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Tree invasions in Italian forests

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Abstract:	<p>Many forest tree species have been moved outside their native range to provide goods and services elsewhere, but some of them have become invasive, causing negative impacts on biodiversity and human activities. The assessment and knowledge on the degree and scale to which forest ecosystems are invaded by non-native trees is of paramount importance for tailored policies and strategies aiming to conserve forest. By reviewing and analyzing main databases and literature and applying a three-level scale of invasion, we assessed the current and potential occurrence of twenty-five invasive non-native trees across forest ecosystems of Italy. Vulnerability to invasion substantially differed across forest categories, with riparian forests being particularly vulnerable to invasion. Many non-native trees invaded a few forest categories but showed their potential to widen their impact on other forest categories. Furthermore, some non-native species showed a recent increase of their invasive status in Italy. This review identifies the Italian forest categories and non-native trees that deserve primary attention and can support management, and allocation of funding to protect forests and limit the spread of invasive tree species.</p>
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Editors-in-Chief of Forest Ecology and Management

Rosà (VI, Italy), March 18 2022

Dear Editors-in-Chief of *Forest Ecology and Management*,

This cover letter accompanies a copy of the manuscript entitled "Tree invasions in Italian forests" that was written in collaboration with Emilio Badalamenti and Tommaso La Mantia of the Università degli Studi di Palermo, Tommaso Sitzia of the Università degli Studi di Padova, Giuseppe Brundu of the Università degli Studi di Sassari, and Sabina Burrascano and Laura Celesti-Grapow of the Sapienza Università di Roma.

We propose here a national-wide assessment of the most invasive non-native trees in Italian forests. To the best of our knowledge, this is the first attempt for Italy or for any European country. We adopted a three-level scale with the aim to evaluate not only the current invasion level but also the potential invasion of Italian forest ecosystems by invasive non-native trees. As such, our study has relevant management implications allowing both increasing the efficiency of early detection and control strategies of the most invasive non-native trees in forests and prioritizing funding and silvicultural practices. The assessment approach adopted in this study can be easily updated as long as new information and knowledge about invasive species and forests are gained. We deem that a similar approach is of interest for international readers and stakeholders as it can be replicated in other countries allowing comparisons and information exchange, which are particularly useful for neighboring countries or affected by high trade exchanges. Furthermore, outcomes are of interest for those areas in which the analyzed tree species are non-native.

I trust that you will find our submission complete, thorough and to the appropriate standard for publication in *Forest Ecology and Management*. I confirm that the manuscript nor any parts of its content are currently under consideration or published in another journal. Should you have any questions regarding our submission, please do not hesitate to contact me.

The first and the last authors of this manuscript contributed equally and I would like to ask if both could be considered corresponding authors.

Yours sincerely,

Thomas Campagnaro (on behalf of our co-authors)

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1 **Tree invasions in Italian forests**

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22

23 **Highlights**

- 24 • Non-native tree species invade a wide range of forest categories in Italy
- 25 • Riparian forests are particularly vulnerable to non-native tree invasion
- 26 • *Robinia pseudoacacia* and *Ailanthus altissima* are by far the most frequent invaders
- 27 • *Eucalyptus camaldulensis*, *Populus ×canadensis* and *Ulmus pumila* have recently spread
- 28 • Tailored strategies for the management of non-native tree species are needed

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30 **Abstract**

31 Many forest tree species have been moved outside their native range to provide goods and services
32 elsewhere, but some of them have become invasive, causing negative impacts on biodiversity and
33 human activities. The assessment and knowledge on the degree and scale to which forest
34 ecosystems are invaded by non-native trees is of paramount importance for tailored policies and
35 strategies aiming to conserve forest. By analyzing main databases and literature and applying a
36 three-level scale of invasion, we assessed the current and potential occurrence of twenty-five
37 invasive non-native trees across forest ecosystems of Italy. Vulnerability to invasion substantially

38 differed across forest categories, with riparian forests being particularly vulnerable to invasion.
39 Many non-native trees invaded a few forest categories but showed their potential to widen their
40 impact on other forest categories. Furthermore, some non-native species showed a recent increase
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42 trees that deserve primary attention and can support management, and allocation of funding to
43 protect forests and limit the spread of invasive tree species.

44

45 **Key words**

46 invasive alien species; exotic trees, non-native trees; silviculture; biodiversity; global change

47

48 **Introduction**

49 Non-native tree species are increasingly becoming dominant components in forest ecosystems in
50 many parts of the world (FAO 2020), thus stimulating debates and concerns about the risks
51 associated with their generalized or unregulated use (Dickie et al. 2014, Pötzelsberger et al. 2020).
52 A decade ago, 434 invasive non-native tree species have been recorded globally (Richardson and
53 Rejmánek 2011, Rejmánek and Richardson 2013), but this figure requires constant updating
54 (Rejmánek 2014). New introductions of non-native species are increasing, and this trend is expected
55 to continue in the next future (Seebens et al. 2017, 2020). Furthermore, a higher number of non-
56 native species, among which trees, is projected to become invasive, although after a lag phase
57 (Early et al. 2016, Richardson et al. 2014). In Europe, 150 non-native tree species were found to
58 cover around 8.54 million hectares, 4% of the total forest area (Brus et al. 2019). Notwithstanding
59 these broad scale syntheses, detailed and updated information for individual countries are often
60 missing, and this hampers early warning systems as well as concrete containment actions.

61 Successful non-native tree invaders most likely high metabolic rate, fast growth rate, great height,
62 abundant seed production, early flowering, robust leaves of longer lifespan, greater duration of
63 annual carbon gain, and wide native range (Lamarque et al. 2011, Pyšek et al. 2014, van der Sande
64 et al. 2020, Fridley et al. 2022). However, other factors play a role in fostering the invasiveness of
65 non-native tree species; for instance, introduction pathways and planting history can be crucial in
66 determining the status and extent of the spread of a species in a given region (Krivánek et al. 2006,
67 Donaldson et al. 2014) or in a specific context (Catford et al. 2022, Sapsford et al. 2020). Globally,
68 non-native trees are introduced and spread mainly through forestry and agroforestry activities
69 (Richardson 1998, Richardson et al. 2014, Brundu and Richardson 2016, Sitzia et al. 2016,
70 Badalamenti et al. 2018, Brundu et al. 2020a) because of their resistance to abiotic and biotic
71 stresses – a crucial aspect to produce valuable timber (Kjær et al. 2014, Sitzia 2014, Keča et al.
72 2019) – especially in times of climate change.

73 Even though only a minor proportion of introduced tree species become invasive, they may cause
74 large, harmful, and long-lasting effects, whose impacts on biodiversity and human societies deserve
75 primary attention (Pötzelsberger et al. 2020). Non-native tree species effects on ecosystem
76 (dis)services are characterized by synergies and trade-offs among them that, in turn, are influenced
77 by context dependencies and local factors (Castro-Díez et al. 2019, 2021). Also, the duration of the
78 invasion can determine the impacts and spread by non-native trees (Mostert et al. 2017) hampering
79 the timely detection of the most harmful species.

80 Non-native trees may invade a wide range of natural and semi-natural ecosystems, encompassing
81 forests as well as open habitats. In Europe, non-native trees are increasingly reported to invade

82 forest habitats (Berg et al. 2016), with some of them being reported to exert a variety of impacts on
83 forest structure and processes, as well as on the assemblage of different taxa (Campagnaro et al.
84 2018, Gentili et al. 2019, Langmaier and Lapin 2020). European forests vulnerability to plant
85 invasions increases with the level of disturbance, fragmentation, and non-native propagule pressure
86 (Wagner et al. 2017). The habitat-based approach is commonly adopted to understand the
87 preferential patterns of invasive plants spread, thus having a strong influence on management and
88 control strategies (Chytrý et al. 2008). There is large evidence that floodplain and oak-dominated
89 woodlands, many of which are of conservation interest, are among the most invaded forest habitats
90 (Essl et al. 2011, Berg et al. 2016, Wagner et al. 2017, Campagnaro et al. 2018, Guarino et al.
91 2021). Furthermore, recent research has highlighted that information on the native habitats of non-
92 native species can provide insights into the forest types that are more likely to be invaded (van der
93 Sande et al. 2020). Despite these findings, the context-dependency of non-native trees status, spread
94 and impacts, indicates the need for studies conducted at fine spatial scales, such as regions and
95 countries (Sádlo et al. 2017), finding associations between non-native trees and invaded forest
96 types. This would allow to address control and monitoring efforts towards the most vulnerable
97 habitats (Bonari et al. 2021) and the most invasive non-native trees (Lapin et al. 2019), thus moving
98 towards tailored and adaptive approaches in non-native tree species management (van Wilgen and
99 Richardson 2014).

100 Here, we combined the national inventory of the Italian non-native flora (Celesti-Grapow et al.
101 2009) and its updates, which report data for each of the 21 administrative regions, with a literature
102 search focused on invasive non-native tree species and on their links with Italian forest categories.
103 Based on this approach, we provide a comprehensive overview on the occurrence of invasive non-
104 native tree species in Italy and an assessment of tree invasions in Italian forests. Our ultimate goal is
105 to inform the prioritization of both non-native tree species and forest categories for their
106 management at the country scale.

107

108 **Materials and methods**

109 *Selection of invasive non-native tree species*

110 We identified and selected the invasive non-native trees (INNTs) for the 21 administrative Italian
111 regions based on: the checklist of the non-native flora of Italy (Celesti-Grapow et al. 2009); its
112 updates (Galasso et al. 2018a, 2018b, 2018c, 2019a, 2019b, 2020a, 2020b, 2021a, 2021b), most of
113 which included in the *Portale della Flora d'Italia* (2021) and *Acta Plantarum* (Longo et al. 2021)
114 online sources; and additional non-indexed sources (e.g., Alessandrini et al. 2018). We included
115 trees (i.e., scapose phanerophytes), woody plants bearing the resting buds higher than 200 cm from

116 the ground and characterized by a clearly identifiable main stem (Raunkiær 1934; Ellenberg and
117 Mueller-Dombois 1967), that were neophytes, species introduced after 1492 (Pyšek et al. 2004).
118 Life forms were retrieved from Acta Plantarum (Longo et al. 2021) and the global database of
119 invasive trees and shrubs (Richardson and Rejmánek 2011) or, if necessary, upon expert opinion.
120 According to these criteria, 25 INNTs were identified for Italy (Table 1).

121 Table 1 Overview of the 25 invasive non-native tree species considered in this study.

Invasive non-native tree species	Species code	Family	Native range ^a	Year of first introduction ^b	Documented or potential negative impact on Italian forests ^c
<i>Acacia dealbata</i> Link subsp. <i>dealbata</i>	Aca.dea	Fabaceae	SE Australia	1835	Impact outcome <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - Soil or sediment modification - Modification of successional patterns - Modification of nutrient pool and fluxes - Modification of food web - Primary production alteration Mechanism <ul style="list-style-type: none"> - Competition - Interaction with other invasive species
<i>Acacia mearnsii</i> De Wild.	Aca.mea	Fabaceae	SE Australia	1821	Impact outcome <ul style="list-style-type: none"> - Reduction in native biodiversity - Soil or sediment modification (GB pers. obs.) - Modification of nutrient pool and fluxes (GB pers. obs.) Mechanism <ul style="list-style-type: none"> - Competition - Interaction with other invasive species^f
<i>Acacia provincialis</i> A.Camus	Aca.pro	Fabaceae	South Australia to Victoria	1862	Not documented up to now

<i>Acacia pycnantha</i> Benth.	Aca.pyc	Fabaceae	SE Australia	1865	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Primary production alteration - Modification of successional patterns - Soil or sediment modification <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species
<i>Acacia saligna</i> (Labill.) H.L.Wendl.	Aca.sal	Fabaceae	SW Australia	1827	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - Modification of successional patterns - Soil or sediment modification - Habitat or refugia replacement/loss <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species^c
<i>Acer negundo</i> L.	Ace.neg	Sapindaceae	Canada to Honduras	1780	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - Modification of food web - Habitat or refugia replacement/loss <p>Mechanism</p> <ul style="list-style-type: none"> - Competition
<i>Ailanthus altissima</i> (Mill.) Swingle	Ail.alt	Simaroubaceae	China	1760	<p>Impact outcome</p> <ul style="list-style-type: none"> - Habitat or refugia replacement/loss - Habitat degradation - Reduction in native biodiversity

					<ul style="list-style-type: none"> - Modification of successional patterns - Soil or sediment modification - Physical disturbance <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species - Poisoning/toxicity - Rooting/digging
<i>Broussonetia papyrifera</i> (L.) Vent.	Bro.pap	Moraceae	Cambodia, China, India, Japan, Laos, Myanmar, Thailand	1760	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Poisoning/toxicity
<i>Elaeagnus pungens</i> Thunb.	Ela.pun	Elaeagnaceae	South-Central and Southeast China, Japan, Korea, Tadzhikistan	1865	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity <p>Mechanism</p> <ul style="list-style-type: none"> - Competition
<i>Eucalyptus camaldulensis</i> Dehnh. subsp. <i>camaldulensis</i>	Euc.cam	Myrtaceae	Australia	1832 ^c	<p>Impact outcome³</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - Soil or sediment modification (GB pers. obs.) <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species

<i>Juglans nigra</i> L.	Jug.nig	Juglandaceae	Canada, USA	1760	Impact outcome ^g - Habitat degradation Mechanism - Competition
<i>Ligustrum lucidum</i> W.T.Aiton	Lig.luc	Oleaceae	Central & S China to Hainan, S Korea	1831	Impact outcome - Habitat degradation - Modification of nutrient pool and fluxes - Reduction in native biodiversity Mechanism - Competition
<i>Parkinsonia aculeata</i> L.	Par.acu	Fabaceae	S USA to N Argentina, Caribbean	1760	- Not documented up to now
<i>Paulownia tomentosa</i> (Thunb.) Steud.	Pau.tom	Paulowniaceae	Central & E China, S Korea	1843	Impact outcome - Reduction in native biodiversity Mechanism - Competition
<i>Platanus ×hispanica</i> Mill. ex Münchh.	Pla.his	Platanaceae	Hybrid origin	After 1551	Impact outcome ^h - Habitat degradation Mechanism - Competition
<i>Populus ×canadensis</i> Moench	Pop.x.can	Salicaceae	Hybrid origin	1907	Impact outcome ^j - Reduction in native biodiversity

					<ul style="list-style-type: none"> - Habitat degradation <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Hybridization¹
<i>Prunus laurocerasus</i> L.	Pru.lau	Rosaceae	SE Europe to Iran	1563	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - (Allelopathy) <p>Mechanism</p> <ul style="list-style-type: none"> - Competition
<i>Prunus serotina</i> Ehrh.	Pru.ser	Rosaceae	W Canada to NW USA, E Canada to Guatemala	1811	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - (Allelopathy) <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Poisoning/toxicity
<i>Quercus rubra</i> L.	Que.rub	Fagaceae	Canada, USA	1803	<p>Impact outcome</p> <ul style="list-style-type: none"> - Reduction in native biodiversity - Habitat degradation - Habitat or refugia replacement/loss <p>Mechanism</p> <ul style="list-style-type: none"> - Competition

<i>Robinia pseudoacacia</i> L.	Rob.pse	Fabaceae	USA	1662	<p>Impact outcome</p> <ul style="list-style-type: none"> - Habitat degradation - Habitat or refugia replacement/loss - Modification of food web - Modification of nutrient pool and fluxes - Modification of successional patterns - Primary production alteration - Reduction in native biodiversity - Soil or sediment modification <p>Mechanism</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species - Poisoning/toxicity
<i>Robinia viscosa</i> Vent.	Rob.vis	Fabaceae	E USA	1802	Not documented up to now
<i>Sesbania punicea</i> (Cav.) Benth.	Ses.pun	Fabaceae	Central & S. Brazil to NE. Argentina	1842	<p>Impact outcome (EB pers. obs.)</p> <ul style="list-style-type: none"> - Habitat degradation - Reduction in native biodiversity - Modification of nutrient pool and fluxes <p>Mechanism (EB pers. obs.)</p> <ul style="list-style-type: none"> - Competition - Interaction with other invasive species^c
<i>Trachycarpus fortunei</i> (Hook.) H.Wendl.	Tra.for	Areaceae	S Central China to N Myanmar	1834	<p>Impact outcome</p> <ul style="list-style-type: none"> - Habitat degradation - Reduction in native biodiversity <p>Mechanism</p> <ul style="list-style-type: none"> - Competition

<i>Ulmus pumila</i> L.	Ulm.pum	Ulmaceae	Central Asia to S Siberia and Korea	1811	Impact outcome - Reduction in native biodiversity Mechanism - Competition - Hybridization
<i>Vachellia karroo</i> (Hayne) Banfi & Galasso	Vac.kar	Fabaceae	Southern Africa	1821 ^d	Impact outcome - Reduction in native biodiversity Mechanism - Competition

122

^a POWO (2021).

123

^b Maniero (2000), Maniero (2015) and references therein, unless otherwise stated.

124

^c Badalamenti et al. (2018).

125

^d Pasta et al. (2012).

126

^e Lazzaro et al. (2020) and references therein, unless otherwise stated. The terminology on impact mechanisms and outcomes follows Blackburn et al. (2014).

127

^f Brundu et al. (2019)

128

^g Scatamacchia (2011).

129

^h Mason (2004).

130

ⁱ Cicinelli et al. (2018).

131

^j Paffetti et al. (2018).

132 *Assessment of the invasive status*

133 Firstly, we assessed the status of each INNT in Italian regions over time by integrating the
134 information in the checklist of non-native flora of Italy and its updates (see above for references).
135 These lists, compiled by national and local experts, apply a classification system of alien plant
136 species tied to the introduction-naturalisation-invasion continuum approach (Pyšek et al. 2004).
137 Therefore, the current definition of invasive status is predominantly based on such approach and, to
138 date, a comprehensive risk assessment of the non-native flora of Italy is missing. Hence, invasive
139 plant species are defined as those “*that produce reproductive offspring, often in very large numbers,*
140 *at considerable distances from the parent plants, and thus have the potential to spread over a large*
141 *area*” (Pyšek et al. 2004). This definition is neither based on the species actual or potential impacts,
142 nor on prioritisation or risk assessment methods. To overcome this limitation and take into account
143 also impact-based definitions (e.g., European Parliament and Council adopted Regulation (EU) No
144 1143/2014) that imply being ecologically detrimental (Lemoine and Svenning 2022), we applied an
145 integrated approach considering the status in the national list as well as the documented and
146 potential impacts for the Italian territory (see Table 1). Furthermore, when referring to naturalized
147 species, we intend species “*that sustain self-replacing populations for at least 10 years without the*
148 *direct intervention of people [...] by the recruitment of seeds or ramets [...] capable of independent*
149 *growth*” (Pyšek et al. 2004).

150 Secondly, we integrated this assessment with recent studies reporting non-native species
151 management at the country scale to account for species inclusion in regional legislations (Brundu et
152 al. 2020b), and priority species that have been suggested as suitable candidates to be included in a
153 national list according to Regulation (EU) n. 1143/2014 (Lazzaro et al. 2019). Our assessment
154 considered non-native tree species reported spreading outside planting sites in Europe (Brus et al.
155 2019), the checklists of woody invasive non-native species in the world (Richardson and Rejmánek
156 2011, Rejmánek and Richardson 2013) and the global risk score (Randall 2017).

157

158 *Classification of the forest categories*

159 A variety of forest classifications exist and are commonly adopted in Italy, encompassing different
160 spatial scales and objectives: from the European classification used for assessing indicators of
161 sustainable forest management (Barbati et al. 2014), to the national classification system used in the
162 inventory of forests and forest carbon pools (Gasparini and Tabacchi 2011), up to regional
163 classifications (Del Favero 2004, Del Favero 2008, Del Favero 2010). However, research on non-
164 native tree species has rarely referred to encoded forest types or categories, but rather to the
165 dominant tree species. Thus, for the purpose of the present study, we adopted an original

166 classification of forest categories (Table 2). We also included a focus on spontaneous urban forests,
 167 which are particularly susceptible to the spread of non-native tree species, due to high propagule
 168 pressure from ornamental non-native trees and plantations, and the high level of disturbance.
 169 The spatial extent of each forest category was based on Gasparini and Tabacchi (2011), except for
 170 mountain pine forests, for which we considered the area reported in last national reporting
 171 according to art. 17 of the Habitats Directive. We used the quartile distribution of the categories
 172 extent data to define three broad classes: narrow, medium, large.

173

174 Table 2. Forest categories used to assess the invasion of non-native trees in Italy, their dominant
 175 woody species and class of spatial extent.

Forest category	Dominant woody species	Spatial extent
Subalpine and temperate scrublands	<i>Alnus alnobetula</i> , <i>Calluna vulgaris</i> , <i>Rhododendron hirsutum</i> , <i>Salix</i> sp.pl., <i>Vaccinium myrtillus</i>	Medium
Maquis, Mediterranean scrublands	<i>Arbutus unedo</i> , <i>Erica</i> sp.pl., <i>Myrtus communis</i> , <i>Pistacia lentiscus</i> , <i>Phillyrea</i> sp.pl., <i>Prunus spinosa</i> subsp. <i>spinosa</i> , <i>Rhamnus alaternus</i>	Large
Mediterranean pine forests	<i>Pinus halepensis</i> subsp. <i>halepensis</i> , <i>Pinus pinaster</i> subsp. <i>pinaster</i> , <i>Pinus pinea</i>	Medium
Oak-hornbeam and hornbeam forests	<i>Carpinus betulus</i> , <i>Quercus robur</i>	Narrow
Holm oak forests	<i>Quercus ilex</i>	Medium
Cork oak forests	<i>Quercus suber</i>	Medium
Downy oak forests	<i>Quercus pubescens</i> s.l.	Large
Sessile oak forests	<i>Quercus petraea</i>	Narrow
Pedunculate oak forests	<i>Quercus robur</i>	Narrow
Turkey oak forests	<i>Quercus cerris</i>	Large
Other oak forests	<i>Quercus frainetto</i> , <i>Quercus trojana</i>	Narrow
Chestnut forests	<i>Castanea sativa</i>	Large
Manna ash-hornbeam forests	<i>Fraxinus ornus</i> , <i>Ostrya carpinifolia</i> , (<i>Quercus pubescens</i>)	Large
Maple-European ash and Maple-Lime forests	<i>Acer pseudoplatanus</i> , <i>Fraxinus excelsior</i> , <i>Tilia cordata</i> , <i>Tilia platyphyllos</i>	Medium
Pioneer forests	<i>Acer campestre</i> , <i>Betula pendula</i> , <i>Corylus avellana</i> , <i>Populus tremula</i> , <i>Ulmus minor</i> s.l.	Narrow
Beech forests	<i>Fagus sylvatica</i>	Large
Mountain pine forests	<i>Pinus mugo</i>	Medium

Black pine forests	<i>Pinus nigra</i> s.l.	Medium
Scots pine forests	<i>Pinus sylvestris</i>	Medium
Silver fir forests	<i>Abies alba</i>	Narrow
Norway spruce forests	<i>Picea abies</i>	Medium
Larch, Larch-Swiss pine, and Swiss pine forests	<i>Larix decidua</i> , <i>Pinus cembra</i>	Medium
Riparian forests	<i>Alnus glutinosa</i> , <i>Alnus incana</i> , <i>Fraxinus angustifolia</i> , <i>Platanus orientalis</i> , <i>Populus</i> sp.pl., <i>Salix</i> sp.pl.	Medium
Forests dominated by other conifers (including plantations)	<i>Cedrus</i> sp.pl., <i>Cupressus</i> sp.pl., <i>Pinus heldreichii</i> subsp. <i>leucodermis</i> , <i>Taxus baccata</i>	Medium
Forests dominated by other broadleaves	<i>Alnus cordata</i> , <i>Ceratonia siliqua</i> , <i>Ilex aquifolium</i> , <i>Laurus nobilis</i> , <i>Olea europaea</i> var. <i>sylvestris</i>	Medium

176

177 *Assessment of the current and potential invasion level of Italian forests by invasive non-native trees*
178 The current and potential invasion level of forest categories by INNTs was ranked from 0 (category
179 with no evidence of current or potential invasion) to 3 (category massively invaded). Intermediate
180 values have been attributed when an INNT is not widely invading the category in Italy (score = 2)
181 or it has solely been observed invading the category outside Italy (score = 1). The difference
182 between score 3 and score 2 accounted for the level and the temporal dynamics of the invasion
183 process in different Italian regions. The score 1 was based on the *invasive elsewhere* criterion, i.e.,
184 the documented establishment of a given non-native tree species in one category outside Italy was
185 considered as a reliable proxy of the potential invasion of the same category in Italy.

186

187 Table 3 - Scale for the level of invasion of forest categories by non-native trees applied in this
188 assessment for Italy.

Score	Brief description	Full description
0	Not currently invaded and with low invasibility	Forest category not invaded in Italy, Europe or elsewhere and with no or very low potential to be invaded in the future
1	Potentially invasible	Forest category invaded outside Italy and prone to be invaded in Italy
2	Moderately invaded	Forest category currently invaded in a limited portion of

		the national range of the forest category
3	Massively invaded	Forest category commonly invaded over large portions of the national range of the forest category

189

190 The score was associated with each combination of category and species based on bibliographic
191 sources and expert knowledge acquired through on-purpose meetings (Supplementary material 1).
192 Primarily, we considered scientific and grey literature reporting information on the spread of INNTs
193 in specific categories in Italy: (i) manuals or atlases, (ii) large scale studies about introduced tree
194 species including Italy in their study area; (iii) studies addressing the invasion by individual INNTs,
195 and (iv) technical documents. Secondly, especially for non-native trees with a current limited
196 range in Italy, the scientific literature mostly derived from an *ad-hoc* search on Google scholar of
197 their scientific and common names, including the most used synonyms, in association with the
198 forest categories of this study. Only scientific articles explicitly referring to spontaneous (*i.e.*, not
199 planted) individuals of INNTs were considered.

200 We also evaluated INNTs found in urban forests and whether they were the stand-dominant species
201 or not, based on Italian literature and personal observations.

202

203 *Statistical analysis of the invasion level scores*

204 We used non-metric multidimensional scaling (NMDS) based on the Cao index of dissimilarity in
205 the level of invasion of each category (function *metaMDS* in the R package *vegan*; R Development
206 Core Team 2021). We correlated (Spearman rho coefficient) the number of INNTs with each score
207 different from 0 per category with the area of that category. We also correlated the number of
208 categories per score to which an INNT was combined with: (i) the date of first introduction; (ii) the
209 number of regions where the tree species was reported as invasive, (iii) or as with self-replacing
210 populations in 2021 (Table 4).

211

212 **Results**

213 *Invasive non-native tree species in Italy*

214 Our systematic overview pointed out an increase in the number of INNTs, from 13 to 25, in 13
215 years (Table 4, Supplementary material 2). Fabaceae family was by far the most represented (10
216 taxa, 40% of all tree species); in fact, except from Rosaceae (represented by 2 taxa), all the other
217 families were represented by a single species. Origin was almost equally distributed among

218 temperate Asia (32%), Australia and North America (24%). The only two species with tropical and
219 subtropical origin were *Parkinsonia aculeata* and *Sesbania punicea*. The majority of tree species
220 was introduced to Italy during the nineteenth century (> 60%), and 20% of the species during the
221 eighteenth century. Older introductions (before 1700) accounted for slightly more than 10% of the
222 species, while only one species was introduced in the last century.

223 *Ailanthus altissima* and *Robinia pseudoacacia* were invasive in most of the national territory (more
224 than 17 regions), while most species (about 40%) were invasive in a single Italian region. When
225 considering naturalized and invasive status together, six species (*Acer negundo*, *Ailanthus altissima*,
226 *Broussonetia papyrifera*, *Ligustrum lucidum*, *Platanus hispanica* and *Robinia pseudoacacia*)
227 occurred in most regions, while only two species were limited to a single region (*Acacia mearnsii*
228 and *Robinia viscosa*). The species which registered the highest increase in the number of regions
229 with self-replacing populations were *Ulmus pumila* (+6 regions), *Eucalyptus camaldulensis* and
230 *Populus ×canadensis* (+4 regions for both). About half of the species were found at least as
231 naturalized in two or three additional regions compared to the 2009 situation.

232 Importantly, about half of the species undergo some type of restrictions or are listed according to
233 regional legislation (56%; Table 4), while less were listed among the national priorities with few
234 overlaps between these two categories. Five priority species were still not legally considered
235 (*Acacia mearnsii*, *Acacia pycnantha*, *Elaeagnus pungens*, *Parkinsonia aculeata* and *Vachellia*
236 *karroo*). *Acacia mearnsii* and *Acacia pycnantha* were recorded as invasive in few regions only
237 recently.

238 Except for *Robinia viscosa* (that was not assessed), all species have been reported as invasive
239 globally, mostly with high or extreme risk scores (64%), with only two species (*Elaeagnus pungens*
240 and *Platanus hispanica*) reported with a low risk.

241

242 **Table 4.** Status of the non-native tree species considered in this study with additional information about legislative and management issues as well
 243 as their global relevance. §: self-replacing populations include “naturalized” and “invasive” status according to Pyšek et al. (2004). In round
 244 brackets, the number of regions where each tree species is regarded as invasive in Italy is reported.

Species	Family	Number of Italian regions with self-replacing populations§ (2009) ^a	Number of Italian regions with self-replacing populations§ (2018) ^b	Number of Italian regions with self-replacing populations§ (2021) ^c	Regulated by regional legislation ^d	Candidate non-native plants for national list ^e	European Countries reporting spread outside planting sites ^f	Invasive non-native species of Union concern ^g	Global invasive species ^h	Global Risk Score (rating) ⁱ
<i>Acacia dealbata</i> subsp. <i>dealbata</i>	Fabaceae	3 (2)	4 (2)	5 (2)	-	-	1	-	X	25.9 (high)
<i>Acacia mearnsii</i>	Fabaceae	1 (1)	1 (1)	1 (1)	-	X	n.r.	-	X	25.9 (high)
<i>Acacia provincialis</i>	Fabaceae	1 (1)	2 (1)	2 (1)	-	-	0	-	X	7.9 (medium)
<i>Acacia pycnantha</i>	Fabaceae	0 (0)	2 (1)	2 (1)	-	X	0	-	X	9.6 (medium)
<i>Acacia saligna</i>	Fabaceae	7 (3)	8 (5)	8 (5)	X	X	1	X	X	34.6 (extreme)
<i>Acer negundo</i>	Sapindaceae	10 (4)	12 (6)	12 (6)	X	-	8	-	X	34.6 (extreme)
<i>Ailanthus altissima</i>	Simaroubaceae	20 (20)	20 (20)	20 (20)	X	-	10	X	X	43.2 (extreme)

<i>Broussonetia papyrifera</i>	Moraceae	13 (3)	14 (4)	14 (4)	X	-	1	-	X	25.9 (high)
<i>Elaeagnus pungens</i>	Elaeagnaceae	2 (0)	4 (1)	4 (1)	X	X	n.r.	-	X	3.6 (low)
<i>Eucalyptus camaldulensis</i> subsp. <i>camaldulensis</i>	Myrtaceae	2 (0)	6 (1)	6 (1)	X	-	0	-	X	25.9 (high)
<i>Juglans nigra</i>	Juglandaceae	0 (0)	2 (1)	3 (1)	-	-	0	-	X	12.9 (medium)
<i>Ligustrum lucidum</i>	Oleaceae	9 (0)	12 (2)	12 (2)	X	-	0	-	X	43.2 (extreme)
<i>Parkinsonia aculeata</i>	Fabaceae	1 (1)	2 (1)	3 (2)	-	X	n.r.	-	X	64 (extreme)
<i>Paulownia tomentosa</i>	Paulowniaceae	5 (0)	7 (1)	7 (1)	X	-	0	-	X	34.6 (extreme)
<i>Platanus hispanica</i>	Platanaceae	8 (0)	10 (0)	11 (1)	-	-	0	-	X	3.36 (low)
<i>Populus ×canadensis</i>	Salicaceae	5 (0)	8 (0)	9 (2)	-	-	2	-	X	16.8 (high)
<i>Prunus laurocerasus</i>	Rosaceae	2 (1)	4 (3)	5 (3)	X	-	n.r.	-	X	19.2 (high)
<i>Prunus serotina</i>	Rosaceae	3 (3)	4 (3)	4 (3)	X	X	9	-	X	43.2

										(extreme)
<i>Quercus rubra</i>	Fagaceae	3 (2)	4 (2)	4 (2)	X	X	3	-	X	25.9 (high)
<i>Robinia pseudoacacia</i>	Fabaceae	20 (17)	20 (17)	20 (17)	X	-	17	-	X	64.0 (extreme)
<i>Robinia viscosa</i>	Fabaceae	0 (0)	1 (1)	1 (1)	-	X	0	-	-	Not assessed
<i>Sesbania punicea</i>	Fabaceae	0 (0)	2 (0)	3 (2)	-	-	-	-	X	8.64 (medium)
<i>Trachycarpus fortunei</i>	Arecaceae	3 (2)	5 (3)	5 (3)	X	X	0	-	X	17.3 (high)
<i>Ulmus pumila</i>	Ulmaceae	2 (0)	7 (2)	9 (2)	X	-	1	-	X	25.9 (high)
<i>Vachellia karroo</i>	Fabaceae	4 (0)	5 (0)	5 (1)	-	X	0	-	X	38.4 (extreme)

245

^a Celesti-Gradow et al. (2009)

246

^b Galasso et al. (2018a)

247

^c Longo et al. (2021), Portale della Flora d'Italia (2021), Galasso et al. (2021a), and Galasso et al. (2021b)

248

^d Brundu et al. (2020b)

249

^e Lazzaro et al. (2019)

250

^f Brus et al. (2019); n.r.: not reported

251

^g pursuant to Regulation (EU) n. 1143/2014, and included in the Commission Implementing Regulation (EU) 2016/1141, 2017/1263 or 2019/1261

252

^h Richardson and Rejmánek (2011) and Rejmánek and Richardson (2013)

253

ⁱ Randall (2017)

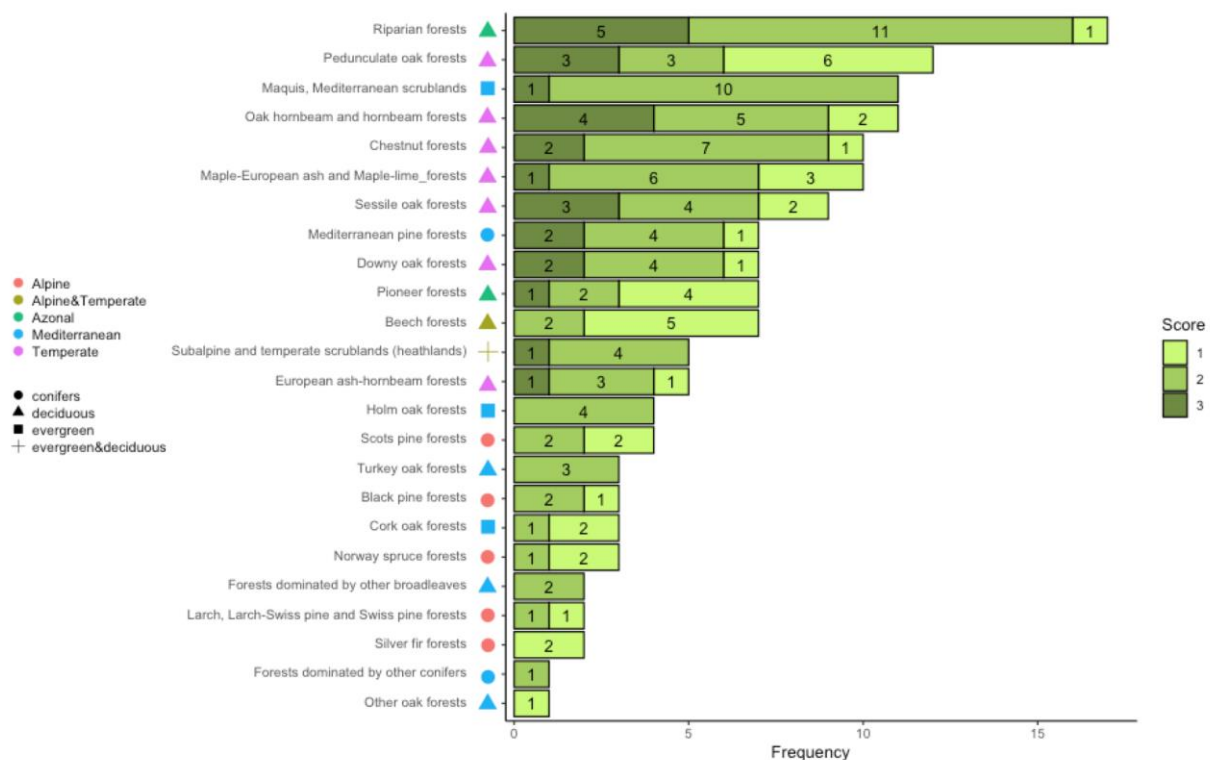
254 *Invasion of forest categories by non-native trees in Italy*

255 Forest categories largely differed in the number of associated INNTs. Riparian forests were by
 256 far the most invaded forest category with 16 INNTs, accounting for more than 60% of all INNTs
 257 (Figure 1). Pedunculate oak forests, oak-hornbeam and hornbeam forests, maple-European ash
 258 and maple lime forests, Mediterranean maquis and chestnut forests hosted more than 10 INNTs.
 259 Only three categories were invaded by only one species (other oak forests, forests dominated by
 260 other conifers) or by none (mountain pine forests).

261 The categories with the highest number of potentially invading species (score 1) were:
 262 pedunculate oak forests (n = 6), beech forests (n = 5), pioneer forests (n = 4), maple-European
 263 ash and maple-lime forests (n = 3). For two categories (other oaks forests and silver fir forests),
 264 we only found one potentially invading tree species each.

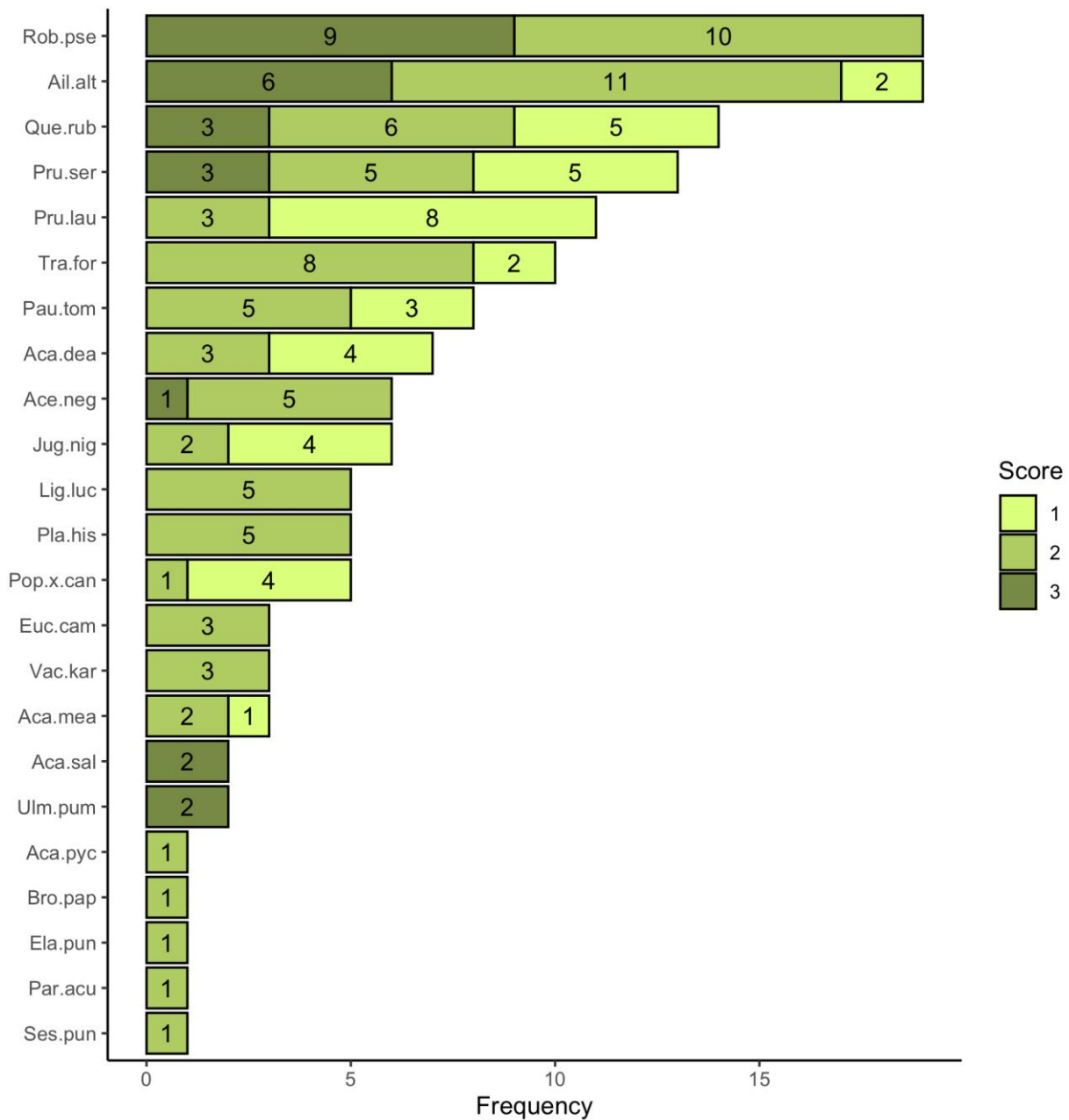
265

266 Figure 1: Number of non-native tree species invading or potentially invading each forest
 267 category in Italy. The scores associated with each group of species are indicated by different
 268 shades of green.



269

270 Figure 2: Number of forest categories invaded or potentially invaded by non-native tree species
 271 (see codes reported in Table 1) in Italy. The scores associated with each group of species are
 272 indicated by different shades of green.



273

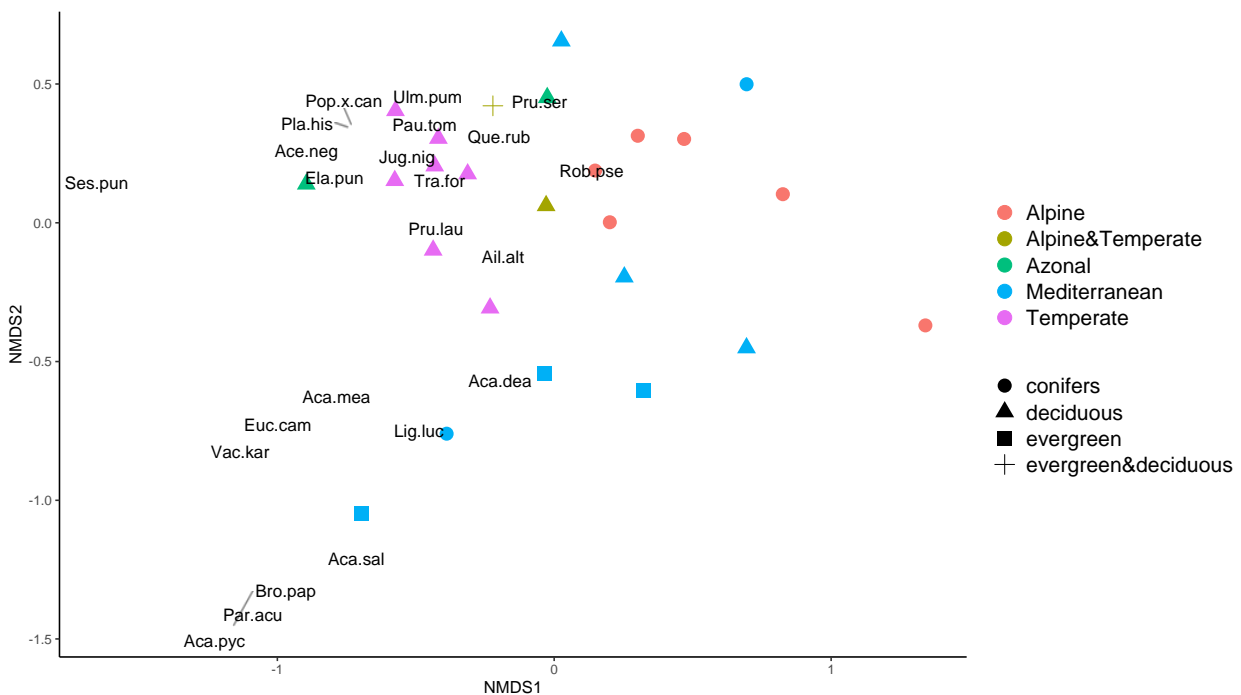
274

275 The non-native tree species invading the highest number of categories were *Robinia*
 276 *pseudoacacia* and *Ailanthus altissima* (n=19 and n=17, respectively), affecting about three
 277 quarters of all forest categories (Figure 2). The other INNTs listed by decreasing number of
 278 invaded categories were *Quercus rubra* (n=9), *Prunus serotina* (n=8), *Trachycarpus fortunei*
 279 (n=8), *Acer negundo* (n=6), and *Ligustrum lucidum*, *Paulownia tomentosa*, and *Platanus*
 280 *hispanica* (n=5 each). Fourteen INNTs are currently invading 1 to 3 forest categories. Finally,
 281 two species did not occur in Italian forests (*Acacia provincialis* and *Robinia viscosa*).

282 Some species have mainly the potential to invade rather than being currently invading Italian
 283 forests. For instance, *Prunus laurocerasus* invaded eight forest categories outside Italy, thus
 284 representing a potential threat for the same forest categories in Italy. The other tree species for

285 which we found a higher number of potentially invaded categories were *Prunus serotina* and
 286 *Quercus rubra* (n=5 for both). For *Acacia dealbata*, *Juglans nigra* and *Populus ×canadensis*,
 287 which were invasive only in a few forest categories, we found a relatively high number of forest
 288 categories potentially invasible (n=4). Importantly, we also found evidence of potential invasion
 289 even by widespread invasive species such as *Ailanthus altissima*. Urban forests were commonly
 290 composed of INNTs. In fact, we found that 20 out of the 25 species (80%) can be found in forest
 291 stands in urban settings, where seven of them were found to be able to dominate the tree layer.
 292 The NMDS ordination highlighted a group of non-native species mostly invading Mediterranean
 293 evergreen broadleaf and conifer categories (*Acacia dealbata*, *Acacia saligna* and *Ligustrum*
 294 *lucidum*) while a more conspicuous group of non-native trees were linked to temperate
 295 deciduous forests (Figure 3). The majority of the forest categories occurred in Alpine areas, as
 296 well as Mediterranean deciduous were scarcely associated with INNTs.

297
 298 Figure 3: NMDS ordination of forest categories (symbols), classified based on main
 299 biogeographical occurrence (Alpine, Alpine and temperate, azonal, Mediterranean, temperate
 300 indicated by different colors) and main tree species character (conifers, deciduous, evergreen,
 301 evergreen and deciduous indicated by different symbols), and non-native trees (see codes
 302 reported in Table 1) invasion level.



303
 304

305 The number of categories with a score 3 for a specific INNT was positively correlated with the
 306 number of regions where the same species was indicated as invasive. Whereas, the number of
 307 categories with a score 2 and a score 3 for a specific INNT was positively correlated with the
 308 number of regions where the same species was reported as naturalised or invasive. The number
 309 of INNTs with score 1 per category was negatively correlated to the spatial extent of the
 310 category (Table 5), indicating that forest categories with a narrow range are also those subject to
 311 higher potential invasion risk.

312

313 Table 5: Spearman correlation (rho) values ($p \leq 0.05$ as *; $p \leq 0.01$ as **; $p \leq 0.001$ as ***)
 314 between the number of invasive non-native trees and forest categories by invasion scores, and
 315 introduction time, occurrence and status in administrative regions, and spatial extent
 316 respectively.

	Ys. from 1 st introduction	No. of regions as invasive	No. of regions as self-replacing populations		Category's spatial extent
Score 1	0.050	0.152	-0.099	Score 1	-0.402*
Score 2	0.324	0.352	0.452*	Score 2	0.263
Score 3	0.355	0.651***	0.416*	Score 3	-0.113

317

318 Discussion

319 We found increasing trends in the number of invasive non-native tree species and invaded
 320 regions by INNTs in Italy. Given these trends, understanding the invasion patterns of non-native
 321 tree species in different forest categories at the country scale is crucial to plan and adopt
 322 adequate management strategies to prevent and/or mitigate their detrimental effects on
 323 biodiversity and ecosystem services. Here, we identified the INNTs already established within a
 324 high number of forest categories in different Italian regions, along with those which registered
 325 the highest increase in the number of affected regions, thus suggesting an increasing trend in
 326 their spread in Italy. By combining information on invasive non-native tree species and invaded
 327 forest categories in Italy, we highlighted specific associations between invasive non-native
 328 species and forest categories that may help identify species and forest categories with a higher
 329 need for prevention and control measures to tackle invasive spread (Catford et al. 2012).
 330 Furthermore, here we give an insight of the potential impacts that the main non-native tree
 331 species have on forest ecosystems.

332

333 *Most invaded forests*

334 We found evidence of a large variability in the vulnerability of different forest categories to the
335 invasion by non-native tree species in Italy. This was partly known (see focus on habitat types of
336 Community interest by Angelini et al. (2016) and Lazzaro et al. (2020)), but still not
337 comprehensively assessed. In general, forest categories occurring in plain and hilly areas hosted
338 a higher number of non-native trees than montane forests.

339 Deciduous, broadleaved-dominated forests were the most invaded. Maquis and Mediterranean
340 scrublands, and riparian, chestnut, and oak-hornbeam and hornbeam forests registered the
341 highest number of non-native species. Most of these forest categories have been already
342 highlighted as those most invaded in Europe (Campagnaro et al. 2018, Wagner et al. 2017) and
343 in other parts of the globe (Yelenik et al. 2004). For instance, the particularly worrying condition
344 of riparian forests, that have by far the highest number of non-native tree species in this study,
345 represents a widely observed pattern at the global and European scale (Langmaier and Lapin
346 2020, Railoun et al 2021). Indeed, several factors make riparian ecosystems particularly
347 vulnerable to the spread of INNTs such as the peculiar disturbance regime, high light
348 availability, frequent nutrient input, and the proximity to water that acts as a vector of transport
349 for plant propagules (Hood and Naiman 2000, Chytrý et al. 2019). Conversely, other highly
350 invaded forest categories in this study have been less studied and reported in the literature. For
351 example, maquis has been seldom reported among those highly invasible (Guarino et al. 2021)
352 and Mediterranean forests were recently reported to host fewer non-native species than
353 temperate forests (Wagner et al. 2021). Indeed, the biotic resistance of Mediterranean woodlands
354 to invasive plant species has been often invoked (FAO and Plan Bleu 2018). The contrasting
355 evidence reported in our study points to the need for cautious assessments of the invasion risk of
356 natural ecosystems.

357 Many forest categories were found to have the potential of hosting some of the non-native tree
358 species we analysed, pointing to an invasion threat for the future especially in light of the
359 forecasted climate and land use changes. Pedunculate oak and beech forests may be the most
360 affected and have been recently highlighted among hotspots of non-native species in Europe
361 (Wagner et al. 2021). While pedunculate oak forests cover a small area mainly in plain areas and
362 low hills, beech forests are among the most widespread forests in Italian mountain areas
363 (Gasparini and Tabacchi 2011). On the one hand, the narrow cover and potential changes in
364 disturbance regimes are very worrying for pedunculate oak forests; on the other hand, changes in
365 climatic conditions could trigger stronger invasions in beech forests (Kleinbauer et al. 2010).

366

367 *Most common invasive alien trees in Italian forests*

368 As expected, the species most frequently recorded in Italian forest categories were *Ailanthus*
369 *altissima* and *Robinia pseudoacacia*, which are responsible for different documented impacts on
370 biodiversity and ecosystem services in Italy (e.g., Benesperi et al. 2012, Montecchiari et al.
371 2020, Celesti-Grapow and Ricotta 2021). Both species, in addition to being the most widespread
372 over the national territory, received policy and legal attention, with *Ailanthus altissima* being
373 listed as of Union concern. Indeed, they dominate in large parts of Italian forested landscapes,
374 even within protected areas (e.g., Euganean hills), and show little or no potential to invade other
375 forest categories, suggesting that are occupying most of their current non-native range in Italy.
376 Differently, many species still not widespread in Italian forests were found to have the potential
377 to spread into different forest categories. *Prunus laurocerasus* was the species with the highest
378 number of potentially invasible forest categories, but also other species with different ecological
379 characteristics, such as *Acacia dealbata*, *Juglans nigra*, *Quercus rubra*, *Populus ×canadensis*,
380 and *Prunus serotina* showed the potential to invade several other forest categories. Overall,
381 except for some species that, up to now, mainly invaded open habitats (*Acacia provincialis*,
382 *Robinia viscosa*) or have already invaded most of the forest categories (*Robinia pseudoacacia*),
383 most non-native tree species could expand on different forest categories.
384 Climate change may drive further invasion by many non-native trees, as already highlighted for
385 *Acer negundo*, *Quercus rubra* and *Robinia pseudoacacia* in Europe (Kleinbauer et al. 2010,
386 Dyderski et al. 2018). The forest categories most likely to be at risk in Italy would be those
387 reported to be invaded in other European countries or those where a further expansion may be
388 possible over the entire national territory. However, under a progressively warmer climate, non-
389 native species are likely to increase their range where the limiting factor is represented by low
390 temperatures rather than in areas where plants are limited by high temperatures. Therefore, while
391 certain species may expand in Mediterranean forests, others, such as *Robinia pseudoacacia*
392 (Puchalka et al. 2021) or *Acer negundo*, may face a decline in potential niches within these forest
393 categories. In fact, several thermophilous tree species could expand in currently mild
394 environments (*Acacia dealbata* and *Prunus laurocerasus*), also favored by the reduction in frost
395 risk, while others may be favored by the increase in temperature in more mesic conditions and in
396 mountain areas (*Prunus serotina* and *Quercus rubra*), as already suggested by different studies
397 investigating relationships between non-native plants invasion and climate (Celesti-Grapow et al.
398 2016, Dainese et al. 2017). We also found evidence of very recent naturalization processes by
399 INNTs both in temperate and Mediterranean ecosystems. This is the case for *Ulmus pumila*
400 (Prosser and Bertolli 2015) and *Eucalyptus camaldulensis* (Badalamenti et al. 2018),
401 respectively.

402 On the one hand, the invasion by some species (e.g., *Populus ×canadensis*, *Sesbania punicea*
403 *and Ulmus pumila*), can be just a matter of time, with their limited invasive range depending on
404 the short time since introduction, and the likely presence of a lag phase. On the other hand, the
405 high number of potentially invasible forest categories (>70% of INNTs in at least one forest
406 category) defines a wide range of potential future invasions and range expansions. For instance,
407 *Eucalyptus camaldulensis*, *Juglans nigra*, *Ligustrum lucidum*, *Platanus hispanica*, *Populus*
408 *×canadensis*, *Prunus laurocerasus*, and *Sesbania punicea* showed a relatively large increase in
409 the occurrence in Italian regions during the last 12 years (at least +3 regions) and have the
410 potential to expand in different forest categories.

411

412 *Management implications*

413 Given the potential benefits of many non-native trees, a general ban of their use in the whole
414 country would generate socio-economic conflicts. For this reason, strategies for a sustainable use
415 of non-native tree species should be tailored at the country scale, besides accounting for the
416 global and European guidelines (Brundu and Richardson 2016, Brundu et al. 2020a). Measures
417 aiming at reducing the invasion risk, mitigating the related impacts and restoring habitat
418 structure and functions should be planned accounting for the degree of invasibility and the
419 conservation concern of forest categories, as well as for the characteristics of the invasive non-
420 native species related to that specific forest category (Richardson et al. 2007; Sádlo et al. 2017;
421 Sitzia et al. 2016).

422 At least three main management options could be considered depending on the initial condition.
423 For forest categories that are already affected by several non-native trees, silvicultural practices
424 and management options should mainly consider controlling and reducing biotic and abiotic
425 factors favoring their further diffusion, e.g., limit vegetation and soil disturbance. While forest
426 categories currently moderately invaded, i.e., by a few non-native trees, local eradication and
427 prevention actions could stop, reduce or avoid the development and spread of non-native trees.
428 Finally, for forest categories showing only a potential for non-native trees invasion, monitoring
429 and prevention measures should be prioritized.

430 Also the information on invasive non-native trees, and on the number of forest categories they
431 invaded has strong management implications, e.g., in risk assessment procedures (Bindewald et
432 al. 2021). Non-native trees that are occurring in many forest categories, due to their wide
433 presence, may be the focus of a general control of their spread. Species invading one or few
434 forest categories should be prioritized when these are forest habitat types of particular
435 conservation and scientific interest. When information on individual forest categories is lacking,
436 attention should be given to tree taxa that are showing an increasing occurrence in Italian regions

437 since they may have the potential of invading several forest categories. These species should be
438 monitored in experimental sites that allow for testing the effects of local eradication or control
439 silvicultural actions.

440 Our assessment could provide useful information for all stakeholders involved in the
441 management of forest areas, within protected areas (e.g., Natura 2000 sites), or in the definition
442 of advanced national or regional legislative tools to prevent the spread or mitigate the negative
443 effects of invasive plant species. Indeed, the array of species and invaded forest category indicate
444 that the control of non-native trees requires an integrated management strategy including
445 different silvicultural interventions (Sitzia et al. 2016, Badalamenti et al. 2018b).

446

447 *Limitations of the study*

448 We acknowledge that the classification of forest resources here reported is subjective and
449 different forest classification schemes may give different outcomes. However, we adopted a
450 classification sufficiently detailed at the tree species level, considering and taking into account
451 the large variability in structure, composition and bioclimate, which characterizes Italian forests.
452 Furthermore, the need to adopt a slender classification suggests merging some forest categories
453 including different woody species. While in some cases we included species belonging to the
454 same bioclimatic belt (e.g., Mediterranean pines), in other cases this was not possible (e.g.,
455 shrublands occurring in different biogeographical contexts).

456 The reliability of our analysis depended on the availability and quality of information across the
457 multiple sources we considered. The spatial extent and the geographical distribution of the forest
458 categories in Italy and elsewhere, certainly influenced the likelihood of the interaction with
459 INNTs. Thus, the finding of some forest categories hosting a very low number of invasive non-
460 native trees may be simply the result of a narrow distribution range. Another critical issue was
461 the attribution of INNTs to the forest categories occurring outside Italy. This was particularly
462 true for some tree species (e.g., *Fagus sylvatica* or *Pinus sylvestris*), which take part in very
463 different forest communities throughout their European distribution range, thus making it
464 difficult to use the information retrieved from literature outside Italy.

465

466 **Conclusions**

467 This is the first study that assessed the invasion status of Italian forest categories by considering
468 the main INNTs. The integration of national databases and scientific evidence reported in the
469 literature on the invasion of non-native trees can give an overview of this pressure on different
470 forest categories. Furthermore, the identification of potentially invasible forests can give an idea
471 of the level of future threat posed by non-native trees. Our results help identify which forest

472 categories and non-native trees should deserve primary attention and could inform management
473 and funding priorities to protect forests from invasive non-native tree species uncontrolled
474 spread. This information will be the basis for a shared framework to tackling tree invasions in
475 forests at the national level that integrates EU and national regulations. This approach could be
476 replicated in other countries where an assessment of forest invasions is lacking. Future studies
477 should target the impact on biodiversity, particularly by focusing on the most vulnerable forest
478 categories and the most worrying non-native tree species, as well as on the potentially most
479 vulnerable forest categories.

480

481 **CRedit authorship contribution statement**

482 **Thomas Campagnaro:** Conceptualization; Methodology; Investigation; Data curation;
483 Validation; Formal analysis; Visualization; Project administration; Roles/Writing - original draft;
484 Writing - review & editing. **Giuseppe Brundu:** Methodology; Validation; Writing - review &
485 editing. **Sabina Burrascano:** Methodology; Validation; Formal analysis; Writing - review &
486 editing. **Laura Celesti-Grapow:** Methodology; Validation; Writing - review & editing.
487 **Tommaso La Mantia:** Conceptualization; Validation; Writing - review & editing. **Tommaso**
488 **Sitzia:** Conceptualization; Validation; Writing - review & editing. **Emilio Badalamenti:**
489 Conceptualization; Methodology; Investigation; Data curation; Validation; Formal analysis;
490 Visualization; Project administration; Roles/Writing - original draft; Writing - review & editing.

491

492 **Declaration of Competing Interest**

493 The authors declare that they have no known competing financial interests or personal
494 relationships that could have appeared to influence the work reported in this paper.

495

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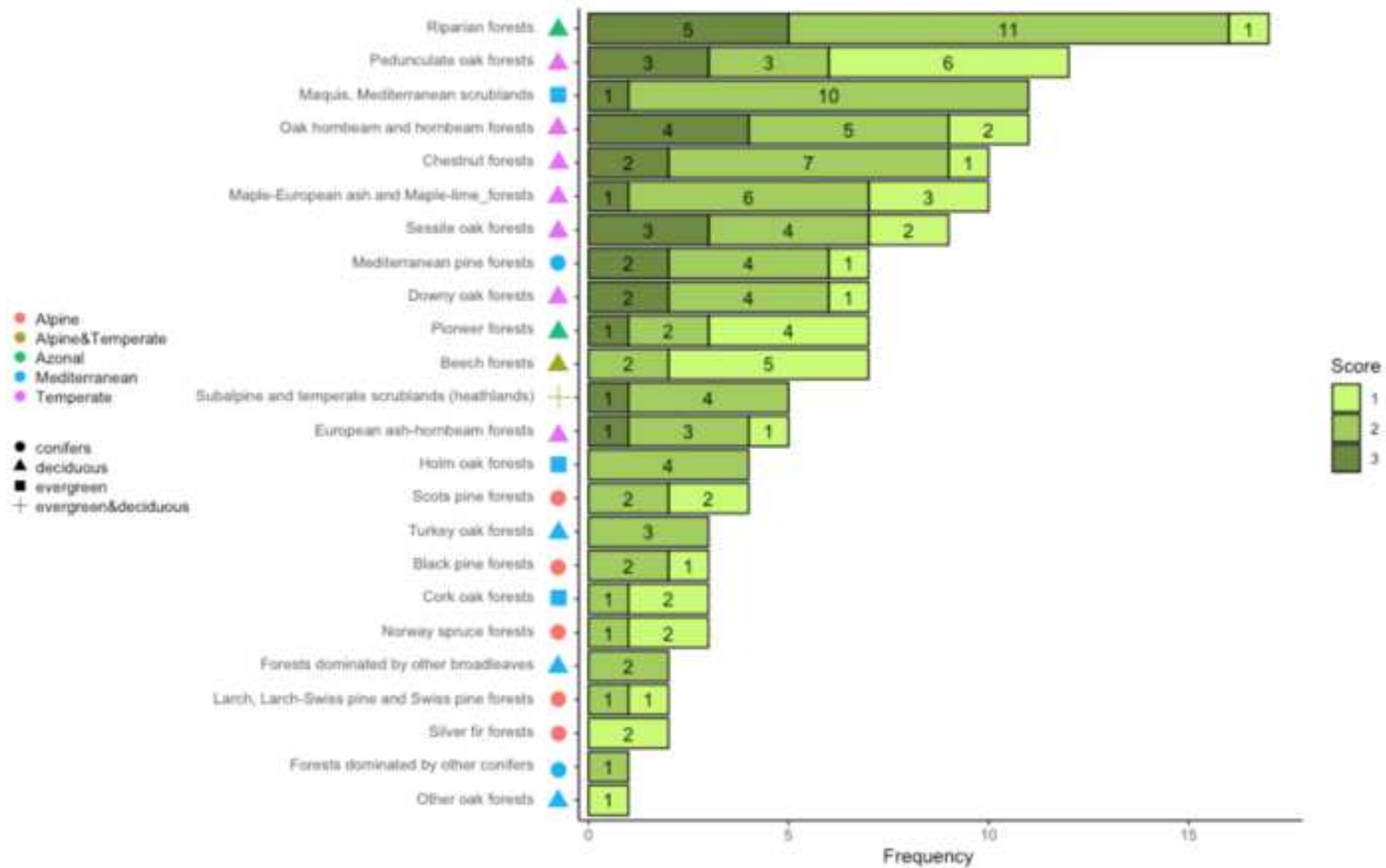
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Figure 1



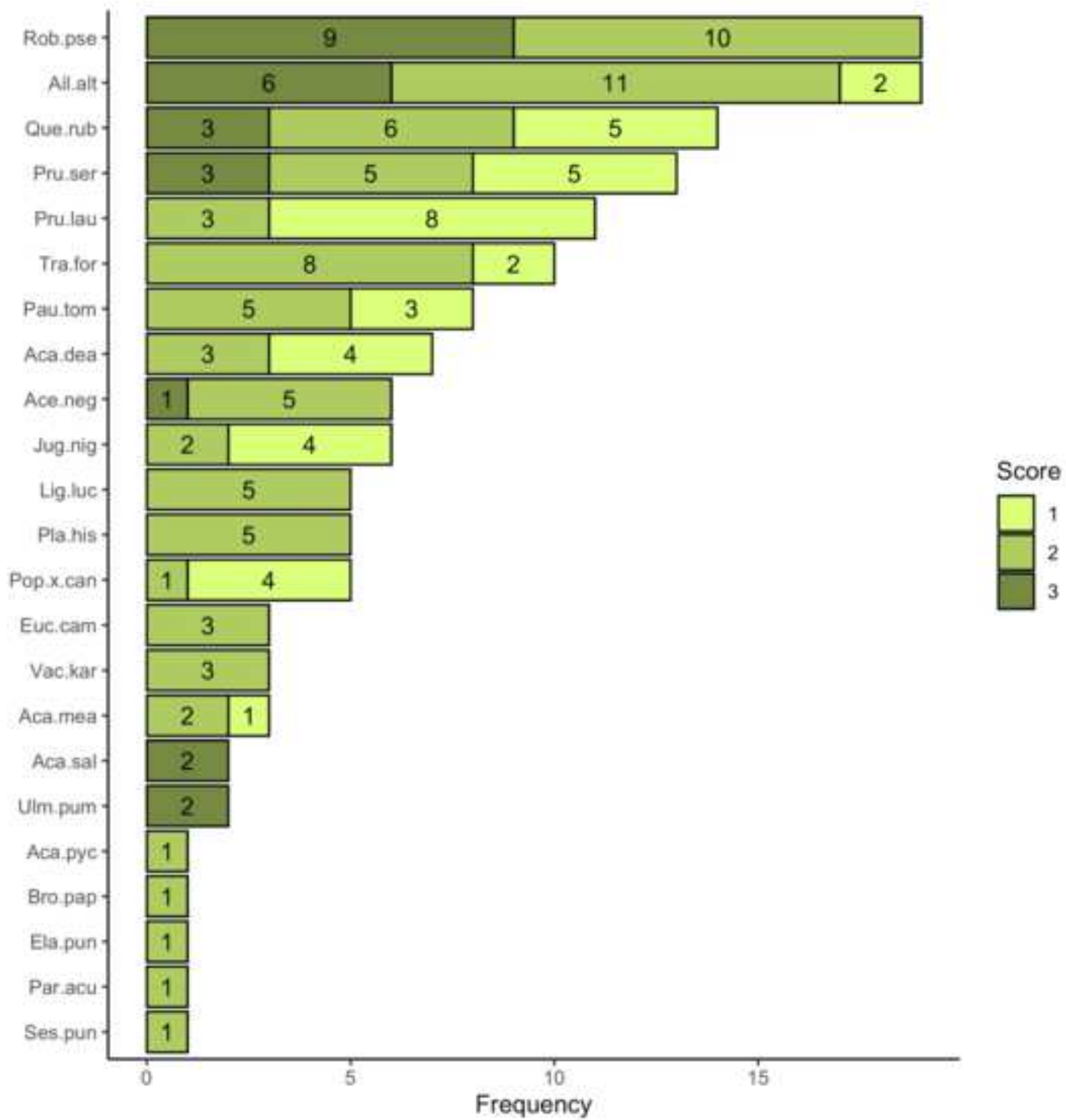
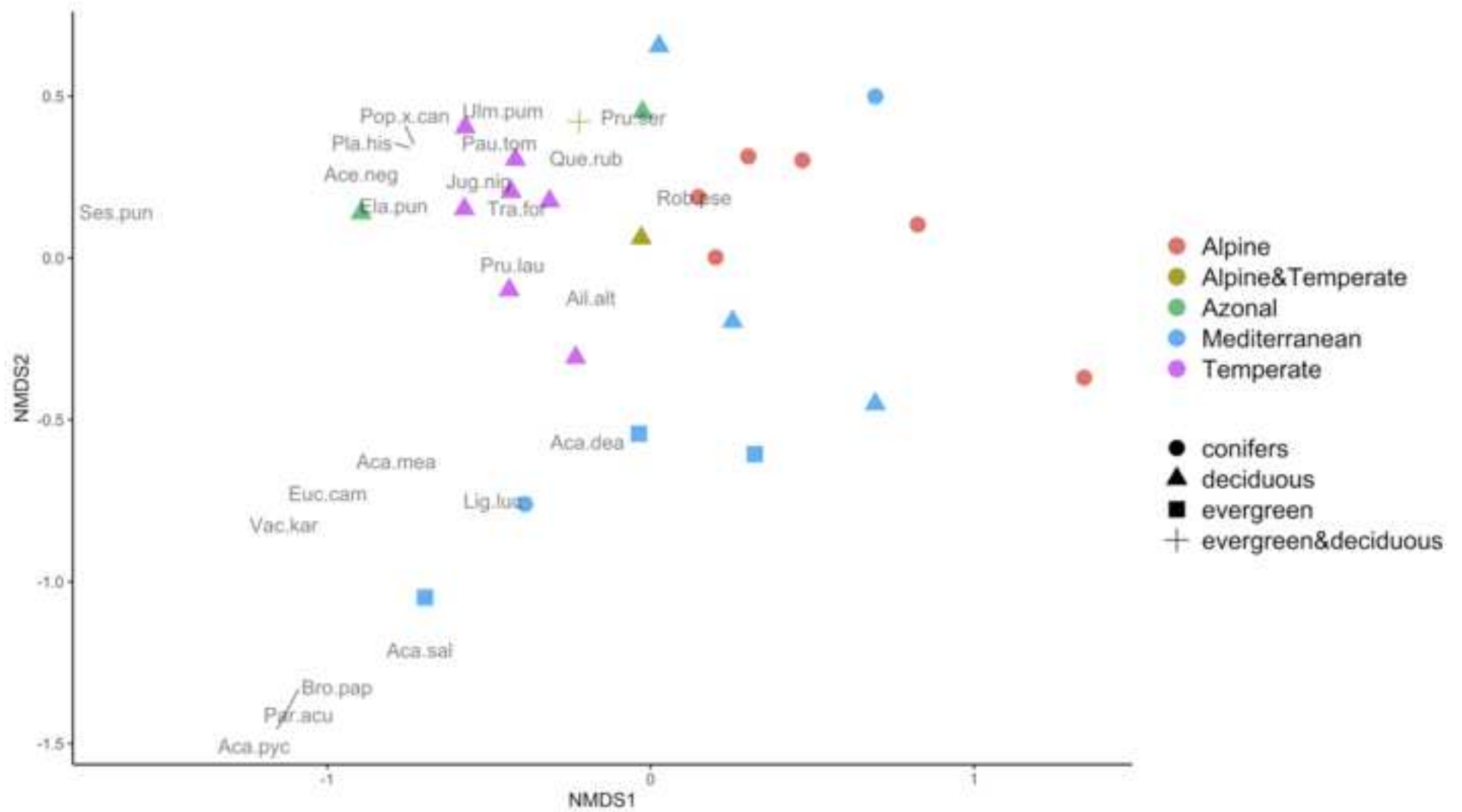


Figure 3

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1 **Tree invasions in Italian forests**

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23 **Declaration of Competing Interest**

24 The authors declare that they have no known competing financial interests or personal relationships
25 that could have appeared to influence the work reported in this paper.

Tree invasions in Italian forests

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Explanation of the spreadsheets in this excel file

"Matrix - invasion level"

We report the values

"References"

List of the full references

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Forest Category	Non-native tree species		
	<i>Acacia dealbata</i>		<i>Acacia mearnsi</i>
	Scale value	Reference	Scale value
Subalpine and temperate scrublands (heathlands)	0		0
Maquis, Mediterranean scrublands	2	Lazzaro et al. (2014), Lazzaro et al. (2020)	2
Mediterranean pine forests	1	Rodriguez et al. (2017)	0

Oak-hornbeam and hornbeam forests	0		0
Holm oak forests	2	Lazzaro et al. (2014), Lazzaro et al. (2020)	0
Cork oak forests	1	Lorenzo et al. (2013)	1
Downy oak forests	2	Lazzaro et al. (2020)	0
Sessile oak forests	0		0
Pedunculate oak forests	1	González-	0

Turkey oak forests	0		0
Other oak forests	0		0
Chestnut forests	1	González-	0
European ash-hornbeam forests	0		0
Maple-European ash and Maple-Lime forests	0		0
Pioneer forests	0		0
European beech forests	0		0
Mugo pine forests	0		0

Black pine forests	0		0
Scots pine forests	0		0
Silver fir forests	0		0
Norway spruce forests	0		0
Larch, Larch-Swiss pine and Swiss pine forests	0		0
Riparian forests	0		2
Forests dominated by sporadic conifers	0		0
Forests dominated by sporadic broadleaves	0		0
Urban forests	p		a

<i>Total assigned to 1</i>	4		1
<i>Total assigned to 2</i>	3		2
<i>Total assigned to 3</i>	0		0

<i>i</i>	<i>Acacia provincialis</i>		<i>Acacia pycnantha</i>		<i>Acacia saligna</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
	0		0		0
Brundu G. (personal observation)	0		2	Lazzaro et al. (2015), Lazzaro et al. (2020)	3
	0		0		3

	0		0		0
	0		0		0
Boudiaf et al. (2013)	0		0		0
	0		0		0
	0		0		0
	0		0		0
	0		0		0

	0		0		0
	0		0		0
	0		0		0
	0		0		0
	0		0		0
Brundu et al.	0		0		0
	0		0		0
	0		0		0
	a		a		d

	0		0		0
	0		1		0
	0		0		2

	<i>Acer negundo</i>		<i>Ailanthus altissima</i>		<i>Broussonetia p</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
	0		2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013a), Lazzaro et al. (2020)	0
Biondi et al. (2009), Del Vecchio et al. (2013), Iamónico et al. (2014), Lazzaro et al. (2020)	0		2	Lazzaro et al. (2020)	2
Biondi et al. (2009), Del Vecchio et al. (2013), Bonari et al. (2017), Badalamenti et al. (2018a), Lazzaro et al. (2020)	0		3	Badalamenti et al. (2012), Angelini et al. (2016), Lazzaro et al. (2020)	0

	2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014a)	3	Ebone et al. (2011), Angelucci (2011), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013a), Lazzaro et al. (2020)	0
	0		2	Badalamenti et al. (2012)	0
	0		2	Badalamenti et al. (2012)	0
	0		3	Allegrezza et al. (2006), Biondi et al. (2009), Badalamenti et al. (2012), Lazzaro et al. (2020)	0
	2	Gruppo di	3	Biondi et al.	0
	2	Gruppo di	2	Del Favero	0

	0		2	Badalamenti et al. (2012)	0
	0		1	Dinucă (2021)	0
	2	Gruppo di	3	Badalamenti	0
	0		2	Radtke et al. (2013)	0
	2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014a)	2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013a), Angelini et al. (2016)	0
	0		0		0
	0		2	Badalamenti et al. (2012)	0
	0		0		0

	0		2	Badalamenti	0
	0		1	Motard et al.	0
	0		0		0
	0		0		0
	0		0		0
	3	Banfi and	3	Badalamenti	0
	0		0		0
	0		2	Badalamenti	0
	d	Sitzia et al.	d	Celesti-	d

	0		2		0
	5		11		1
	1		6		0

<hr/>							
<i>aprifera</i>		<i>Elaeagnus pungens</i>		<i>Eucalyptus camaldulensis</i>		<i>Juglans nigra</i>	
Reference	Scale value	Reference	Scale value	Reference	Scale value	Reference	Scale value
	0		0		0		0
Lazzaro et al. (2020)	0		2	Guarino et al. (2021)	0		0
	0		2	Biondi et al. (2009)	0		0

	0		0		2
	0		0		0
	0		0		0
	0		0		2
	0		0		0
	0		0		1

	0		0		0
	0		0		0
	2	Antonietti and	0		0
	0		0		0
	0		0		1
	0		0		1
	0		0		0
	0		0		0

	0		0		0
	0		0		0
	0		0		0
	0		0		0
	0		0		0
	0		2	Badalamenti	1
	0		0		0
	0		0		0
Scuderi and	p		p		p

	0		0		4
	1		3		2
	0		0		0

	<i>Ligustrum lucidum</i>		<i>Parkinsonia aculeata</i>		<i>Paulownia tom</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
	0		0		2
	2	Lazzaro et al. (2020)	2	Angelini et al. (2016)	0
	2	Galasso et al. (2016), Lazzaro et al. (2020)	0		0

Scatamacchia (2011)	0		0		2
	2	Prosser and Bertolli (2015), Lazzaro et al. (2020)	0		0
	0		0		0
Cicinelli et al. (2018)	0		0		0
	0		0		1
Šebesta et al.	0		0		1

	0		0		0
	0		0		0
	0		0		2
	2	Prosser and Bertolli (2015)	0		0
Šebesta et al. (2021)	0		0		2
Šebesta et al. (2021)	0		0		0
	0		0		1
	0		0		0

	0		0		0
	0		0		0
	0		0		0
	0		0		0
	0		0		0
Šebesta et al.	2	Lastrucci et al.	0		2
	0		0		0
	0		0		0
	p		d	Scuderi and	p

	0		0		3
	5		1		5
	0		0		0

<i>mentosa</i>	<i>Platanus hispanica</i>		<i>Populus x canadensis</i>		<i>Prunus lauroce</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2020)	0		0		0
	0		0		2
	0		0		0

Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2020)	2	Del Favero (2002), Mason (2004)	1	Šebesta et al. (2021)	1
	0		0		0
	0		0		0
	0		0		1
Maringer et al.	2	Del Favero	0		1
Maringer et al.	2	Del Favero	1	Šebesta et al.	1

	0		0		0
	0		0		0
Gruppo di	0		0		2
	0		0		1
Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2020)	2	Del Favero (2002)	1	Šebesta et al. (2021)	1
	0		1	Šebesta et al. (2021)	0
Maringer et al. (2012)	0		0		1
	0		0		0

	0		0		0
	0		0		0
	0		0		1
	0		0		0
	0		0		0
Gruppo di	2	Del Favero	2	Cicinelli et al.	2
	0		0		0
	0		0		0
Essl (2007),	p		p		p

	0		4		8
	5		1		3
	0		0		0

<i>rasus</i>	<i>Prunus serotina</i>		<i>Quercus rubra</i>		<i>Robinia pseudo</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
	2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013b)	2	Biondi et al. (2009), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013c)	3
Lucchese (2017)	0		0		0
	0		0		2

Hättenschwiler and Körner (2003)	3	Del Favero (2002), Biondi et al. (2009), Angelucci (2011), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013b), Terwei et al. (2013), Angelini et al. (2016), Lazzaro et al. (2020)	3	Biondi et al. (2009), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014b), Prosser and Bertolli (2015), Lazzaro et al. (2020)	3
	0		0		2
	0		0		0
Conedera et al. (2018)	0		2	Lazzaro et al. (2020)	3
Conedera et	2	Del Favero	3	Del Favero	3
Descombes et	3	Del Favero	3	Del Favero	3

	2	Annighöfer et al. (2012)	2	Annighöfer et al. (2012), Lazzaro et al. (2020)	0
	0		0		0
Antonietti and	2	Del Favero	2	Del Favero	3
Fehr and Burga (2016)	0		0		3
Conedera et al. (2018)	0		2	Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2013c)	3
	1	Vanhellemont (2009)	1	Major et al. (2013), Wagner et al. (2017)	2
Rusterholz et al. (2018)	1	Maringer et al. (2012)	1	Wagner et al. (2017), Medvecká et al. (2018), Bindewald et al. (2021)	2
	0		0		0

	1	Godefroid et	0		2
	2	Del Favero	1	Halarewicz	2
Wagner et al.	0		1	Wagner et al.	0
	1	Starfinger et	1	Wagner et al.	2
	1	Vanhellemont	0		2
Gruppo di	3	Gruppo di	2	Gruppo di	3
	0		0		2
	0		0		2
Sitzia et al.	p		p		d

	5		5		0
	5		6		10
	3		3		9

<i>Acacia</i>	<i>Robinia viscosa</i>		<i>Sesbania punicea</i>		<i>Ulmus pumila</i>
Reference	Scale value	Reference	Scale value	Reference	Scale value
Biondi et al. (2009), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014b), Prosser and Bertolli (2015), Lazzaro et al. (2020)	0		0		0
	0		0		0
Angelini et al. (2016)	0		0		0

Angeles-Gras (1991; as cited in Sitzia 2009), Del Favero (2002), Angelucci (2011), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014b), Prosser and Bertolli (2015), Angelini et al. (2016), Lazzaro et al. (2020)	0		0		0
Lazzaro et al. (2020)	0		0		0
	0		0		0
Del Favero (2002), Sitzia (2009), Prosser and Betolli (2015), Angelini et al. (2016), Lazzaro et al. (2020)	0		0		0
Angeles-Gras	0		0		0
Del Favero	0		0		0

	0		0		0
	0		0		0
Del Favero	0		0		0
Del Favero (2002), Del Favero (2004), Sitzia (2009), Odasso et al. (2018)	0		0		0
Del Favero (2002), Sitzia (2009), Gruppo di Lavoro Specie Esotiche della Regione Piemonte (2014b), Prosser and Betolli (2015), Angelini et al. (2016), Lazzaro et al. (2020)	0		0		0
Sitzia (2009), Sitzia et al. (2012)	0		0		3
Prosser and Betolli (2015), Angelini et al. (2016), Lazzaro et al. (2020)	0		0		0
	0		0		0

Sitzia (2009)	0		0		0
Del Favero	0		0		0
	0		0		0
Campagnaro	0		0		0
Sitzia (2009)	0		0		0
Del Favero	0		2	Brunel et al.	3
Lazzaro et al.	0		0		0
Biondi et al.	0		0		0
Sitzia et al.	a		0		p

	0		0		0
	0		1		0
	0		0		2

		<i>Trachycarpus fortunei</i>		<i>Vachellia karroo</i>
Reference	Scale value	Reference	Scale value	Reference
	0		0	
	0		2	Biondi et al. (2009), Angelini et al. (2016), Lazzaro et al. (2020)
	0		2	Biondi et al. (2009), Gioia (2021)

	2	Lazzaro et al. (2020)	0	
	0		0	
	0		0	
	2	Fehr and Burga (2016)	0	
	2	Fehr and Burga	0	
	1	Conedera et al.	0	

	0		0	
	0		0	
	2	Fehr and Burga	0	
	2	Fehr and Burga (2016)	0	
	2	Fehr and Burga (2016), Lazzaro et al. (2020)	0	
Brunet et al. (2013), Prosser and Bertolli (2015), Mainetti and Lonati (2017)	2	Fehr and Burga (2016)	0	
	1	Maringer et al. (2012)	0	
	0		0	

	0		0	
	0		0	
	0		0	
	0		0	
	0		0	
Prosser and	2	Bona and Federici	2	La Mantia T.
	0		0	
	0		0	
Maineti and	p	Sitzia et al. (2016)	d	

	2		0
	8		3
	0		0

Total assigned to 1	Total assigned to 2	Total assigned to 3
0	4	1
0	10	1
1	4	2

2	5	4
0	4	0
2	1	0
1	4	2
2	4	3
6	3	3

0	3	0
1	0	0
1	7	2
1	3	1
3	6	1
4	2	1
5	2	0
0	0	0

1	2	0
2	2	0
2	0	0
2	1	0
1	1	0
1	11	5
0	1	0
0	2	0

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1 **Tree invasions in Italian forests**

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13 **Supplementary Information 2**

14 Table SI2: Status (CAS: casual, NAT: naturalized, INV: invasive) of the selected 24 non-native tree species in 2009, 2018, and 2021. Numbers
15 indicate the number of regions with such status in Italy.

Invasive non-native tree	Celesti-Grapow et al. 2009			Total (2009)	Galasso et al. 2018			Total (2018)	Acta Plantarum 2021			Total (2021)
	CAS	NAT	INV		CAS	NAT	INV		CAS	NAT	INV	
<i>Acacia dealbata</i> subsp. <i>dealbata</i>	5	1	2	8	5	2	2	9	5	3	2	10
<i>Acacia mearnsii</i>	0	0	1	1	0	0	1	1	0	0	1	1
<i>Acacia provincialis</i>	0	0	1	1	1	1	1	3	2	1	1	4

<i>Acacia pycnantha</i>	2	0	0	2	0	1	1	2	0	1	1	2
<i>Acacia saligna</i>	2	4	3	9	2	3	5	10	2	3	5	10
<i>Acer negundo</i>	7	6	4	17	7	6	6	19	7	6	6	19
<i>Ailanthus altissima</i>	0	0	20	20	0	0	20	20	0	0	20	20
<i>Broussonetia papyrifera</i>	5	10	3	18	4	10	4	18	5	10	4	19
<i>Elaeagnus pungens</i>	2	2	0	4	6	3	1	10	6	2	2	10
<i>Eucalyptus camaldulensis</i> subsp. <i>camaldulensis</i>	7	2	0	9	5	5	1	11	5	5	1	11
<i>Juglans nigra</i>	6	0	0	6	7	1	1	9	8	2	1	11
<i>Ligustrum lucidum</i>	7	9	0	16	5	10	2	17	5	10	2	17
<i>Parkinsonia aculeata</i>	2	0	1	3	4	1	1	6	3	1	2	6
<i>Paulownia tomentosa</i>	5	5	0	10	8	6	1	15	10	6	1	17
<i>Platanus hispanica</i>	8	8	0	16	7	10	0	17	7	10	1	18

<i>Populus ×canadensis</i>	12	5	0	17	10	8	0	18	9	7	2	18
<i>Prunus laurocerasus</i>	9	1	1	11	10	1	3	14	11	1	4	16
<i>Prunus serotina</i>	2	0	3	5	3	1	3	7	3	1	3	7
<i>Quercus rubra</i>	4	1	2	7	5	2	2	9	7	2	2	11
<i>Robinia pseudoacacia</i>	0	3	17	20	0	3	17	20	0	3	17	20
<i>Robinia viscosa</i>	1	0	0	1	0	0	1	1	0	0	1	1
<i>Sesbania punicea</i>	1	0	0	1	3	2	0	5	2	1	2	5
<i>Trachycarpus fortunei</i>	3	1	2	6	5	2	3	10	7	2	3	12
<i>Ulmus pumila</i>	5	2	0	7	3	5	2	10	3	7	2	12
<i>Vachellia karroo</i> (= <i>Acacia karroo</i>)	0	4	0	4	2	5	0	7	2	4	1	7
	94	64	60	218	99	86	78	263	108	88	81	277

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