

1 **Biodegradation potential of bacterial isolates related to the genus *Thalassospira* in**
2 **hypersaline conditions**

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25 **ABSTRACT**

26 Three bacterial species related to the genus *Thalassospira* (*T. lucentensis*, *T. xianhensis*
27 and *T. profundimaris*), isolated from polluted sediment and seawater samples collected from
28 Priolo Bay (eastern coast of Sicily, Ionian Sea), were analyzed for their biotechnological
29 potential. For this purpose, were evaluate the presence/absence of specific catabolic genes
30 associated to aliphatic and aromatic hydrocarbons metabolism, the production of biosurfactants
31 and emulsification activity, the capability to degrade oil derived linear, branched, cyclic alkanes
32 and polycyclic aromatic hydrocarbons (PAHs). Alkane hydroxylase gene (alkano-
33 monooxygenase *alkB* and citochrome *P450*) and biphenyl dioxygenase gene (*bphA* and *xylC1*)
34 were present in genome of strains, confirming their hydrocarbons degrading capability. All
35 strains of *Thalassospira* produced biosurfactants and showed emulsification activity. The two-
36 dimensional gas chromatography analysis (GC×GC) showed that they were able to degrade oil
37 fractions with the capacity ranged between 77-91%.

38 Data obtained in this study demonstrated the potential of *Thalassospira* for oil
39 bioremediation in oil-contaminated coastal area, suggesting that the three bacterial strains
40 isolated could be used for bioremediation purposes and treatment of contaminated ecosystems.

41

42 **Keywords:**

43 Bioremediation, Marine Environment, Marine Oil Pollution, *Thalassospira*, Comprehensive
44 two-dimensional gas chromatography

45 **1. Introduction**

46 Bioremediation consist in the conversion of environmental pollutants in non-toxic
47 compounds by microorganisms. During this transformation, organic and toxic compounds were
48 convert into less complex and toxic metabolites, and with the mineralization process in
49 inorganic compounds, H₂O and CO₂ (if the microorganisms are in aerobic condition) and CH₄
50 (if the microorganisms are in anaerobic condition) (Nikolopoulou and Kalogeraki, 2010).. The
51 successes of bio-conversion depend on many factors, such as: pH, temperature, oxygen
52 concentration, availability of nutrients and/or chemicals, type of pollutants, cellular transport
53 properties, and chemical partitioning in growth medium (Singh and Ward, 2004). On the
54 foregoing basis, bioremediation techniques have proved advantageous, and the study and the
55 optimization of this eco-friendly technique can be a valid alternative to chemical and physical
56 techniques for cleaning up of oil polluted marine environments (Alkatib et al., 2011). Oil
57 derived hydrocarbon-degrading microorganisms are widely distributed in marine, freshwater
58 and soil habitats (Chaillan et al., 2004; Yakimov et al., 2007, Catania et al., 2015, Catania et al,
59 2017, Djahnit et al., 2019). Despite the chemical stability, very low water solubility and
60 bioavailability of hydrocarbon compounds, numerous bacterial species have been described as
61 efficient break-downer of oil-pollutant as widely evidenced in literature (Cerniglia 1993;
62 Sutherland et al. 1995; Kanaly and Harayama 2000; Peng et al. 2008; Haritash and Kaushik
63 2009, Catania et al., 2020a). Scientific interests and biotechnological applications about
64 hydrocarbon-utilizing bacteria increased in recent years (Margesin and Schinner 2001; Le
65 Borgne et al. 2008, Scaffaro et al., 2017, Catania et al., 2020b) and several studies on the
66 biodegradation of crude oil and/or PAHs and other aromatic compounds have been performed
67 (Kasai et al. 2002; Melcher et al. 2002; García et al. 2005; Li and Bai 2005; McKew et al. 2007,
68 Catania et al., 2018). Bacterial degradation of PAHs was firstly described by Zhao et al. (2009),
69 in a phenanthrene degrading consortium from a saline soil which was developed at 10% salinity

70 and included several strains belonging to different genera, among which *Thalassospira spp.*
71 (Hassanshahian et al., 2010 and 2012). Members of *Thalassospira* genus are mesophilic,
72 halophilic, Gram-negative, oxidase and catalase positive, aerobic and often facultative
73 anaerobic α -Proteobacteria with the ability to utilize hydrocarbons or xenobiotics compound
74 such as polyvinyl-alcohol as source of carbon and energy. (Baldani et al., 2014; Kiseleva et al.,
75 2015). Recent studies revealed and demonstrated widely as the genus *Thalassospira* is
76 correlated to hydrocarbons pollution in marine environment (Lopez Lopez, A. et al 2002, Chen,
77 et al. 2007, Kodoma et. al. 2008, Zhao et al. 2010, Thompson et al. 2017, Catania et al., 2018,),
78 Furthermore, numerous scientific evidences showed that some *Thalassospira* species
79 were the dominant members in the hydrocarbon-degrading consortia, suggesting a central role
80 in biodegradation and restoration of marine ecosystems.(Muangchida, et al. 2018).

81 however, the knowledge on some species of the genus *Thalassospira* is still limited, and
82 the capabilities and mechanisms of degradation of contaminants are not well known.

83 In this study, for the first time, the biotechnological potential (for bioremediation
84 application) of species. *lucentensis*, *xianhensis* and *profundimaris* related to the genus
85 *Thalassospira (T)* were analyzed. In particular were evaluate: i) presence/absence of specific
86 catabolic genes, ii) production of biosurfactants and iii) capability to degrade different oil
87 fractions (linear, branched, cyclic alkanes and polycyclic aromatic hydrocarbons). The potential
88 of hydrocarbons degradation was evaluate using a comprehensive two-dimensional gas
89 chromatography method (GC×GC) with a dual detection, flame ionization detector (FID) and
90 single quadrupole (Q MS). The use of this instrument has proved essential in order to achieve
91 a detailed characterization and quantification of the different chemical families (Tranchida et
92 al., 2009, Zoccali et al., 2015, 2018).

93

94 2. Materials and methods

95 2.1. Isolation of bacterial strains

96 Bacterial strains were isolated from enrichment cultures obtained from seawater and
97 sediment samples collected in Priolo Bay (South-East Sicily, Italy, Mediterranean Sea) (Catania
98 et al., 2015). As previously reported (Catania et al., 2015) different enrichment sets were
99 performed by supplementing the ONR7a medium (Dyksterhouse et al., 1995), with n-alkanes
100 mixture (C₁₆, C₁₈ and C₂₄), PAHs (phenanthrene, pyrene, biphenyl and dibenzothiophene) or
101 crude oil (Arabian Light Crude Oil). Phenotypically different colonies obtained from the plates
102 were purified, and transferred to fresh medium with hydrocarbons as unique carbon sources
103 (crude oil, n-Alkanes mixture, PAHs). Plate without hydrocarbons was prepared to eliminate
104 autotrophs and agar-utilizing bacteria. The procedure was repeated, and only isolates exhibiting
105 pronounced growth on crude oil, *n*-alkanes mixture or PAHs were stored in stock media with
106 glycerol at -20±1°C for further characterization (Hassanshahian et al., 2012). On the base of the
107 data present in literature and 16S rDNA analysis, the isolates more interest were selected to
108 following analysis.

109

110

111 2.2. Identification of the isolates

112

113 2.2.1. Molecular characterization of isolates (16S rDNA sequencing)

114 Total DNA extraction of bacterial strains obtained was performed through the MasterPure
115 Complete DNA&RNA Purification Kit (Epicenter, Biotechnologies, Madison, WI) in
116 accordance with manufacture's protocol. The 16S rDNA loci were amplified using the 27F (5'-
117 AGAGTTTGATCCTGGCTCAG-3', Lane, 1991) primer and the 1492R (5'-
118 TACGGYTACCTTGTTACGACT-3', Lane, 1991) universal primer. Polymerase chain

119 reaction (PCR) was carried out in 50 μ L of reaction mixture containing 1x reaction buffer, 1x
120 solution Q (both from QIAGEN), 1 μ M of each primer, 200 μ M dNTP (Gibco), 1 μ L of
121 template and 2.5 U of Qiagen *Taq* polymerase. The PCR conditions were as follows: 95 $^{\circ}$ C for
122 5 min (1 cycle); 94 $^{\circ}$ C for 1 min, 50 $^{\circ}$ C for 1 min and 72 $^{\circ}$ C for 2 min (35 cycles); with a final
123 extension step at 72 $^{\circ}$ C for 10 min. PCR products were sequenced using Macrogen Service
124 (Macrogen, Europe). The analysis of the sequences was performed as previously described by
125 Yakimov et al. (2005). Sequences similarity of individual inserts was analyzed by the FASTA
126 program Nucleotide Database Query, available through the EMBL-European Bioinformatics
127 Institute. The phylogenetic affiliation of the sequenced strains was performed as described by
128 Yakimov et al. (2006). 16S rDNA were deposited in GeneBank with Accession Number:
129 KT348286, KT348281 and KT348287.

130

131 2.2.2. *Biochemical identification*

132

133 In addition to the morphological examination carried out in Catania et al. 2015 was
134 performed: i) Oxidase/Catalase test; ii) cellular shape and mobility. These tests done according
135 to the Bergey's manual for identification taxonomy (Holt et al. 1998).

136

137 2.2.3. *Qualitative analysis of catabolic genes*

138 Qualitative analysis (presence/absence) of catabolic genes [cytochrome (*P450*), alkane
139 hydroxylase (*alkB*) and biphenyl dioxygenase iron sulfur protein (*bphA* gene) were carried out
140 using different couples of primers. The presence/ absence of *P450* gene (~800 bp product) was
141 assayed applying the primers P450F (5'TGTCGGTTGAAATGTTTCATYGCNMTGGAYCC-
142 3') and P450R (5'TGCAGTTCGGCAAGGCGGTTDCCSRYRCAVCKRTG-3')(Kubota *et al.*,
143 2005); while to assay the presence/ absence of alkane-monoxygenase gene (~550 bp product)

144 was used degenerate primers alkBwf (5'-AAYACNGCNCAYGARCTNGGVCAYAA-3' and
145 alkBwr (5'-GCRTGRTGRTCHGARTGNCGYTG-3') (Kloos et al., 2006), and couple primers
146 TS2S (5'-AAYAGAGCTCAYGARYTRGGTCAYAAG-3') and deg1RE (5'-
147 GTGGAATTCGCRTGTCIGARTG-3') (Smits et al., 1999); the primers described by Ryslava
148 et al. (2003) 352F (5'-TTCACCTGCASCTAYCACGGC -3') and 1178R (5'-
149 ACCCAGTTYTCDCCRTCGTCCTGC-3') were used for the detection of *bphA* gene (~800 bp
150 product). The PCR reaction mixtures contained 50 ng of DNA, 1× Qiagen reaction buffer, 1×
151 solution Q (Qiagen), 0.4 mM P450 primers (or 0.8 mM for *alkB* primers and/ or 10 pmol for
152 *bphA* primers), 200µM dNTPs (Gibco) e 2.5 U di Qiagen Taq Polymerase in a final volume of
153 50 ml; while the PCR reaction mixtures employed with the primers TS2S and deg1RE was
154 carried out as described by Innis et al (1990). The PCR amplification of alkane-monoxygenase
155 gene and P450 gene was carried out on a Mastercycler Gradient (Eppendorf) with the following
156 conditions: initial denaturation 4 min at 94°C (1 cycle); 30 s at 94°C, 30 s at 52°C (for the
157 degenerate primers alkBwf and alkBwr, it was 30 s at 55°C) and 1 min at 72°C (32 cycles);
158 and a final extension at 72°C for 10 min; while for the amplification of alkane-monoxygenase
159 gene with the primers TS2S and deg1RE and *bphA* gene were used respectively the condition
160 described by Smits et al. (1999) (1 cycle 4 min at 95°C; 25 cycles: 45 s at 95°C, 1 min at 40°C,
161 1 min at 72°C; 1 cycle 5 min at 72°C, indefinitely at 4°C) and Dudášová et al. (2014)(at 95 °C
162 for 2 min; 30 cycles: 95 °C for 1 min, 60 °C for 1 min, 72 °C for 1 min and 30 s; and at 72 °C
163 for 10 min). The gene amplicons obtained with PCR reaction were separated by electrophoresis
164 on agarose (1.5%) gel (Whang et al.2010).

165 2.4.1. Starter culture and growth in laboratory conditions

166 The initial cultures were prepared by inoculating one loop of microbial cells into 10 mL
167 of ONR7a mineral medium based on seawater (Dyksterhouse et al., 1995) supplied with 5% of
168 Sodium Acetate. After growing in a rotary shaker (New Brunswick C24KC, Edison NJ, USA;

169 150 ×g) at 25±1 °C for two days the culture was used to Biosurfactant Production,
170 Emulsification Activity and Liquid Surface Tension, besides 500 µL of the seed-culture broth
171 were washed (×2) with PBS 1× and ready for Biodegradation experiment.

172

173 *2.3. Biosurfactant Production, Emulsification Activity and Liquid Surface Tension*

174

175 Biosurfactant production was screened by two distinct methods: *i*) The Drop collapse test
176 and *ii*) Oil Spreading. In the Drop collapse test, 7 µl of (sterile) mineral oil was added to each
177 well of a 48-well microtiter plate. Covered plates were equilibrated for 24 hours at room
178 temperature before adding culture supernatant. 20 µl of the culture supernatant was added on
179 the (sterile) oil surface in the microtiter well. Sodium dodecyl sulfate (SDS, Sigma-Aldrich,
180 Milan) and sterile distilled water were used respectively as positive and negative control
181 suspensions (Manjoubi et al., 2013). In the Oil Spreading, 20 ml of distilled water was added
182 to an empty Petri dish followed by addition of 10 µl of crude oil to the surface of the water. An
183 aliquot of 10 microliters of bacterial culture were then added to the oil surface. The diameter of
184 the clear zone around the bacterial suspension corresponds to the biosurfactant concentration
185 (Youssef et al., 2004). Emulsification activity was measured with the Emulsification index
186 (E_{24}). The E_{24} was determined by adding 2 ml of hydrocarbon (benzene- C_6H_6 and diesel) to
187 same volume of different bacterial culture. After mixing with a vortex for 2 min and leaving to
188 stand for 24 h, the E_{24} index is given as percentage of height of emulsified layer (in millimeters)
189 divided by total height of the liquid column (in millimeters) (Iqbal et al. 1995; Cappello et al.,
190 2012). The Surface Tension of bacterial cultures was measured after 10 days of growth, by the
191 Wilhelmy plate method, using a digital tensiometer (Gibertini, Italy), in accordance with the
192 manufacturer's instructions. The surface tension was expressed in units of $mN m^{-1}$.

193

194 *2.4. Biodegradation experiment*

195 *2.4.2. Experimental set-up of microcosm systems*

196

197 The microcosm systems were performed in 250 mL sterilized Erlenmeyer flasks.
198 Microcosms were incubated at $25\pm 1^\circ\text{C}$ for 15 days with shaking ($100 \times g$) and were realized in
199 100 mL of mineral ONR7a medium supplemented with 0.1% (v/v) -of sterile crude oil (Arabian
200 Light Crude Oil; ENI Technology S.p.A.) and inoculating single bacterial strains. Crude oil
201 used in the experimentation was introduced in microcosm systems after physical weathering
202 ($100 \times g$, $25\pm 1^\circ\text{C}$ for 48 h), and supplemented with 0.1% (v/v) of heptamethylnonane ($\text{C}_{16}\text{H}_{34}$,
203 Sigma-Aldrich, Milan) as internal spike for measure of bio-degradation rate. Abiotic
204 microcosms untreated (without inoculation) were used as negative control. All
205 experimentations were carried out in twice.

206

207 *2.4.3. Hydrocarbons extraction of microcosm system*

208 After acidification TERHCs from microcosm system were extracted at room temperature
209 on a shaking table using dichloromethane (CH_2Cl_2 , Sigma-Aldrich, Milan, Italy; 10% v/v). This
210 procedure was repeated thrice, and the CH_2Cl_2 phase was combined and treated with sodium
211 sulfate anhydrous (Na_2SO_4 , Sigma-Aldrich, Milan, Italy) in order to remove any residual water
212 (Ehrhardt et al., 1991; Wang et al., 1998; Dutta and Harayama, 2001; Denaro et al., 2005). The
213 extracts were concentrated by rotary evaporation (Rotavapor model R110; Büchi Labortechnik
214 AG, Switzerland) at room temperature ($\sim 30^\circ\text{C}$), followed by evaporation under a stream of
215 nitrogen.

216

217 *2.4.4. Cryogenic Modulation GC×GC-Q MS/FID Analyses*

218

219 Quali-, quantitative analysis of hydrocarbons from microcosm were analyzed by
220 Cryogenic Modulation (CM) GC×GC-Q MS/FID consisting of two GC-2010 gas
221 chromatographs (Shimadzu Corporation, Kyoto, Japan), a TQ8040 triple quadrupole mass
222 spectrometer (Shimadzu) and a flame ionization detector (FID). Protocol use for extraction was
223 following the 3510 EPA (Environmental Protection Agency). After 15 days of incubation
224 sample were collected from microcosms to analyze the residual hydrocarbons. All Data were
225 collected by the GCMS Solution v.4.41 software (Shimadzu); the ChromSquare ver 2.3
226 software was used for GC × GC data processing (Shimadzu). MS spectral matching was
227 performed automatically by using the ChromSquare ver 2.3 software and the Wiley Registry
228 11th Edition/NIST 2017 Mass Spectral Library (Wiley, NIST). Moreover, the identification
229 was performed by comparison with linear retention index, and elution order. The first gas
230 chromatograph (GC1) was equipped with an AOC-20i auto-injector, and a split–splitless
231 injector (350°C). The primary column (¹D) was an SLB-35 ms [(silphenylene polymer, which
232 can be considered equivalent in polarity to poly(35%diphenyl/65% dimethylsiloxane)] with
233 dimensions 30 m × 0.25 mm id × 0.25 μm *d_f*. A 1.0 m × 0.10 mm id uncoated column was used
234 to create the modulator loop. The second dimension (²D) column was an SLB-5 ms [(polymer
235 which can be considered equivalent in polarity to poly(5%diphenyl/95% dimethylsiloxane)]
236 with dimensions 1.0 m × 0.10 mm id × 0.10 μm *d_f*. All the columns used were provided by
237 Merck Life Science (Merck KGaA, Darmstadt, Germany). The connections between the ¹D and
238 ²D columns, and the modulator loop, were made by using two SilTite mini unions (Trajan,
239 Ringwood, Victoria, Australia). The outlet of the ¹D column was connected to a MXT “Y”-
240 Union (Restek); the latter was then linked to a 0.4 m × 0.10 mm id (for FID analysis) and to a
241 1.0 m × 0.10 mm id uncoated column (for MS analysis). An average of 63% of the column
242 effluent was directed to the FID with the objective of quantification, while the remaining
243 fraction reached the MS for identification purposes. Helium was supplied at the GC1 inlet at a

244 pressure of 289.5 kPa (constant average linear velocity mode); volume and mode of injection:
245 1 μL in the split mode (25:1). GC1 temperature program: 80–350°C (20 min) at 3°C min⁻¹; GC2
246 temperature program: 60-350 °C (13.3 min) at 3 °C min⁻¹. Modulation was performed by using
247 a cryogenic fluid-free modulator; modulation period was 4 s (the heating step was performed
248 at 400 °C, for 0.35 s). MS conditions: the temperature of the interface was 330°C; the ion source
249 temperature was 280 °C, with analyte fragmentation induced by electron ionization (70 eV). A
250 spectral generation frequency of 33 *Hz* was applied with the following mass range *m/z* 45-360.
251 FID (360 °C) sampling frequency was 125 *Hz*.
252

253 **3. Results**

254

255 3.1 Taxonomic and Biochemical characterization of isolates in study

256 Three different bacterial strains HC-degrading were isolated by sediment and seawater
257 samples in Priolo Bay as previously reported in Catania et al. (2015). The isolates were
258 identified by 16S rDNA analysis and were affiliated to genus *Thalassospira* (Alpha-
259 proteobacteria class), species *T. xianhensis*, *T. lucentensis* and *T. profundimaris*. The results of
260 phylogenetic identification are shown in **Fig.1**. The three bacterial strains formed an
261 independent branch related to other type strains HC degrading affiliated to other genus among
262 Alpha proteobacteria. In particular, The sequences showed considerable evolutionary
263 divergence from the sequences of *Aquaspirillum* sp., *Terasakiella* sp., Colonies were positive
264 for catalase and oxidase. Phenotypic and biochemical characteristic analyzed were showed in
265 **Table 1**.

266

267 3.2. *Detection of catabolic genes*

268 Positive PCR amplification with alkBwf-alkBwr, TS2S- deg1RE, and/or P450F-P450R
269 couple primers evidenced the presence of alkane-monoxygenase (*alkB*) and/or cytochrome
270 *P450* genes, while the biphenyl dioxygenase (*bphA*) didn't show amplification in the strains in
271 study. The presence /absence (positive/ negative amplification) of these genes is an indirect
272 index of potential capability of the strains in study of degradation of hydrocarbons, but in the
273 same time, it isn't express efficiency in degradation process.

274

275 3.3 *Biosurfactant production and emulsification activity*

276

277 The results obtained from the Surface Tension, drop collapse, oil spreading test and
278 Emulsification Activity (E24%)
279 have been reported in the **Table2**. The *thalassospira* spp. strains are producers of biosurfactants
280 and showed a good emulsification activity. Data obtained with drop collapse test identified
281 positive results for all isolates, but this contrast with data were obtained to other test. This
282 mismatch can be absolutely correlated to the different type and concentration of biosurfactant
283 produced and / or to the different experimental conditions (Youssef et al., 2004; Rodrigues et
284 al 2006).

285

286 3.4. GC×GC-Q MS/FID Analyses

287 As shown in **Figure 1**, the Arabian Light Crude Oil represents a very complex sample
288 containing different chemical classes. To achieve a reliable identification and quantification of
289 the different chemical classes the separation power of a mono-dimensional GC system is not
290 enough. For such a reason, in the present research, a GC × GC was employed due to its higher
291 peak capacity and the capability to couple columns with different selectivity. Moreover, in such
292 a kind of samples, the use of an MS detector is mandatory for reliable identification. However,
293 the drawback of this detector is the different response according to the analyte, do not allowing
294 quantification in case of lack of standard materials. For this purpose, FID is the only detector
295 providing virtually the same response according to the chemical class (Zoccali et al., 2016).
296 Thus, in the present paper, we developed a GC × GC with a dual detection system able to
297 perform both identification and quantification of the different chemical classes in a single run.

298 The goal of the GC × GC method optimization is to occupy as much as possible of the
299 two-dimensional space (without occurring in the wrap around effect) by achieving the best
300 separation possible and an ordered chromatogram. A medium polar column (35%) was used in
301 the first dimension and a non polar column (5%) was used in the second dimension, with such

302 a kind of configuration the non-polar components will occupy the upper part of the plot, while
303 the more polar ones the lower region. Compounds identification were carried out according to
304 MS spectral similarity, linear retention indices, and 2D position. In **Figure 1**, reporting the
305 Arabian Light Crude Oil analysis, the upper part of the 2D space was occupied by linear and
306 branched alkanes forming a fairly horizontal row of spots, and cyclic alkanes are positioned
307 below the row of the *n*-alkanes. As can be seen in **Figure 1** the retention in the second dimension
308 decrease according to the increase of the aromatic ring number, while increase in the first
309 dimension also considering the length of the alkyl chain. The use of the developed GC×GC-Q
310 MS/FID method, has allowed the quantification of the residual HC after incubation of the
311 different bacterial *Thalassospira* isolates. Totally, 6 different chemical classes were evaluated
312 namely: *i*) linear and branched alkanes, *ii*) cyclic alkanes, *iii*) C6-aromatics, *iv*) di-aromatics, *v*)
313 tri-aromatics + sulphur compounds and *vi*) tetra-aromatics + sulphur compounds. All the
314 analysis were carried out by using heptamethylnonane as internal standard and the values were
315 calculated considering the GC×GC-FID response. The obtained results were reported in the
316 histogram in **Figure 2**, expressed as percentage of degradation. Considering the total
317 contamination interesting result were obtained for all the studied bacterial strains, in fact the
318 degradation ranged of Arabian Light Crude Oil was between 77-91%. A important results was
319 obtained in the microcosm inoculated with the strain affiliated to *T. xianhensis* that gives the
320 best results with a % degradation ranging between 87% for the Cyclic alkanes and 94% for
321 Tetra-aromatics + Sulphur compounds. These compounds, in particular, Tetra-aromatics, are
322 recalcitrant, toxic, persistent and this result could be used for future study.

323

324 **4. Discussion**

325 Increasingly of environment pollution had involved scientific community to extend the
326 research in this sector and in particular in bioremediation and biorecovery strategies. This

327 “new” technology exploits microorganisms’ catabolism; indeed this last uses the pollutant
328 compound as carbon source and energy. Various studies demonstrated as this microbial process
329 converts pollutant compounds in atoxic or less toxic products, and as the bioremediation
330 technologies are one of benefit strategies to recovery polluted environment. Low cost and high
331 efficiency of bioremediation technology identify this strategy “*green technology*” and “eco-
332 friendly” (Nicolopoulou et al, 2013, Crisafi et al., 2016). Several data about different kind of
333 microorganisms with hydrocarbons degrading capability are present and described a lot of in
334 literature (Yakimov 2004; Rooling et al., 2004; Cappello et al., 2007; Catania et al. 2015;
335 Bargiela et al. 2015). The three stains isolated in this study were affiliated to the class of
336 *Alphaproteobacteria* and precisely to *Thalassospira* genus, particularly: isolates AU5AG4,
337 AU3AA1 and AU6SG1 were correlated respectively to *T. xianhensis* (KT348286), *T.*
338 *lucentensis* (KT348281) and *T. profundimaris* (KT348287). The focus of this work has been
339 the evaluation to features of this three species correlated with hydrocarbon degradation
340 capability, in particular were carried out test of biosurfactant production, presence of specific
341 catabolic genes involved in biodegradation processes (*alkB*, *CYP450* and *bphA* genes) and
342 biodegradation test in microcosms and analysis of the hydrocarbons fractions degraded. Several
343 test of biosurfactant production are detected, because these molecules have different properties,
344 besides low concentrations of this molecules could give incorrect values (Mahjoubi et al 2013).
345 Results of drop collapse and oil spreading test showed that all isolates in study are biosurfactant
346 producers; however, divergent data are obtained, in particular data of superficial tension put in
347 evidence that the isolate AU5AG4 *T. xiahnensis* is the best producer of biosurfactant, while
348 Emulsification activity (E24%) showed good results for AU6SG1 *T. profundimaris*. Zhan et al.
349 (2017) in their study isolated from *T. profundimaris* a siderophore affiliated to Thalassomide
350 family. These results could be dependent by the nature of biosurfactant, by the concentration of
351 this molecule and to the test applied. In the study of Youssef et al. (2004) the assessment of

352 surface tension is time- depending and showed a correlation coefficient stronger than drop
353 collapse and oil spreading test, which are indicated as more reliable detector of low
354 concentration of biosurfactant. Zhang et al. (2017), in their study, reported that this widespread
355 genus produces a siderophore affiliated to thalassosamide family, that is efficacy against
356 *Pseudomonas aeruginosa*. Results about the analysis of catabolic genes (alkano-monoxygenase
357 and citochrome P450) confirmed as this three bacterial species present in their metabolism
358 hydrocarbons degrading capability. However, the rates of biodegradation of hydrocarbons
359 substrate in the same isolates are different, in particular isolate AU5AG4 *T. xiahnensis* resulted
360 to have better biodegradation capability than others two isolates in study. Analysis of Oil
361 degradation evidence as the strains in study are capable both degrade linear and branches
362 alkanes, both fractions of molecules more complex and toxic as Cyclic alkanes and Di-, Tri-
363 and Tetra-aromatics compounds. The molecular structure of this kind of pollutants is usually
364 difficult to covert in molecular less toxic and recalcitrant, the strains of *Thalassospira* applied
365 in this study had able to transform and biodegrade this pollutants in few time. Recovery of oil
366 polluted areas can be achieved by physic-chemical or biological methods. Due to negative
367 aspect of the physic-chemical approach (Chibueze Azubuike et al., 2016), more attention is
368 being given to the exploitation of biological alternatives (Okoh, 2006). Biological treatments
369 are becoming more important, mainly due of the low environmental impact, costs (in general
370 cheaper than other cleanup technologies), the capability to destroy/degrade organic
371 contaminants and the possibility of advantageous use of treated sediments (Rulkens and
372 Bruning, 2005). Several studies have shown better results using bioremediation strategies
373 (Beolchini et al., 2010; Rocchetti et al., 2011, 2012; Cappello et al., 2015). In general,
374 bioremediation is often based on *in situ* stimulation of the microbial community
375 (biostimulation) or amending the microbial community with an inoculum of hydrocarbon-
376 degrading bacteria (bioaugmentation). In both cases, the successful result of bioremediation

377 depends on appropriate hydrocarbon-degrading consortia and environmental conditions. In this
378 study, interesting results suggest as the isolates of *Thalassospira* species are important and
379 relevant not only in the next studies of biotechnological processes, as the production and the
380 extraction of significative biomolecules used in several industries including organic chemicals,
381 petroleum, petrochemicals, mining, metallurgy (mainly bioleaching), agrochemicals,
382 fertilizers, foods, beverages, cosmetics, pharmaceuticals and many others (Vijayakumar S. and
383 Saravanan V. 2015), but also in the understanding of the bioremediation processes and in the
384 application of the same microorganisms in bioremediation techniques (Scaffaro, et al., 2017,
385 Catania et al., 2020).

386

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397

398 **Conflict of Interest**

399 No conflict of interest is declared by the authors and none of the authors of this paper has a
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401

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403 **References**

404

405 Ajay Singh, O.P.W., 2004. Biodegradation and Bioremediation. Springer-Verlag Berlin
406 Heidelberg.

407 Alkhatib, M.a.F., Alam, Z., Muyibi, S.A., Husain, A., 2011. An isolated bacterial consortium
408 for crude oil biodegradation. African Journal of Biotechnology 10, 18763-18767.

409 Bai, Z., Chen, H., Li, B., Li, W., 2005. Catalytic decomposition of methane over activated
410 carbon. Journal of analytical and applied pyrolysis 73, 335-341.

411 Baldani, J.I., Videira, S.S., dos Santos Teixeira, K.R., Reis, V.M., de Oliveira, A.L.M., Schwab,
412 S., de Souza, E.M., Pedraza, R.O., Baldani, V.L.D., Hartmann, A., 2014. The family
413 *rhodospirillaceae*, The Prokaryotes. Springer, pp. 533-618.

414 Beolchini, F., Rocchetti, L., Regoli, F., Dell'Anno, A., 2010. Bioremediation of marine
415 sediments contaminated by hydrocarbons: experimental analysis and kinetic modeling.
416 Journal of hazardous materials 182, 403-407.

417 Cappello, S., Calogero, R., Santisi, S., Genovese, M., Denaro, R., Genovese, L., Giuliano, L.,
418 Mancini, G., Yakimov, M.M., 2015. Bioremediation of oil polluted marine sediments: A
419 bio-engineering treatment. International microbiology : the official journal of the Spanish
420 Society for Microbiology 18, 127-134.

421 Cappello, S., Caruso, G., Zampino, D., Monticelli, L., Maimone, G., Denaro, R., Tripodo, B.,
422 Troussellier, M., Yakimov, M., Giuliano, L., 2007. Microbial community dynamics
423 during assays of harbour oil spill bioremediation: a microscale simulation study. Journal
424 of Applied Microbiology 102, 184-194.

425 Catania, V., Santisi, S., Signa, G., Vizzini, S., Mazzola, A., Cappello, S., ... & Quatrini, P.
426 (2015). Intrinsic bioremediation potential of a chronically polluted marine coastal area.
427 Marine pollution bulletin, 99(1-2), 138-149.

428 Catania, V., Sarà, G., Settanni, L., & Quatrini, P. (2017). Bacterial communities in sediment of
429 a Mediterranean marine protected area. Canadian Journal of Microbiology, 63(4), 303-
430 311.

431 Catania, V., Cappello, S., Di Giorgi, V., Santisi, S., Di Maria, R., Mazzola, A., ... & Quatrini,
432 P. (2018). Microbial communities of polluted sub-surface marine sediments. Marine
433 pollution bulletin, 131, 396-406.

434 Catania, V., Diliberto, C. C., Cigna, V., & Quatrini, P. (2020a). Microbes and Persistent
435 Organic Pollutants in the Marine Environment. Water, Air, & Soil Pollution, 231(7), 1-
436 10.

437 Catania, V., Lopresti, F., Cappello, S., Scaffaro, R., & Quatrini, P. (2020b). Innovative,
438 ecofriendly biosorbent-biodegrading biofilms for bioremediation of oil-contaminated
439 water. New Biotechnology.

440 Cerniglia, C.E., 1993. Biodegradation of polycyclic aromatic hydrocarbons. Current opinion in
441 biotechnology 4, 331-338.

442 Chaillan, F., Le Flèche, A., Bury, E., Phantavong, Y.-h., Grimont, P., Saliot, A., Oudot, J., 2004.
443 Identification and biodegradation potential of tropical aerobic hydrocarbon-degrading
444 microorganisms. Research in microbiology 155, 587-595.

445 Chen, L., Fang, Y., Zhu, T., Gu, Q., Zhu, W., 2007. Gentisyl alcohol derivatives from the
446 marine-derived fungus *Penicillium terrestre*. Journal of natural products 71, 66-70.

447 Crisafi, F., Genovese, M., Smedile, F., Russo, D., Catalfamo, M., Yakimov, M., Giuliano, L.,
448 Denaro, R., 2016. Bioremediation technologies for polluted seawater sampled after an

449 oil-spill in Taranto Gulf (Italy): A comparison of biostimulation, bioaugmentation and
450 use of a washing agent in microcosm studies. *Marine pollution bulletin* 106, 119-126.

451 Djahnit, N., Chernai, S., Catania, V., Hamdi, B., China, B., Cappello, S., & Quatrini, P. (2019).
452 Isolation, characterization and determination of biotechnological potential of oil-
453 degrading bacteria from Algerian centre coast. *Journal of applied microbiology*, 126(3),
454 780-795.

455 Dudášová, H., Lukáčová, L., Murínová, S., Puškárová, A., Pangallo, D., Dercová, K., 2014.
456 Bacterial strains isolated from PCB-contaminated sediments and their use for
457 bioaugmentation strategy in microcosms. *J. Basic Microbiol.* 54, 253-260.

458 Dyksterhouse, S.E., Gray, J.P., Herwig, R.P., Lara, J.C., Staley, J.T., 1995. *Cycloclasticus*
459 *pugetii* gen. nov., sp. nov., an aromatic hydrocarbon-degrading bacterium from marine
460 sediments. *International Journal of Systematic and Evolutionary Microbiology* 45, 116-
461 123.

462 García-Falcón, M.S., Cancho-Grande, B., Simal-Gándara, J., 2005. Minimal clean-up and rapid
463 determination of polycyclic aromatic hydrocarbons in instant coffee. *Food chemistry* 90,
464 643-647.

465 Haritash, A., Kaushik, C., 2009. Biodegradation aspects of polycyclic aromatic hydrocarbons
466 (PAHs): a review. *Journal of hazardous materials* 169, 1-15.

467 Hassanshahian, M., Emtiazi, G., Cappello, S., 2012. Isolation and characterization of crude-oil-
468 degrading bacteria from the Persian Gulf and the Caspian Sea. *Marine pollution bulletin*
469 64, 7-12.

470 Hassanshahian, M., Emtiazi, G., Kermanshahi, R.K., Cappello, S., 2010. Comparison of oil
471 degrading microbial communities in sediments from the Persian Gulf and Caspian Sea.
472 *Soil and Sediment Contamination* 19, 277-291.

473 Holt, J.G., 1977. The shorter Bergey's manual of determinative bacteriology. The shorter
474 Bergey's manual of determinative bacteriology. 8th edition.

475 Innis, M.A., Gelfand, D.H., Sninsky, J.J., White, T.J., 1990. PCR Protocols: A Guide to
476 Methods and Applications. San Diego: Academic Press.

477 Iqbal, S., Khalid, Z., Malik, K., 1995. Enhanced biodegradation and emulsification of crude oil
478 and hyperproduction of biosurfactants by a gamma ray-induced mutant of *Pseudomonas*
479 *aeruginosa*. Letters in Applied Microbiology 21, 176-179.

480 Kanaly, R.A., Harayama, S., 2000. Biodegradation of high-molecular-weight polycyclic
481 aromatic hydrocarbons by bacteria. Journal of bacteriology 182, 2059-2067.

482 Kasai, Y., Kishira, H., Harayama, S., 2002. Bacteria belonging to the genus *Cycloclasticus* play
483 a primary role in the degradation of aromatic hydrocarbons released in a marine
484 environment. Applied and environmental microbiology 68, 5625-5633.

485 Kiseleva, L., Garushyants, S.K., Briliute, J., Simpson, D.J., Cohen, M.F., Goryanin, I., 2015.
486 Genome sequence of the electrogenic petroleum-degrading *Thalassospira* sp. strain HJ.
487 Genome announcements 3, e00483-00415.

488 Kloos, K., Munch, J.C., Schloter, M., 2006. A new method for the detection of alkane-
489 monooxygenase homologous genes (alkB) in soils based on PCR-hybridization. Journal
490 of Microbiological Methods 66, 486-496.

491 Kodama, Y., Stiknowati, L.I., Ueki, A., Ueki, K., Watanabe, K., 2008. *Thalassospira*
492 *tepidiphila* sp. nov., a polycyclic aromatic hydrocarbon-degrading bacterium isolated
493 from seawater. International journal of systematic and evolutionary microbiology 58,
494 711-715.

495 Kubota, A., Iwata, H., Tanabe, S., Yoneda, K., Tobata, S., 2005. Hepatic CYP1A induction by
496 dioxin-like compounds, and congener-specific metabolism and sequestration in wild

497 common cormorants from Lake Biwa, Japan. *Environmental science & technology* 39,
498 3611-3619.

499 Le Borgne, S., Paniagua, D., Vazquez-Duhalt, R., 2008. Biodegradation of organic pollutants
500 by halophilic bacteria and archaea. *Journal of molecular microbiology and biotechnology*
501 15, 74-92.

502 López-López, A., Pujalte, M.J., Benlloch, S., Mata-Roig, M., Rosselló-Mora, R., Garay, E.,
503 Rodríguez-Valera, F., 2002. *Thalassospira lucentensis* gen. nov., sp. nov., a new marine
504 member of the alpha-Proteobacteria. *International journal of systematic and evolutionary*
505 *microbiology* 52, 1277-1283.

506 Mahjoubi, M., Jaouani, A., Guesmi, A., Ben Amor, S., Jouini, A., Cherif, H., Najjari, A.,
507 Boudabous, A., Koubaa, N., Cherif, A., 2013. Hydrocarbonoclastic bacteria isolated from
508 petroleum contaminated sites in Tunisia: isolation, identification and characterization of
509 the biotechnological potential. *New biotechnology* 30, 723-733.

510 Margesin, R., Schinner, F., 2001. Biodegradation and bioremediation of hydrocarbons in
511 extreme environments. *Applied microbiology and biotechnology* 56, 650-663.

512 McKew, B.A., Coulon, F., Yakimov, M.M., Denaro, R., Genovese, M., Smith, C.J., Osborn,
513 A.M., Timmis, K.N., McGenity, T.J., 2007. Efficacy of intervention strategies for
514 bioremediation of crude oil in marine systems and effects on indigenous
515 hydrocarbonoclastic bacteria. *Environmental microbiology* 9, 1562-1571.

516 Melcher, R.J., Apitz, S.E., Hemmingsen, B.B., 2002. Impact of irradiation and polycyclic
517 aromatic hydrocarbon spiking on microbial populations in marine sediment for future
518 aging and biodegradability studies. *Applied and environmental microbiology* 68, 2858-
519 2868.

520 Muangchinda, C., Yamazoe, A., Polrit, D., Thoetkiattikul, H., Mhuantong, W., Champreda, V.,
521 Pinyakong, O., 2017. Biodegradation of high concentrations of mixed polycyclic

522 aromatic hydrocarbons by indigenous bacteria from a river sediment: a microcosm study
523 and bacterial community analysis. *Environmental science and pollution research*
524 *international* 24, 4591-4602.

525 Nicolopoulou-Stamati, P., Hens, L., Howard, V.C., 2013. Health Impacts of Waste
526 Management Policies: Proceedings of the Seminar 'Health Impacts of Waste Management
527 Policies' Hippocrates Foundation, Kos, Greece, 12–14 November 1998. Springer Science
528 & Business Media.

529 Nikolopoulou, M., Kalogerakis, N., 2010. Biostimulation strategies for enhanced
530 bioremediation of marine oil spills including chronic pollution, *Handbook of hydrocarbon*
531 *and lipid microbiology*. Springer, pp. 2521-2529.

532 Okoh, A.I., 2006. Biodegradation alternative in the cleanup of petroleum hydrocarbon
533 pollutants. *Biotechnology and Molecular Biology Reviews* 1, 38-50.

534 Peng, R.-H., Xiong, A.-S., Xue, Y., Fu, X.-Y., Gao, F., Zhao, W., Tian, Y.-S., Yao, Q.-H.,
535 2008. Microbial biodegradation of polyaromatic hydrocarbons. *FEMS microbiology*
536 *reviews* 32, 927-955.

537 Ryslava, E., Krejčík, Z., Macek, T., Novakova, H., Denmerova, K., and Mackova, M., 2003.
538 Study of PCB biodegradation in real contaminated soil. *Fresh Environ. Bull.*, 12, 296-
539 301.

540 Rocchetti, L., Beolchini, F., Ciani, M., Dell'Anno, A., 2011. Improvement of bioremediation
541 performance for the degradation of petroleum hydrocarbons in contaminated sediments.
542 *Applied and Environmental Soil Science* 2011.

543 Rodrigues, J.d.A., Cardoso, F.d.P., Lachter, E.R., Estevão, L.R., Lima, E., Nascimento, R.S.,
544 2006. Correlating chemical structure and physical properties of vegetable oil esters.
545 *Journal of the American Oil Chemists' Society* 83, 353-357.

546 Röling, W.F., Milner, M.G., Jones, D.M., Fratepietro, F., Swannell, R.P., Daniel, F., Head,
547 I.M., 2004. Bacterial community dynamics and hydrocarbon degradation during a field-
548 scale evaluation of bioremediation on a mudflat beach contaminated with buried oil.
549 *Applied and environmental microbiology* 70, 2603-2613.

550 Rulkens, W.H., Bruning, H., 2005. Clean-up Technologies for Dredged Fine Sediments Review
551 and Future Challenge.

552 Scaffaro, R., Lopresti, F., Catania, V., Santisi, S., Cappello, S., Botta, L., & Quatrini, P. (2017).
553 Polycaprolactone-based scaffold for oil-selective sorption and improvement of bacteria
554 activity for bioremediation of polluted water: Porous PCL system obtained by leaching
555 melt mixed PCL/PEG/NaCl composites: Oil uptake performance and bioremediation
556 efficiency. *European Polymer Journal*, 91, 260-273.

557 Smits, T.H.M., Röthlisberger, M., Witholt, B., van Beilen, J., B., 1999. Molecular screening for
558 alkane hydroxylase genes in gram-negative and gram-positive strains. *Environ.*
559 *Microbiol.*, 1, 307-317.

560 SUTHERLAND, J.B., 1995. Mechanisms of polycyclic aromatic hydrocarbon degradation.
561 *Microbial transformation and degradation of toxic organic chemicals*, 269-306.

562 Thompson, H., Angelova, A., Bowler, B., Jones, M., Gutierrez, T., 2017. Enhanced crude oil
563 biodegradative potential of natural phytoplankton-associated hydrocarbonoclastic
564 bacteria. *Environmental microbiology* 19, 2843-2861.

565 Tranchida, P.Q., Franchina, F.A., Zoccali, M. Bonaccorsi, I., Cacciola, F., Mondello, L.,
566 2013. A direct sensitivity comparison between flow-modulated comprehensive 2D and
567 1D GC in untargeted and targeted MS-based experiments. *Journal of Separation Science*
568 36, 2746-2752.

569 Tranchida, P.Q., Purcaro, G., Conte, L., Dugo, P., Dugo, G., Mondello, L., 2009. Optimized
570 use of a 50 µm internal diameter secondary column in a comprehensive two-dimensional
571 gas chromatography system. *Analytical chemistry* 81, 8529-8537.

572 Wang, L., Wang, W., Lai, Q., Shao, Z., 2010. Gene diversity of CYP153A and AlkB alkane
573 hydroxylases in oil-degrading bacteria isolated from the Atlantic Ocean. *Environmental*
574 *microbiology* 12, 1230-1242.

575 Wong, M., 2003. Ecological restoration of mine degraded soils, with emphasis on metal
576 contaminated soils. *Chemosphere* 50, 775-780.

577 Yakimov, M.M., Cappello, S., Crisafi, E., Tursi, A., Savini, A., Corselli, C., Scarfi, S., Giuliano,
578 L., 2006. Phylogenetic survey of metabolically active microbial communities associated
579 with the deep-sea coral *Lophelia pertusa* from the Apulian plateau, Central Mediterranean
580 Sea. *Deep Sea Research Part I: Oceanographic Research Papers* 53, 62-75.

581 Yakimov, M.M., Denaro, R., Genovese, M., Cappello, S., D'Auria, G., Chernikova, T.N.,
582 Timmis, K.N., Golyshin, P.N., Giuliano, L., 2005. Natural microbial diversity in
583 superficial sediments of Milazzo Harbor (Sicily) and community successions during
584 microcosm enrichment with various hydrocarbons. *Environmental microbiology* 7, 1426-
585 1441.

586 Yakimov, M.M., Giuliano, L., Denaro, R., Crisafi, E., Chernikova, T.N., Abraham, W.-R.,
587 Luensdorf, H., Timmis, K.N., Golyshin, P.N., 2004. *Thalassolituus oleivorans* gen. nov.,
588 sp. nov., a novel marine bacterium that obligately utilizes hydrocarbons. *International*
589 *journal of systematic and evolutionary microbiology* 54, 141-148.

590 Yakimov, M.M., Timmis, K.N., Golyshin, P.N., 2007. Obligate oil-degrading marine bacteria.
591 *Current opinion in biotechnology* 18, 257-266.

592 Youssef, N.H., Duncan, K.E., Nagle, D.P., Savage, K.N., Knapp, R.M., McInerney, M.J., 2004.
593 Comparison of methods to detect biosurfactant production by diverse microorganisms.
594 Journal of microbiological methods 56, 339-347.

595 Zhao, B., Wang, H., Li, R., Mao, X., 2010. *Thalassospira xianhensis* sp. nov., a polycyclic
596 aromatic hydrocarbon-degrading marine bacterium. International journal of systematic
597 and evolutionary microbiology 60, 1125-1129.

598 Zoccali, M., Barp, L., Beccaria, M., Sciarrone, D., Purcaro, G., Mondello, L., 2016.
599 Improvement of mineral oil saturated and aromatic hydrocarbons determination in edible
600 oil by liquid-liquid-gas chromatography with dual detection. Journal of Separation
601 Science 39, 623-631.

602 Zoccali, M., Bonaccorsi, I.L., Tranchida, P.Q., Dugo, P., Mondello, L., Dugo, G.,
603 2015. Analysis of the sesquiterpene fraction of citrus essential oils by using the off-line
604 combination of high performance liquid chromatography and gas chromatography-based
605 methods: A comparative study. Flavour and Fragrance Journal 30, 411-422.

606 Zoccali, M., Cappello, S., Mondello, L., 2018. Multilevel characterization of marine microbial
607 biodegradation potentiality by means of flow-modulated comprehensive two-dimensional
608 gas chromatography combined with a triple quadrupole mass spectrometer. Journal of
609 Chromatography A 1547, 99-106.

610

611 **Tables**

612

613 **Tab. 1.** Closest relatives of the 16S rRNA gene sequences of bacteria isolated in this study and
614 results of the main classical biochemical.

615

616

Isolate code	Closest type strain	Identity (%)	Oxidase test	Catalase test	Mobility	Shape
AU5AG4	<i>T. lucentensis</i>	98	Positive	Positive	Yes	Curved-rod
AU3AA1	<i>T. profundimaris</i>	99	Positive	Positive	Yes	Curved-rods to spirilla
AU6SG1	<i>T. xianhensis</i>	98	Positive	Positive	no	Curved-rod

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Tab. 2. Superficial Tension, Drop Collapse test, Oil spreading test and Emulsification Activity of strains in study. Positive data of Drop Collapse and Oil spreading test are indicated by the symbol “+”, in contrast negative data are indicated with the symbol “-”.

Isolate code	Closest type strain	Superficial tension (mN m-1)	Drop collapse test	Oil spreading test	E24 test (%)
AU5AG4	<i>T. lucentensis</i>	49.7 ±0.5	+	+	22.2 ±0.5
AU3AA1	<i>T. profundimaris</i>	53.8 ±0.5	+	+	21.9 ±0.5
AU6SG1	<i>T. xianhensis</i>	53.2 ±0.5	+	+	11.1 ±0.5

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627

628 **Figure Captures**

629

630 **Fig. 1.** Phylogenetic tree based on 16S rRNA gene sequence analysis of isolates in study
631 (AU3AA1, AU5SG4 and AU6SG1) and relationship with different species of the genus
632 *Thalassospira*, some genera related to the class of Alpha-proteobacteria (*Terasakiella* and
633 *Aquaspirillum*), and others hydrocarbons degrading bacteria affiliated to the class of Gamma-
634 proteobacteria. As Outgroup was used the 16S rRNA gene of *Methanococcus jannaschii*
635 (M59126).

636

637 **Fig. 2.** GC×GC-Q MS analysis of Arabian Light Crude Oil. Abbreviations: B, benzene; N,
638 naphthalene; F, fluorene; DBT, dibenzothiophene; P, phenanthrene; Py, pyrene; BNT,
639 benzonaphthothiophene; Cry, chrysene; Per, perylene. The abbreviations C1, C2, etc., refer to
640 the number of carbons in the alkyl substituents. The intensity of the peaks is related to color
641 (white minimum, red maximum).

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643

644 **Fig. 3.** Rate (%) of biodegradation (calculated by using the GC×GC-FID response) of the
645 different bacterial strains studied in the present work: isolate AU6SG1 (*T. profundimaris*; white
646 bars), isolate AU3AA1 (*T. lucentensis*; light gray bars) and isolate AU5AG4 (*T. xianhensis*;
647 dark gray bars).

648