FTIR library. Only spectra matched over 80% with the standard database were accepted.

199 To identify the natural, semi-synthetic and synthetic textile materials in the fibre samples, the spectral  
200 data collected by Peets and collaborators (Peets et al., 2017) were used. In this way, we were able to  
201 distinguish different kinds of single- and two-component mixed textiles.

202

203

204 Data analysis

205 The Wilcoxon-Mann-Whitney test was performed to detect significant differences in microplastics  
206 abundance between the two fish species and between the life stage group of each species ( $p < 0.05$ ). A  
207 Kruskal-Wallis one-way ANOVA and Tukey's test was performed to determine whether  
208 microparticle characteristics (size, colour and polymer composition) were significantly different  
209 between the fish groups investigated. Univariate statistical analysis was performed using Sigmaplot  
210 V. 14.5. Non-parametric multi-dimensional scaling (nMDS) were performed to highlight any  
211 microparticle feature similarities between the fish groups. After data square root transformation, the  
212 Bray-Curtis similarities were calculated. Statistical analyses were performed using PRIMER6-E.

213

214

## 215 **Results**

216 Number of specimens analysed and morphological characteristics, including the standard length of  
217 larvae (SL, mm), total body length (TL, cm) and the body weight (W, g) of fry and adults of both  
218 species are reported as means  $\pm$  SD in Table 1. The number and the corresponding chemical types of  
219 the identified items found in the two species, are summarized in Tables 2 and 3.

220 Throughout the rest of the paper, we will use the term "microparticle" as a neutral term to refer to  
221 both microplastics (filaments or fragments) and microfibers. Furthermore, microfibers that are  
222 identified as blend of synthetic and non-synthetic materials have been included with the synthetic  
223 microfibers.

224

### 225 Microplastic in *Sparus aurata*

226 In total, 39 microparticles were isolated from the GITs of both fry and adult specimens (0.48  
227 items/specimen), while no particles were detected at larval stage. All of them appeared to be fibers,  
228 ranging in size from 0.24 to 8.86 mm (Table 1). Representative images of microparticles found are  
229 shown in Figure 1. 33.3% were isolated from fry, while 66.6% from adult individuals ( $p > 0.05$ ). The  
230 most dominant colours were black (46.15%), followed by azure (20.5%) (Fig. 2 a,b) ( $p > 0.05$ ). A total  
231 of 25 microfibers isolated from GSB were identified. Regarding the microparticles composition,

232 natural (cotton, linen), semi-synthetic cellulose-based (rayon, lyocell), and synthetic (polyamide,  
233 nylon, polyester, polyacrylic and PTFE) polymers were identified in GSB (Table 2). Some  $\mu$ -FT-IR  
234 example spectra of the fibers found in GSB-are shown in Figure 4.

235 In detail, isolated fibers from fry showed a polymeric composition consisting of 22.2% of fibers of  
236 natural origin (cotton), 33.3% of semi-synthetic fibers (rayon) and 44.4% from synthetic fibers  
237 (polyester, nylon, cotton:polyamide) (Table 2). Polymeric composition of microfibers isolated from  
238 adults was characterized by a 18.75% of natural fibers (cotton and linen). Semi-synthetic fibers  
239 accounted for 25% (rayon and lyocell), while synthetic polymers were the most abundant (56.25%),  
240 presenting additionally cotton:polyester, wool:polyester, PTFE and polyacrylic. No significant  
241 differences were observed between the chemical composition of microparticle isolated from fry and  
242 adults of GSB ( $p>0.05$ ).

243

#### 244 Microplastic in *Cyprinus carpio*

245 Even then, no MPs were observed in common carp at larval stage. A total of 9 microparticles were  
246 isolated (0.11 items/specimen) from the GITs of common carp specimens, whose representative  
247 images are shown in Figure 1. 44.4% were isolated from fry, while 55.5% from adult individuals (Fig.  
248 3). Microfibers represented 55.5%, while the micro-fragments constituted 44.4%, all of them ranging  
249 in size from 0.07 to 2.23 mm. The dominant colour was azure (55.5%), followed by black (22.2%),  
250 light blue (11.1%) and blue (11.1%) (Fig. 3 a, b, c). From the examination of the hepatopancreas no  
251 microparticles were revealed. 8 of 9 -items isolated from common carp were identified as natural and  
252 artificial cellulose-based polymers (cotton, rayon, lyocell, linen), polyester, and PTFE (Table 2).  
253 Some  $\mu$ -FT-IR example spectra of the different microparticles found in carps are shown in Figure 5.

254 As with GSB, common carp specimen also showed a higher percentage of synthetic polymers in both  
255 age groups investigated (Table 3).

256 In adult individuals no items of natural origin were observed. Semi-synthetic polymer was  
257 represented by lyocell (25%), while synthetic polymers were the most representative chemical class  
258 identified (75%, including cotton:polyester and PTFE).

259

#### 260 Comparison between microparticles detected in *Sparus aurata* and *Cyprinus carpio*

261 Microparticles abundance and size were significantly different between GBS and common carp  
262 ( $p<0.05$ ). However, no relevant differences were detected between fry and adult specimens of both  
263 species ( $p>0.05$ ) (Fig. 6), as such as in microparticle color distribution among examined samples  
264 ( $p>0.05$ ).



265 Microparticles polymer composition was significantly different between adult individuals of GSB  
266 and common carp ( $p<0.005$ ) and between adult of GBS and common carp fingerling ( $p<0.05$ ). The  
267 nMDS showed a high similarity of polymer composition (60%) between fry and adult of GBS, while  
268 a 40% similarity was found between them in common carp fry (Fig.7).

269

270

## 271 **DISCUSSION**

272 The widespread presence of microplastics in aquatic environments, both marine and freshwater, has  
273 attracted the attention of the scientific community. Microplastics may severely impact biotic and  
274 abiotic compartments of aquatic ecosystems. The ingestion of microplastics has been well  
275 documented in several marine and freshwater fish species (Bottari et al., 2019; Capillo et al., 2020;  
276 S. Savoca et al., 2019, Steer et al., 2017) and the literature on the subject is constantly increasing.  
277 However, only few studies evaluated the microplastic and man-made fiber pollution in farmed fish  
278 species (Ma et al., 2020, Wu et al., 2020; Lv et al., 2020). As such, the aquaculture industry may  
279 suffer from environment and farming derived microplastic pollution, especially as plastic products  
280 are widely used for aquaculture. This study, to the best of our knowledge, is the first report on the  
281 presence of non-synthetic and synthetic microfibers and fragments in the two farmed fish, gilthead  
282 sea bream and common carp, in European waters.

283 In terms of number of microitems, both fish species showed lower abundance of microparticles than  
284 their wild counterparts (Güven et al., 2017; Zheng et al., 2019), although GSB showed a greater  
285 accumulation than carp (39 and 9 items respectively) ( $p<0.05$ ). In both species, no microparticles  
286 were found at larval stage. This is not surprising considering that larvae, in both farms, are raised  
287 inside a hatchery, equipped with filtration systems that probably mitigate the entry of microplastics  
288 through the water. This finding is in contrast to what has been reported in open water studies, where  
289 microplastics have recently been found in the digestive tract of wild fish larvae and juveniles  
290 belonging to commercially important species of the English Channel and the Mediterranean Sea  
291 (Savoca et al., 2020; Steer et al., 2017).

292 In GSB, the number of microfibers found in the fry (0.21 items/specimen) is lower than in adult  
293 specimens (1.3 items/specimen). This difference can be linked to the production phases, in fact, the  
294 fingerlings are raised in raceways or in tanks within the hatchery facility, while the adults are  
295 intensively reared in offshore sea cages. Therefore, adult specimens are more exposed to  
296 environment-derived microdebris. Existing data on the ingestion of microplastics by GSB are relative  
297 to wild specimens in the Turkish Mediterranean waters (Güven et al., 2017). The authors reported a

298 microparticle abundance of 1.53 items/specimen in wild GSB, that was significantly lower than  
299 microparticle abundance (0.48 items/specimen) found in the present study.

300 A low number of microparticles was found in common carp specimens (fry: 0.06 items/specimen;  
301 adult: 0.25 items/specimens) showing no significant difference in their number between fry and adult.  
302 Studies carried out in natural waters have shown in many cases low ingestion levels of plastic debris  
303 (0.2 items/individual) in wild common carp (Pazos et al., 2017; Zheng et al., 2019).

304 According to previous observations (Bottari et al., 2019; Savoca et al., 2019; Wu et al., 2020) the  
305 MPs were found mostly in fiber shape in both species (100% in gilthead sea bream and 56 % in  
306 common carp). Jabeen et al. 2017 in a study on Chinese common carp highlighted that only fibres  
307 (100%) and no one fragment or other type of MPs were found. In addition, in the specimens that  
308 weighed between 270 ÷150 g and measured between  $28 \pm 5.7$  cm, they found  $2.5 \pm 1.3$   
309 items/individual. This is probably linked to the site pollution. The low number of microparticles found  
310 could depend on two factors: 1) the location of the fish rearing plant and the level of contamination  
311 of the supplied water (Pazos et al., 2017) and 2) the level of plastic contamination present in  
312 commercial feed (Hanachi et al., 2019). We, therefore, assume that in this study, both the farming  
313 environment and the feed presented low concentrations of microplastics.

314 Regarding particles size, much of the debris (30.7%) found in GSB ranged between 1-2 mm, while  
315 66.6% of the MIs found in common carp specimens were smaller than 1mm.

316 So, assuming that the species could not discern the size of particles for ingestion, such differences  
317 may be mediated by biological processes, such as mastication or digestion, which could modify the  
318 size of microplastics (ref).

319 Our most surprising result was that numerous fibers, initially visually identified as microplastic  
320 (actually 50% of identified fibers), were instead classified as semi-synthetic (30% of identified fibers)  
321 or non-synthetic fibers (20%) through more detailed analysis (Table 2). This indicates that semi-  
322 synthetic and non-synthetic microfibers could be a significant and overlooked pollutant in aquatic  
323 environment.

324 The composition of most polymers is the typical one of textiles fibers. Microfibers are generally  
325 identified as secondary microplastics, which are mainly released from synthetic clothing during  
326 washing processes. Typically, these microfibers are made up of materials such as nylon, polyethylene  
327 terephthalate and polypropylene (Gago et al., 2018). Most of the microfibers accumulated in the  
328 aquatic environment are released by textile industries, recycling processes, regular domestic drainage,  
329 direct discharge of garments into the sea or rivers (Almroth et al., 2017).

330 Thus, we suggest that the source of this microfiber pollution might be mainly from rivers, in the case  
331 of common carp contamination, and maritime activities in the case of GSB specimens, as well,  
332 obviously from the production system themselves (Lv et al., 2020)

333 The presence of a higher percentage of cellulose-based polymers fibers is in accordance with what  
334 has been recently observed in a study on the accumulation of microplastics in farmed aquatic species  
335 (Wu et al., 2020). It is interesting to note that in the present study polyethylene (PE), one of the most  
336 used polymers in aquaculture for ropes and floating rigs (Andrady, 2011), was not found in the tested  
337 fishes. Polyethylene has a low density ( $0.857 - 0.975 \text{ g cm}^{-3}$ ), and, rather than sink on the seabed, it  
338 tends to float on the water surface, thus being for instance, unavailable for the feeding behavior of  
339 *Cyprinus carpio* species. Conversely, the specific density of cotton/cellulose ( $1.54 - 1.63 \text{ g cm}^{-3}$ ) is  
340 higher than that of polyester ( $1.37 - 1.46 \text{ g cm}^{-3}$ ) and nylon/acrylics ( $1.14 - 1.18 \text{ g cm}^{-3}$ ), so this could  
341 be explain the higher ingestion of cellulosic fibers by the studied fish species.

342 In any case, the degree of contamination of the geographic location appears to have a greater influence  
343 on high MP abundance values. For example, it has been shown that MPs found in China's inland  
344 waters were much more abundant than European ones (Wang et al., 2017).

345

## 346 **Conclusion**

347 As emerging contaminants, microplastics and microfibers have been found ubiquitously in both  
348 farmed sea and freshwater fish species, indicating their widespread distribution and contamination.  
349 This study provides the first investigation on the ingestion and characteristics of plasti and non-plastic  
350 microparticles in farmed gilthead seabream and common carp from European waters. Moreover the  
351 abundance level of microparticles is lower in farmed species than that reported in other natural and  
352 aquaculture areas worldwide. No differences of microparticles abundance were observed among fish  
353 life stages investigated, although this was significantly different between the two species analysed in  
354 this study. Microplastics were mainly observed in fibrous shape, consisting mainly of semi-synthetic  
355 (30%) and synthetic materials (50%). Future research needs more extensive monitoring of  
356 microfibers in aquaculture products for a better understanding of the role of aquaculture activity in  
357 microparticles accumulation. These results represent an important baseline in assessing cultured  
358 species food safety in term of microplastic ingestion demonstrating that fish farming could help in  
359 the reduction of human consumptions of MP contaminated fish.

360

361 TABLES

362

363 **Table 1.** Data (length and weight) of the analysed samples of *Sparus aurata* and *Cyprinus carpio*, and corresponding number of  
 364 microplastic particles (MPs)  
 365

Species		N° of samples	Lenght (cm)	Weight (g)	N° MPs	Item/specimen	Particles size
			(Mean±SD)				(Mean±SD)
<i>Sparus aurata</i>	Larvae	700	7.5±0.3		0		
	Fry	60	6.84±0.49	5.41±1.13	13	0.21	1.84±1.29
	Adult	20	25.6±1.7	253±2.17	26	1.3	1.96±1.72
<i>Cyprinus carpio</i>	Larvae	795	5.81±0.3		0		
	Fry	60	7.11±1.19	10.9±1.17	4	0.06	0.81±0.64
	Adult	20	51.18±2.71	2740±0.43	5	0.25	0.80±1
<b>Total</b>		160			48		

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368 **Table 2.** Polymer composition of the identified items in the two investigated species.

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Species	Sample	Stage	ItēN°ms	Chemical type	Shape
<i>Sparus aurata</i>	A	Fry	5	Cotton: Polyamide, Rayon, Rayon, Polyester, Cotton	Fibers
	B	Fry	2	Cotton: Polyamide, Cotton	Fibers
	C	Fry	2	Rayon, Nylon	Fibers
	D	Adult	5	Rayon, Cotton: Polyamide, Polyacrylic, Cotton, PTFE	Fibers
	E	Adult	3	Nylon, Polyester, PTFE	Fibers
	F	Adult	4	Rayon, Cotton: Polyester, Wool: Polyester, Linen	Fibers
	G	Adult	2	Lyocell, PTFE	Fibers
	H	Adult	2	Rayon, Cotton	Fibers
<i>Cyprinus carpio</i>	I	Fry	3	Polyester, PTFE, Linen	Fiber, Fragment, Fiber
	J	Fry	1	Rayon	Fiber
	K	Adult	1	Cotton: Polyester	Fiber
	L	Adult	1	PTFE	Fragment
	M	Adult	2	Lyocell, PTFE	Fiber, Fragment

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373 **Table 3.** Chemical type of the identified items and their percentages.

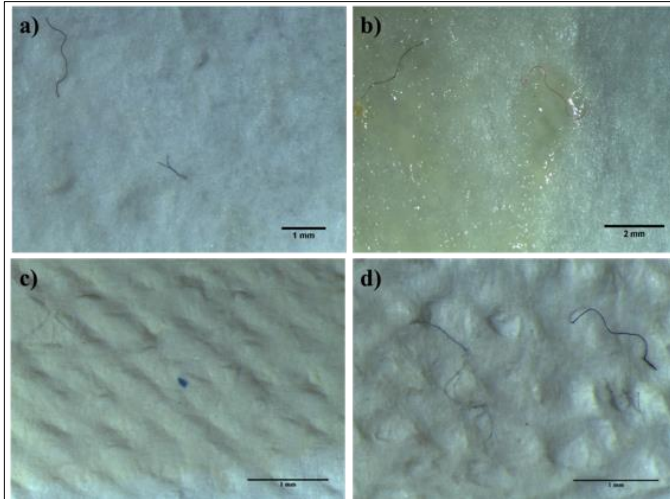
	<b>Chemical type</b>	<b>N° of items</b>	<b>Percent (%)</b>	<b>Percent per class (%)</b>
<b>Natural/Artificial</b>	Linen	2	6.06	18.18
	Cotton	4	12.12	
<b>Semi-synthetic</b>	Rayon	7	21.2	27.27
	Lyocell	2	6.06	
<b>Synthetic/Plastic</b>	Cotton: Polyester	2	6.06	54.54
	Cotton: Polyamide	3	9.09	
	Wool: Polyester	1	3.03	
	Polyester	3	9.09	
	Nylon	2	6.06	
	Polyacrylic	1	3.03	
	PTFE	6	18.18	

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380 **FIGURES**

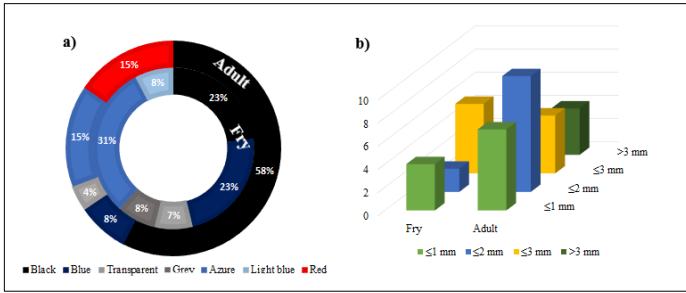


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382 **Figure 1.** Representative images of microplastics found in fry (a) and adult specimens (b) of *Sparus aurata* and in fry (c) and adult  
383 specimens (d) of *Cyprinus carpio*.

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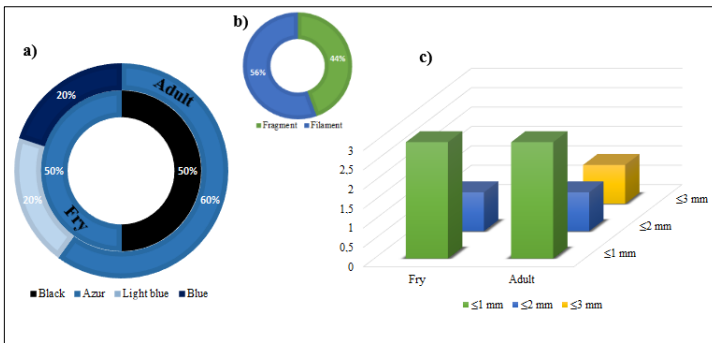
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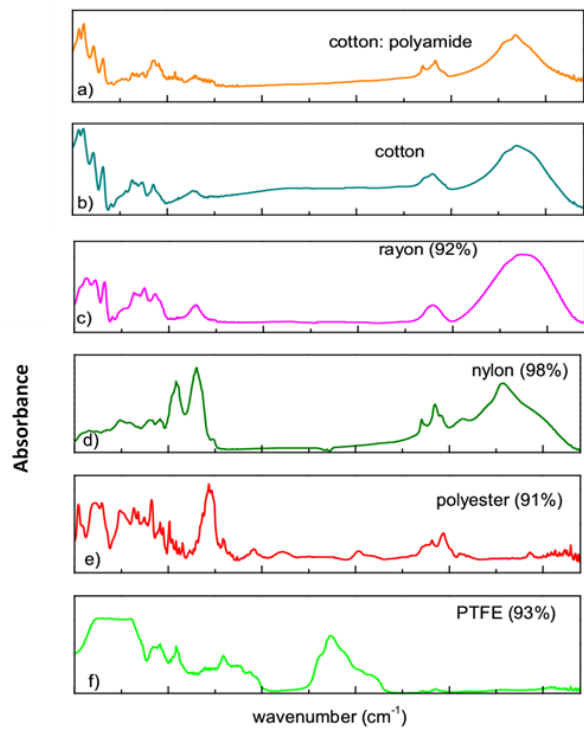
387 **Figure 2.** Percentage (%) of plastic particles classified by colour (a) and size (b) extracted from the gastrointestinal tract of reared fry  
388 and adult *Sparus aurata*.

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391 **Figure 3.** Percentage (%) of plastic particles classified by colour (a) shape (b) and size (c) extracted from the gastrointestinal tract of  
392 fry and adults of *Cyprinus carpio*.

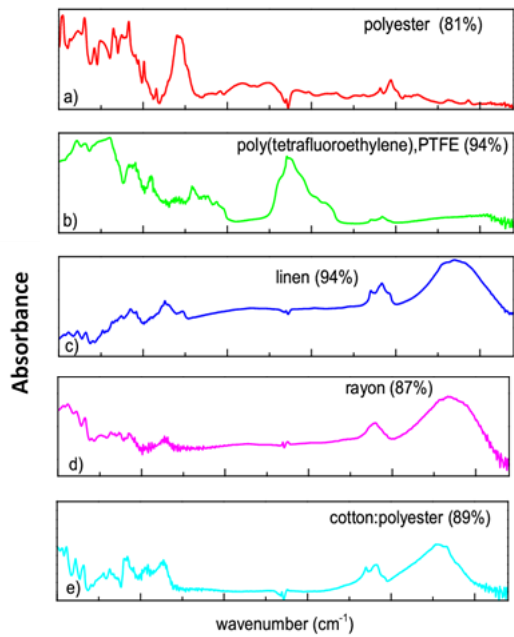


**Commentato [GD1]:** fig a) cotton:polyamide (86%)  
fig b)cotton (89%)

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395 **Figure 4.**  $\mu$ -FT-IR example spectra of the identified items in gilthead sea bream specimens: a) and b) spectra of items found in B  
sample; c) item found in C samples and d) e) and f) spectra of items found in E sample.

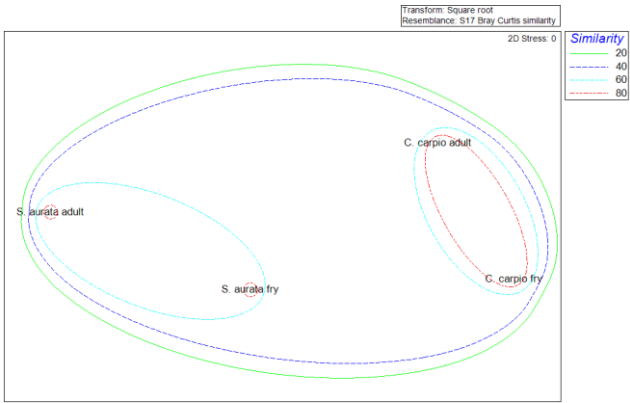
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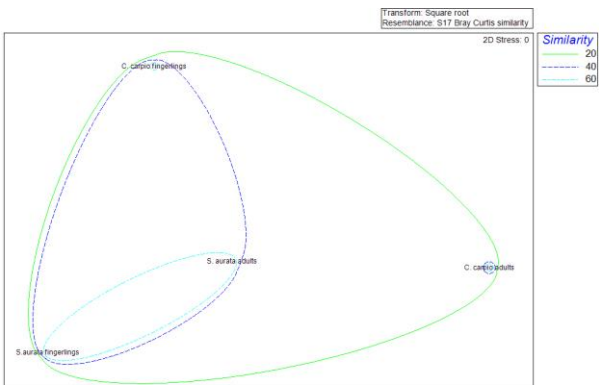


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 398 **Figure 5.** μ-FT-IR example spectra of the identified items in common carp specimens: a) b) c) spectra of items found in I sample; d)  
 399 and e) spectra of items found in sample J and K, respectively.

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 409 **Figure 6.** Similarities in microparticles size between fry and adult specimens of *Sparus aurata* and *Cyprinus carpio*.  
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 415 **Figure 7.** Similarities in microparticles polymer compositions between fry and adult specimens of *Sparus aurata* and *Cyprinus carpio*.  
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421 **REFERENCES**

- 422 Andradý, A. L. (2011). Microplastics in the marine environment. *Marine Pollution Bulletin*, 62(8), 1596–  
423 1605. <https://doi.org/10.1016/j.marpolbul.2011.05.030>
- 424 Arthur, C., Baker, J., & Bamford, H. (2008). International research workshop on the occurrence, effects, and  
425 fate of microplastic marine debris. *Conference Proceedings. Sept*, 9–11.
- 426 Barboza, L. G. A., Vethaak, A. D., Lavorante, B. R. B. O., Lundebye, A.-K., & Guilhermino, L. (2018).  
427 Marine microplastic debris: An emerging issue for food security, food safety and human health. *Marine*  
428 *Pollution Bulletin*, 133, 336–348.
- 429 Barnes, D. K. A., Galgani, F., Thompson, R. C., & Barlaz, M. (2009). Accumulation and fragmentation of  
430 plastic debris in global environments. *Philosophical Transactions of the Royal Society B: Biological*  
431 *Sciences*, 364(1526), 1985–1998.
- 432 Bellas, J., Martínez-Armental, J., Martínez-Cámara, A., Besada, V., & Martínez-Gómez, C. (2016).  
433 Ingestion of microplastics by demersal fish from the Spanish Atlantic and Mediterranean coasts.  
434 *Marine Pollution Bulletin*, 109(1), 55–60.
- 435 Bottari, T., Savoca, S., Mancuso, M., Capillo, G., GiuseppePanarello, G., MartinaBonsignore, M., Crupi, R.,  
436 Sanfilippo, M., D'Urso, L., Compagnini, G., Neri, F., Romeo, T., Luna, G. M., Spanò, N., & Fazio, E.  
437 (2019). Plastics occurrence in the gastrointestinal tract of *Zeus faber* and *Lepidopus caudatus* from  
438 the Tyrrhenian Sea. *Marine Pollution Bulletin*, 146(July), 408–416.  
439 <https://doi.org/10.1016/j.marpolbul.2019.07.003>
- 440 Browne, M. A., Galloway, T., & Thompson, R. (2007). Microplastic—an emerging contaminant of potential  
441 concern? *Integrated Environmental Assessment and Management: An International Journal*, 3(4), 559–  
442 561.
- 443 Capillo, G., Savoca, S., Panarello, G., Mancuso, M., Branca, C., Romano, V., D'Angelo, G., Bottari, T., &  
444 Spanò, N. (2020). Quali-quantitative analysis of plastics and synthetic microfibers found in demersal  
445 species from Southern Tyrrhenian Sea (Central Mediterranean). *Marine Pollution Bulletin*, 150.  
446 <https://doi.org/10.1016/j.marpolbul.2019.110596>
- 447 Feng, Z., Zhang, T., Li, Y., He, X., Wang, R., Xu, J., & Gao, G. (2019). The accumulation of microplastics  
448 in fish from an important fish farm and mariculture area, Haizhou Bay, China. *Science of The Total*  
449 *Environment*, 696, 133948.
- 450 Güven, O., Gökdağ, K., Jovanović, B., & Kıdeys, A. E. (2017). Microplastic litter composition of the  
451 Turkish territorial waters of the Mediterranean Sea, and its occurrence in the gastrointestinal tract of  
452 fish. *Environmental Pollution*, 223, 286–294. <https://doi.org/10.1016/j.envpol.2017.01.025>

- 453 Hanachi, P., Karbalaei, S., Walker, T. R., Cole, M., & Hosseini, S. V. (2019). Abundance and properties of  
454 microplastics found in commercial fish meal and cultured common carp (*Cyprinus carpio* ).  
455 *Environmental Science and Pollution Research*, 26(23), 23777–23787. <https://doi.org/10.1007/s11356->  
456 019-05637-6
- 457 Lusher, A. L., Mchugh, M., & Thompson, R. C. (2013). Occurrence of microplastics in the gastrointestinal  
458 tract of pelagic and demersal fish from the English Channel. *Marine Pollution Bulletin*, 67(1–2), 94–  
459 99.
- 460 Markic, A., Gaertner, J.-C., Gaertner-Mazouni, N., & Koelmans, A. A. (2020). Plastic ingestion by marine  
461 fish in the wild. *Critical Reviews in Environmental Science and Technology*, 50(7), 657–697.
- 462 Moore, C. J. (2008). Synthetic polymers in the marine environment: a rapidly increasing, long-term threat.  
463 *Environmental Research*, 108(2), 131–139.
- 464 Nadal, M. A., Alomar, C., & Deudero, S. (2016). High levels of microplastic ingestion by the semipelagic  
465 fish bogue *Boops boops* (L.) around the Balearic Islands. *Environmental Pollution*, 214, 517–523.
- 466 Pazos, R. S., Maiztegui, T., Colautti, D. C., Paracampo, A. H., & Gómez, N. (2017). Microplastics in gut  
467 contents of coastal freshwater fish from Río de la Plata estuary. *Marine Pollution Bulletin*, 122(1–2),  
468 85–90. <https://doi.org/10.1016/j.marpolbul.2017.06.007>
- 469 Peets, P., Leito, I., Pelt, J., & Vahur, S. (2017). Identification and classification of textile fibres using ATR-  
470 FT-IR spectroscopy with chemometric methods. *Spectrochimica Acta - Part A: Molecular and*  
471 *Biomolecular Spectroscopy*, 173, 175–181. <https://doi.org/10.1016/j.saa.2016.09.007>
- 472 Rochman, C. M., Kurobe, T., Flores, I., & Teh, S. J. (2014). Early warning signs of endocrine disruption in  
473 adult fish from the ingestion of polyethylene with and without sorbed chemical pollutants from the  
474 marine environment. *Science of the Total Environment*, 493, 656–661.
- 475 Savoca, S., Capillo, G., Mancuso, M., Bottari, T., Crupi, R., Branca, C., Romano, V., Faggio, C., D'Angelo,  
476 G., & Spanò, N. (2019). Microplastics occurrence in the Tyrrhenian waters and in the gastrointestinal  
477 tract of two congener species of seabreams. *Environmental Toxicology and Pharmacology*, 67, 35–41.  
478 <https://doi.org/10.1016/j.etap.2019.01.011>
- 479 Savoca, Serena, Bottari, T., Fazio, E., Bonsignore, M., Mancuso, M., Luna, G. M., Romeo, T., D'Urso, L.,  
480 Capillo, G., Panarello, G., Greco, S., Compagnini, G., Lanteri, G., Crupi, R., Neri, F., & Spanò, N.  
481 (2020). Plastics occurrence in juveniles of *Engraulis encrasicolus* and *Sardina pilchardus* in the  
482 Southern Tyrrhenian Sea. In *Science of the Total Environment*.  
483 <https://doi.org/10.1016/j.scitotenv.2020.137457>
- 484 **Savoca, Serena, Capillo, G., Mancuso, M., Faggio, C., Panarello, G., Crupi, R., Bonsignore, M.,**  
485 **D'Urso, L., Compagnini, G., Neri, F., Fazio, E., Romeo, T., Bottari, T., & Spanò, N. (2019). Detection**  
486 **of artificial cellulose microfibrils in *Boops boops* from the northern coasts of Sicily (Central**  
487 **Mediterranean). *Science of the Total Environment*, 691, 455–465.**  
488 <https://doi.org/10.1016/j.scitotenv.2019.07.148>Solomando et al. 2020 Long-term exposure to microplastics

489 induces oxidative stress and a pro-inflammatory response in the gut of *Sparus aurata* Linnaeus, 1758.  
490 *Environmental Pollution*

491

492 Steer, M., Cole, M., Thompson, R. C., & Lindeque, P. K. (2017). Microplastic ingestion in fish larvae in the  
493 western English Channel. *Environmental Pollution*, 226, 250–259.  
494 <https://doi.org/10.1016/j.envpol.2017.03.062>

495 Thompson, R. C., Olsen, Y., Mitchell, R. P., Davis, A., Rowland, S. J., John, A. W. G., McGonigle, D., &  
496 Russell, A. E. (2004). Lost at sea: where is all the plastic? *Science(Washington)*, 304(5672), 838.

497 Wu, F., Wang, Y., Leung, J. Y. S., Huang, W., Zeng, J., Tang, Y., Chen, J., Shi, A., Yu, X., & Xu, X. (2020).  
498 Accumulation of microplastics in typical commercial aquatic species: A case study at a productive  
499 aquaculture site in China. *Science of The Total Environment*, 708, 135432.

500 Zheng, K., Fan, Y., Zhu, Z., Chen, G., Tang, C., & Peng, X. (2019). Occurrence and Species-Specific  
501 Distribution of Plastic Debris in Wild Freshwater Fish from the Pearl River Catchment, China.  
502 *Environmental Toxicology and Chemistry*, 38(7), 1504–1513. <https://doi.org/10.1002/etc.4437>

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