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Zirconium and hafnium fractionation and distribution of rare earth elements in neutral - alkaline waters: Case study of Lake Van hydrothermal system, Turkey --Manuscript Draft--

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Abstract:	<p>We investigated the distribution of Zr, Hf, and rare earth elements (REE) as the sum of lanthanides plus Y in the hydrothermal system in the Lake Van area of south-eastern Turkey. This system is characterised by water with variable pH in alkaline conditions resulting from hydrothermal CO₂ upraise and neoformation of calcite minerals in near equilibrium with the interacting waters. Zr, Hf, and REE determinations were carried out for aqueous phases and suspended solids in lake water and surrounding thermal springs. We found that dissolved Hf is partitioned relative to Zr during calcite formation and that such fractionation is a function of the Ca²⁺ activity in warm water. The observed Zr-Hf fractionation is explained by coulombic interactions that occur between suspended solid particles and dissolved phases at the calcite-water interface. There, the surfaces of carbonate minerals demonstrated greater reactivity towards aqueous Hf-bearing species relative to Zr-complexes. This evidence involves a coulombic mechanism of reactivity at the calcite-water interface because Hf complexes are negatively charged while Zr compounds are uncharged. Thus, authigenic calcite can behave as a suitable host for dissolved metal ion species to adsorb on crystal surfaces to remediate waste waters from mine drainage.</p>
Suggested Reviewers:	Thomas H Darrah darrah.24@osu.edu Geochemist with a large experience in REE Rosa Cidu cidur@unica.it Johan Varekamp jvarekamp@wesleyan.edu lake geochemist Yigal Erel yerel@vms.huji.ac.il Trace element geochemist with a large experience in solid-water interface reactions
Response to Reviewers:	Response to Reviewer 1 Line 41 – the reference was added Line 73 – Yes, this problem was considered but sampling conditions in the studied areas and the limited time elapsed between the sample collection and filtration allowed us to filter waters under clean lab conditions.

Line 131 – only two significant digits were reported in Eh values.
Line 193 – allo scopo di riportare le concentrazioni disciolte sempre in pmol/L alcuni valori maggiori di 1000 pmol/L erano stati riportati con la virgola (1,000) per dividere le migliaia dalle altre cifre intere. Per evitare confusione nel lettore questa scelta è stata evitata.

Response to Reviewer 2

A brief geological summary of the lake Van area was added.

Line 31 – done

Line 37 – two references were added.

Line 45 – the text was corrected

Line 97 – w/w concentrations of HNO₃, NH₄OH and HCl solutions and the mark ULTREX II of Baker chemicals were added.

Line 140-142 – data are in weight %. They were assessed with the Rietveld method as reported at lines 115-120.

Line 198 – the correct text is “without evidence of a positive Ce anomaly”. Positive was lost. Sorry!

Line 241 –in lines 243-247, we list several minerals (Fe-oxyhydroxides, Phosphates, Salt minerals, sulphates) that can fractionate MREE relative to other REE. It is a large and heterogeneous group of minerals where it is hard to identify a possible MREE source during rock-water interactions. Probably MREE fractionation in a so large group of minerals is a process ruled by crystal-chemical reasons since elements from Sm to Dy show similar ionic radii than Ca²⁺ and Na⁺. Then, the text is limited to suggest possible minerals representing reasonable MREE sources.

Fig. 3. It was modified

Fig. 4. It was modified

Fig. 5. pCa = -Log[Ca²⁺]. [Ca²⁺] is expressed in mol/L, as reported in the caption.

Table 1. The number of digits was modified.

Table 2. No. Those reported in Table 2 is the mineralogical composition expressed in weight % units. These are measured with the TOPAS software carrying out on X-ray spectra according to the Rietveld method (see Materials and Methods chapter).
Supplementary tables have been corrected as suggested by the reviewer.

1 Zirconium and hafnium fractionation and distribution of Rare Earth Elements in neutral–alkaline
2 waters: Case study of Lake Van hydrothermal system, Turkey

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12

13 Abstract

14 We investigated the distribution of Zr, Hf, and rare earth elements (REE) as the sum of lanthanides
15 plus Y) in the hydrothermal system in the Lake Van area of south-eastern Turkey. This system is
16 characterised by water with variable pH in alkaline conditions resulting from hydrothermal CO₂
17 upraise and neof ormation of calcite minerals in near equilibrium with the interacting waters. Zr, Hf,
18 and REE determinations were carried out for aqueous phases and suspended solids in lake water and
19 surrounding thermal springs. We found that dissolved Hf is partitioned relative to Zr during calcite
20 formation and that such fractionation is a function of the Ca²⁺ activity in warm water. The observed
21 Zr-Hf fractionation is explained by coulombic interactions that occur between suspended solid
22 particles and dissolved phases at the calcite-water interface. There, the surfaces of carbonate
23 minerals demonstrated greater reactivity towards aqueous Hf-bearing species relative to Zr-
24 complexes. This evidence involves a coulombic mechanism of reactivity at the calcite-water
25 interface because Hf complexes are negatively charged while Zr compounds are uncharged. Thus,

26 authigenic calcite can behave as a suitable host for dissolved metal ion species to adsorb on crystal
27 surfaces to remediate waste waters from mine drainage.
28
29 **Keywords**
30 Zr/Hf ratio; REE; Alkaline lakes; Turkey

31 **1 Introduction**

32 Zirconium, Hf, and [rare-earth elements \(REE\)](#) ~~REEs~~ are used in specialised technology
33 industries notwithstanding their limited production (Chakhmouradian and Wall, 2012; Jones III et
34 al., 2017). Demand for these elements is high leading an effort by the mineral exploration industry to
35 understand genetic processes and discover new resources. Geochemically, these elements were
36 considered to be 'immobile' in subsurface conditions while some works indicate easy water
37 transport under subsurface hydrothermal conditions– [\(Michard and Albarede, 1986; Michard, 1989\)](#).
38 Accordingly, REE enrichment has been observed in deposits formed from carbonisation in alkaline
39 environments where Zr concentrations were enriched by a factor of 3 to 4 (Estrade et al., 2015). The
40 behaviour of these strategic elements under surface and subsurface conditions is often neglected
41 because of difficulties in accurate analytical determination at the level of pmol/kg in natural water.
42 In a recent survey, the dissolved Zr/Hf ratio was different from that found in the rock minerals
43 (Censi et al., 2020; [Cidu et al., 2013](#); Inguaggiato et al., 2015, 2016; Parisi et al., 2017; Tepe and
44 Bau, 2015; Zuddas et al., 2017, 2018) suggesting that the subsurface cycle is important in the
45 accumulation of these elements. In fact, the Zr/Hf molar value in rock minerals was 71.4 ± 5.6 which
46 corresponds to the so-called 'chondritic' signature (Jochum et al., 1986), while natural water had a
47 wider range of Zr/Hf (Zuddas et al., 2018). The supergene geochemical cycle of Zr and Hf
48 participates in the accumulation and stability of these elements through the process of dissolution
49 and precipitation of the minerals. Because the composition of subsurface and surface waters is often
50 near equilibrium with carbonate minerals like calcite, calcite is expected to play a significant role in
51 the supergene behaviour of these elements. Although carbonates are among the most common
52 minerals in equilibrium with natural water, the formation of neogenic calcite also determined the
53 change in REE distribution observed by Moller and De Lucia (2020) in controlled laboratory
54 investigations. To explain Zr-Hf decoupling during calcite crystallisation and the concurrent
55 behaviour of REEs, we investigated the distribution of these elements in a natural hydrothermal
56 system characterised by the formation of calcite under the neutral and alkaline water system found at

57 Lake Van and characterised by low thermal and lake waters with pH ranging from 6 to 10, resulting
58 from a variable CO₂ flux (Degens and Kurtman, 1978; Moller and Bau, 1993). The Lake Van system
59 is characterised by neogenic precipitation in the form of Ca-Mg carbonates (Moller and Bau, 1993).

60 **2 Material and Methods**

61 Lake Van is located in south-eastern Turkey (Fig. 1) which formed during the collision
62 between European and Arabian plates. The sequence starts with Paleozoic metamorphic rocks
63 overlaid by Permian limestones and schists followed by sediments of marine and continental origin
64 (Upper Cretaceous-Pliocene) closed by basic and felsic volcanic-pyroclastic rocks (Oligocene-
65 Recent) (Degens and Kurtman, 1978, Yilmaz, 1990). Heat flow data suggest the presence of a
66 magmatic body beneath the lake bottom, possibly related to Nemrut Volcano and sustaining input of
67 CO₂ into Lake Van waters (Moller and Bau, 1993). Presently, the Lake Van represents the largest
68 soda lake worldwide resulting from a combined volcanogenic hydrothermal CO₂ input and seasonal
69 evaporitic conditions (Moller and Bau, 1993).

70 **2.1 Field sampling**

71 Lake Van is located in south-eastern Turkey (Fig. 1) which formed during the collision
72 between European and Arabian plates. Field study and water sampling were conducted in November
73 2018 and consisted of a 16 samples collected from lake water and thermal springs. Coordinates and
74 geologic units of the collected samples are given in Table 1. Lake water was sampled at a depth of
75 15–20 cm. Physicochemical parameters (pH, redox potential (Eh), temperature, and electrical
76 conductivity) were measured with an Eh oxytrode Pt probe (Hamilton™) using a standard reference
77 solution buffer at 0.475±0.005 V~~475±5 mV~~. Accuracy of determinations was ±10 mV. For the major
78 ion determinations, water samples were filtered in the field through 0.45 µm Millipore® cellulose
79 acetate filters into 50 mL polyethylene bottles and acidified with ultrapure HNO₃. Water samples for
80 trace and rare element analyses were collected in 2,000 mL Nalgene bottles at each sampling site.

81 **2.2 Water analysis and thermodynamic modelling**

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82 In the laboratory, samples were filtered (Millipore® manifold filter, 47 mm diameter with a
 83 0.45 µm pore size cellulose nitrate membrane), then 5% ultrapure HNO₃ acid solution was added to
 84 reach pH ≈ 2. Zirconium, Hf, and REEs in aqueous samples were analysed according to the methods
 85 of Raso et al. (2013). Briefly, an excess of FeCl₃ (1%) solution was added to each sample (1 L), and
 86 a suitable volume of NH₄OH (25%) solution was added to attain a pH of 8 to induce precipitation of
 87 solid Fe(OH)₃. During this process, Zr, Hf, and REEs were held onto the surface of the crystallising
 88 solid. To ensure that crystallisation of Fe(OH)₃ was complete, the solution was left in a closed flask
 89 for 48–72 h in a stirrer. Next, the iron concentration was measured to assess recovery and was
 90 consistently greater than 95%. Precipitated Fe(OH)₃, together with Zr, Hf, and REEs, was collected
 91 on a membrane filter (Millipore® manifold filter, 47 mm diameter with 0.45 µm pore size). The
 92 solid filtrate was then dissolved in 3 M HCl, and the obtained solution was diluted to 1 M HCl prior
 93 to analysis using a Quadrupole-ICP-MS (Agilent 7800 series) with an external calibration procedure.
 94 The overall strategy and a detailed evaluation of the analytical errors were reported in Raso et al.
 95 (2013). The overall enrichment factor measured for Zr, Hf, and REE concentrations corresponded to
 96 66.7.

97 Lanthanum, Ce, and Eu anomalies were calculated using equations 1–3 (Alibo and Nozaki,
 98 1999):

$$99 \quad \frac{La}{La^*} = \frac{[La]_n}{3*[Pr]_n + 2*[Nd]_n}, \quad (1)$$

$$100 \quad \frac{Ce}{Ce^*} = \frac{2*[Ce]_n}{[La]_n + [Pr]_n}, \text{ and} \quad (2)$$

$$101 \quad \frac{Eu}{Eu^*} = \frac{2*[Eu]_n}{[Sm]_n + [Gd]_n}, \quad (3)$$

102 where the subscript n indicates shale-normalised elemental concentrations calculated relative to the
 103 Post Archean Australian Shale (PAAS, Taylor and McLennan, 1995).

104 All chemicals used during laboratory manipulations were ultrapure grade. Ultrapure water
 105 (resistivity of 18.2 MΩ cm or better) was obtained from an Arium® mini system (Sartorius, Italy).

106 Nitric acid 65% (w/w), NH₄OH (20%) and HCl (30%) solutions ~~ammonia solution, and HCl acid~~

107 were purchased from J.T. Baker ([ULTREX II](#)) chemicals. Working standard solutions of the
108 elements studied were prepared daily by stepwise dilution of standard multi-element stock solutions
109 from DBH, Merck, or CPI International (1,000±5 mg/L) in a 1 M HCl medium. All labware was
110 polyethylene, polypropylene, or Teflon® and the calibration of all volumetric equipment was verified.
111 Assessment of the analytical precision of Zr, Hf, and REEs was performed on five aliquots (500 mL
112 each) of NASS-6 (seawater certified reference material for trace metals) distributed by the National
113 Research Council of Canada. These certified reference materials were treated as water samples
114 according to the procedure mentioned above and the concentrations obtained were compared with
115 those previously reported by Raso et al. (2013) and Lemaitre et al. (2014). Analysed values are
116 reported in Supplementary Data 1.

117 Thermodynamic modelling of the chemical composition of the water performed was using
118 PHREEQC software and adding additional Lawrence Livermore National Laboratory (LLNL)
119 database (version 3.0.6, Parkhurst and Appelo, 2010).

120 **2.3 Suspended solids**

121 X-ray diffraction investigations were carried out with a Philips PW14 1373 X-ray
122 spectrometer (Cu-K α radiation 2 θ range 10–40°, step size 0.02°, and a 1-min step time). The X-ray
123 spectra were analysed by the Rietveld method (Program DiffracPlus TOPAS®, Version 4.0, Bruker
124 AXS Inc., Karlsruhe, Germany) using parameters for the Rietveld refinement method obtained from
125 the Inorganic Crystal Structure Database (ICSD) program. This method consists of fitting the
126 experimental XRD spectrum to the theoretical spectrum (Young, 1993). [A detailed explanation of
127 the method is reported in Scarlett and Madsen \(2006\).](#)

128 Scanning electron microscopy (SEM) of suspended particles was carried out on materials collected
129 after filtration onto cellulose nitrate membranes. Dried solids were mounted on an aluminium stub
130 and gold coated. The SEM analyses were carried out using a LEO 440 SEM equipped with an EDS
131 system OXFORD ISIS Link and Si (Li) PENTAFET detector at the SIDERCEM SRL laboratory
132 (Caltanissetta, Italy). Chemical analyses were carried out by digesting 100 mg of each sample in 10

133 mL of a 1:1 HNO₃:H₂O₂ mixture in a sealed Teflon® TFM bomb using a microwave oven (CEM
134 MARS 5 device). After digestion, the TFM bomb was rinsed, and the solution was filtered onto
135 previously acid-cleaned 0.45 µm Millipore® filters to remove silicate residue. The filtered solution
136 was diluted to 50 mL and stored for chemical analyses.

137 **3 Results and Discussion**

138 **3.1 System description**

139 The concentration of major elements, pH, Eh, electric conductivity, and temperature are
140 reported in the Supplementary Data 2. Water samples had ionic strength ranging from 0.024 to 0.222
141 M in thermal water and between 0.004 and 0.443 M in lake water. Thermal water had pH between
142 5.8 and 7.1 and Eh between 0.013 and 0.172 V, while lake water had pH between 8 and 10 and Eh
143 from -0.2988 and 0.012 V. Figure 2a shows that lake water preferentially falls in the field of alkaline
144 elements (only two samples have higher Ca²⁺ and Mg²⁺ concentrations) while thermal water fell
145 between alkaline and Ca²⁺-rich field (few samples fell in the Mg²⁺ corner).

146 The SEM images of the suspended particulates revealed that suspended particulate matter in
147 lake water consisted of authigenic carbonates associated with biogenic detritus and lithic fragments
148 (Fig. 3a-f), while suspended solids in thermal waters were authigenic carbonates and Fe-rich
149 particles and lithogenic fragments (Fig. 3g-i). Iron-bearing particulates in the samples were often
150 mixed with authigenic calcite and dolomite (see elemental maps in Fig. 3i). The X-ray diffraction
151 analyses confirmed the presence of carbonate minerals identified as calcite and low-Mg calcite
152 representing up to 80% of the authigenic minerals (Table 2). Other phases collected were Fe-
153 oxyhydroxides and detritic silicate particles reaching 10–15% of the total mass.

154 Using the chemical composition of the collected samples, we estimated the saturation state
155 with respect to the mineral in equilibrium and found both lake water and thermal water were
156 saturated or oversaturated with respect to calcite or low-Mg calcite. Some water was also saturated
157 with respect to dolomite and goethite in agreement with the results of the X-ray determinations of
158 the collected suspended matter.

159 **3.2 Zr and Hf behaviour**

160 Zr and Hf concentrations in both aqueous and suspended particulate matter are reported in
161 Supplementary Data 3 and 4, respectively. In the aqueous phase, the concentration of Zr and Hf was
162 not homogenous. Lake waters had Zr concentrations between 79 and 4,958 pmol/L while Hf
163 concentrations fell between 2 and 53 pmol/L, and thermal waters had Zr concentrations between 32
164 and 1.425 pmol/L; whereas Hf concentrations fell between 0.5 and 14 pmol/L. The concentrations of
165 Zr and Hf in thermal waters were 1 to 3 orders of magnitude lower than those found in lake water;
166 however, the concentration of Zr as a function of Hf was linearly correlated with an intercept equal
167 to zero and parametrising a single slope of 90.4 ± 21.7 (Fig. 4a). This value was higher than the Zr/Hf
168 value of 71 ± 5.6 for crustal materials and chondrites (Jochum et al., 1986) indicating that the calcite
169 precipitation fractionates the Zr/Hf ratio in water. In the suspended solids from lake water, Zr
170 concentration ranged from 99.5 to $2,533.2$ nmol/kg, whereas Hf fell between 2.6 and 77.1 nmol/kg.
171 Suspended solids from thermal water had Zr concentrations between 20.7 and 346.7 nmol/kg and Hf
172 between 0.6 and 6.8 nmol/kg. Similarly, Zr and Hf in solids collected from thermal water were lower
173 compared to suspended solids found in lake water. Figure 4B shows the concentration of Zr as a
174 function of Hf in neogenic calcite from lake water and describes a linear slope of 33.5 similar to
175 calcite in equilibrium with thermal water where the slope was 37.

176 These results show that the Zr/Hf ratio in water is different from that in crustal rocks and that
177 Hf is preferentially enriched in authigenic carbonate minerals. Preferential enrichment of Hf relative
178 to Zr in newly-formed suspended carbonates was unexpected and may result from processes
179 occurring at the surface-water interface. The calcite-water interface is generally represented by an
180 electric double-layer of weak mineral solubility buffered by CO_3^{2-} ions with the ζ potential of the
181 {104} carbonate surface independent of pH (Herberling et al., 2011). Laboratory measurements
182 indicate the positive surface potential (up to 24 mV) when the Ca^{2+} concentration in solution is
183 between 10^{-2} and 10^{-4} M (Al-Mahrouqui et al., 2017). Byrne (2002) made a critical review of
184 stability constants for Zr and Hf hydroxyl complexes based on the work of Sillen and Martell (1964),

185 Baes and Mesmer (1976), Ekberg et al. (2004), and Brown et al. (2005) and estimated that the most
186 abundant Zr and Hf complexes in aqueous phase were $[\text{Zr}(\text{OH})_4]^0$ and $[\text{Hf}(\text{OH})_5]^-$, respectively
187 (Koschinsky and Hein, 2003). Reporting Zr/Hf ratios as a function of pCa:

$$pCa = -\text{Log}(a_{\text{Ca}^{2+}}) \quad (4)$$

189 ~~($pCa = -\log[\text{Ca}^{2+}]$)~~ (Fig. 5), we found two distinct correlation lines, one for thermal water and one for
190 lake water, suggesting that negatively-charged $[\text{Hf}(\text{OH})_5]^-$ complexes were preferentially scavenged
191 onto {104} positive carbonate surfaces and were, in turn, removed from the aqueous phase. This
192 potential mechanism explains the Zr/Hf fractionation observed during calcite precipitation under pH
193 conditions between 7 and 10 and should correspond to the higher binding energy between Hf and
194 {104} calcite surface relative to Zr as estimated by theoretical ab initio calculations (Pinto et al.,
195 2017). Our results support the Zr/Hf fractionation observed by Firdaus et al. (2011, 2018) during Fe-
196 oxyhydroxide formation. Fe-oxyhydroxide was in fact positively charged under weak alkaline
197 conditions (Koschinsky and Hein, 2003). On the contrary, Zr-Hf fractionation did not occur onto
198 surfaces of Mn-oxyhydroxides that were negatively charged at the pH up to acidic values (Tripathy
199 et al., 2001). The Zr-Hf fractionation here observed during calcite formation and interpreted as
200 water-interface interaction is consistent with the Zr-Hf fractionation found in stromatolites (Censi et
201 al., 2015), halite (Censi et al., 2017a, 2020), and other groundwater environments (Censi et al.,
202 2017b; Inguaggiato et al., 2015, 2016).

203 3.3 REE behaviour

204 Dissolved REE concentrations in lake water and thermal water in the Lake Van system
205 ranged between 7.135 and 76.140 pmol/L and from 168 to 17.426 pmol/L, respectively
206 (Supplementary Data 3). In lake waters the shale normalised pattern (Fig. 6aA) reveals the presence
207 of La positive anomaly as well an 'ascending' feature leading to enrichment in heavy REE (HREE,
208 from Ho to Lu) while thermal waters have a single slight medium REE (MREE, from Sm to Dy)
209 enrichment (Fig. 6bB). In both types of water, Eu anomalies were 0.3 and 4.8 without evidence of a
210 positive Ce anomaly ($0.1 \leq \text{Ce}/\text{Ce}^* \leq 1$). The progressive increase of REEs observed in the

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211 normalised patterns in these waters may result from the values of the stability constants of di-
212 carbonate and carbonate REE-complex ions (Lee and Byrne, 1993; Liu and Byrne, 1998) potentially
213 producing $[\text{REE}(\text{CO}_3)_2]^-$ complexes (Johannesson et al., 1994; Johannesson and Lyon, 1994; Kerrich
214 et al., 2002; Moller and Bau, 1993). Our data showed a higher abundance of La and therefore
215 inconsistent with earlier REE measurements from Van Lake waters (Moller and Bau, 1993). The
216 water level of Lake Van has dropped due to climate change over the last 30 years while the input of
217 lithogenic solids increased. The main difference between the REE distribution recognised in lake
218 Van waters and that reported by Moller and Bau (1993) is the lack of positive Ce anomaly in our
219 samples. This may be caused by a larger dissolution of lithogenic particles typical of continental and
220 coastal waters. The dissolution of lithogenic solids usually provides La enrichment in the aqueous
221 phase that can conceal the concurrent Ce enrichment completely or partially (Censi et al., 2007;
222 Greaves et al., 1999; Grenier et al., 2018).

223 Suspended particulate material collected from the lake and thermal waters had a total
224 concentration of REEs ranging between 508.2 and 835.2 nmol/kg and from 102.3 to 278.5 nmol/kg,
225 respectively (Supplementary Data 4). These values are 3 to 4 orders of magnitude higher than the
226 concentrations found in related waters. The shale-normalised REE patterns of suspended calcites
227 collected in lake waters (Fig. 6c) and thermal waters (Fig. 6d) display positive Ce anomalies ranging
228 from 1.1 to 3.1, an enrichment in MREEs with positive Eu anomalies ($0.8 < \text{Eu}/\text{Eu}^* < 4.2$), and
229 slight progressive depletion of HREEs. Suspended solids from thermal waters showed ‘bulged’ REE
230 patterns (Fig. 6d) while solids from lake waters had flatter patterns (Fig. 6c). The REE distribution
231 in shale-normalised patterns usually show MREE enrichment mainly ruled by crystal-chemical
232 reasons, as demonstrated by the review of distribution coefficient values reported by Moller and De
233 Lucia (2020). In suspended solids from thermal water, the MREE enrichment was more evident than
234 in materials collected from lake water. This evidence is probably justified by the concurrent presence
235 of a detrital fraction in the ‘bulge’ of suspended solids from lake water with a ‘flatter’ shale-like
236 trend in the normalised concentration (Taylor and McLennan, 1988).

237 The behaviour of Y-Ho twin is different in lake and thermal waters. In lakes, Y and Ho
238 concentrations ranged from 3,845 to 62,279 pmol/L and 36 to 686 pmol/L respectively, whereas in
239 thermal springs, Y ranges up to 3,577 pmol/L, whereas Ho fell between 2 and 68 pmol/L. The
240 related Y/Ho ratios spanned from 51 to 111 in lake water and between 53 and 162 in thermal water.
241 These ratios fell between chondritic and super-chondritic values. On the contrary, in carbonates from
242 suspended solids, the Y/Ho ratio ranged from 32.4 to 49.6 in lake water and from 28.6 and 47.8 in
243 thermal water. Then Y/Ho values in suspended solids spanned between slight subchondritic to
244 chondritic values ($Y/Ho = 52 \pm 5$, Jochum et al., 1986).

245 The role played by the formation of calcite in the distribution of REEs in alkaline waters was
246 highlighted by the super-chondritic value of the Y/Ho ratio. Our study revealed that Ho was 2 to 3
247 orders of magnitude higher than Y during calcite formation causing enrichment. The preferential
248 enrichment of Ho in calcite compared to Y observed in our study may reflect the larger value of the
249 Misono softness parameter of Ho^{3+} relative to Y^{3+} (Thompson et al., 2013) due to the different
250 external electronic configuration of these metal ions (Qu et al., 2009).

251 The different fractionation of these elements found during the formation of calcite also
252 reflects other processes. In thermal water, the intensity of the water-rock interaction may cause
253 MREE enrichment (Fig. 6c). Dissolution of lithogenic rocks may be the source of MREE (Bau,
254 1999; Censi et al., 2019; Greaves et al., 1999; Haley et al., 2004; Inguaggiato et al., 2016; Lin et al.
255 2019) and explain MREE enrichment observed in authigenic hydrothermal carbonates (Jakubowicz
256 et al., 2015; Phan et al., 2019; Wang et al., 2020), phosphates (Censi et al., 2007; Hannigan and
257 Sholkovitz, 2001; Reynard et al., 1999; Zhang et al., 2016), sulphates (Censi et al., 2014; 2018; Kagi
258 et al., 1993; Playà et al., 2007; Toulkeridis et al., 1998), and Fe-oxyhydroxides (Bau, 1999;
259 Davranche et al., 2011).

260 **4 Implications and Conclusions**

261 The results of our investigation of water oversaturated with calcite indicate that calcite plays
262 a crucial role in constraining Zr-Hf fractionation and the geochemical behaviour of REEs in alkaline

263 lake waters. The strong reactivity of positively charged { 104 } surfaces of calcite coupled with the
264 aqueous Zr and Hf speciation lead to preferential accumulation of Hf in calcite relative to Zr. This
265 indication agrees with the larger affinity of Hf towards solid surfaces relative to Zr already
266 evidenced onto the surfaces of Fe-oxyhydroxides. But the Hf removal onto the carbonate surfaces is
267 ruled by pCa, whereas onto the Fe-oxyhydroxide surfaces is ruled by pH. Therefore, changes of the
268 Zr/Hf ratio in natural waters relative to the chondritic signature indicate that these elements are
269 involved in processes occurring at the interface with crystallising authigenic minerals.

270 Differences in the external electronic configuration of Zr and Hf is similar to that between Y
271 and Ho due to differences in softness/hardness between Zr and Hf and between Y and Ho that, in
272 turn, influence hydrolysis during formation of Zr and Hf hydroxyl complexes. This process could be
273 a crucial aspect of the fractionation of these metal ions during calcite formation. Differences in
274 softness between Y and Ho lead increase stability of Ho-O relative to Y-O bonds on solid surfaces
275 and rule Y-Ho fractionation at the solid-liquid interface.

276 Based on previous researches addressed to the larger Hf affinity relative to Zr towards crystal
277 surfaces of halite, potash salts and Fe-oxyhydroxides, this study contributes to demonstrate that the
278 larger surface reactivity of Hf relative to Zr is a widespread phenomenon through the sedimentary
279 processes. The evidence of this phenomenon is detected measuring the Zr/Hf ratio in natural waters
280 and in authigenic minerals.

281

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286

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621

622 TABLE CAPTIONS

623 Table 1 - Sampling sites and geographical localisation of studied waters.

624 Table 2 - Mineralogical composition of studied suspended particulates in lake waters. Values are
625 given in weight %.

626

627 FIGURE CAPTIONS

628 Figure 1 – Geological sketch map of the studied area.

629 Figure 2 – Major element concentrations analysed in studied waters.

630 Figure 3 – SEM images of suspended solids in studied water samples filtered with membrane

631 porosity 0.45 μm . Fig. 3a, 3b, 3c, and 3d: Different amounts of build-up in suspended solids mixed

632 to authigenic carbonates in lake waters from different sites. Fig. 3e and 3f: Some particulars of well-

633 developed calcite crystals together with biologic remnants in lake waters. Fig. 3g and 3h: Authigenic

634 solids and lithic fragments are suspended in thermal springs. Fig. 3i: Fine-grained Ca-Mg carbonates

635 and Fe-bearing solids in suspended particles from thermal springs. The open square indicates the

636 area where data to depict X-ray Fluorescence maps of Fe, Ca, and Mg were collected.

637 Figure 4 – a: Zr and Hf concentration measured in studied water samples and the linear trend

638 depicted by these analyses. b: Zr and Hf concentration measured in suspended solids in studied

639 water samples. c: Zr and Hf concentration measured in halite crystallising from the Dead Sea brine

640 (data from Censi et al., 2020). Measured Zr and Hf concentrations are compared with the range of

641 Zr/Hf values characteristic of crustal rocks and chondrite ($\text{Zr/Hf} = 71 \pm 5.6$, Jochum et al., 1986).

642 Figure 5– Zr/Hf values measured in the dissolved phase of the lake and thermal waters reported

643 towards Ca^{2+} concentration (mol/L) expressed as pCa (see equation 44).

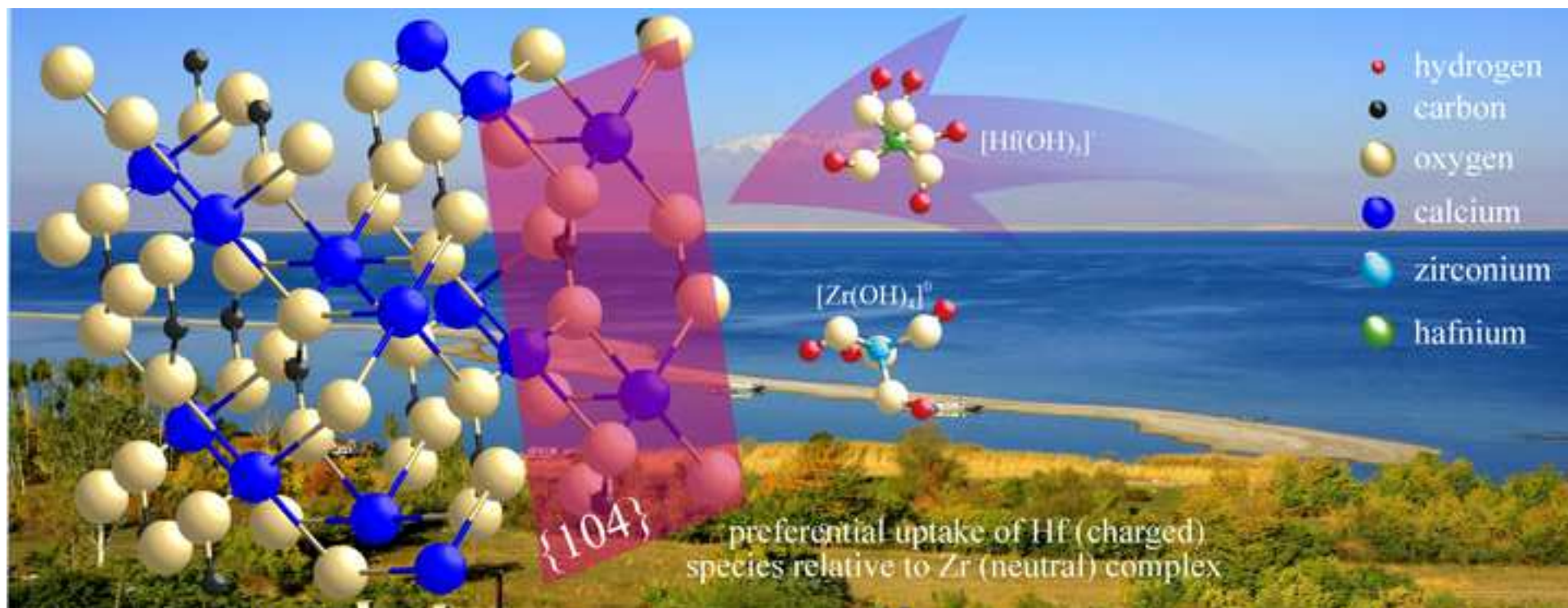
644 Figure 6 – Shale-normalised REE pattern calculated in the dissolved phase of the lake (a) and

645 thermal waters (b). These features are also depicted in terms of MREE/HREE and LREE/MREE

646 ratios (c). Shale-normalised REE pattern calculated for suspended solids in the lake (d) and thermal

647 waters (e). These features are also reported in terms of MREE/HREE and LREE/MREE ratios (f).

- Weathering and authigenic reactions rule the Zr/Hf ratio in natural waters
- Hf shows larger affinity than Zr towards the surface of authigenic calcite
- Calcite crystallisation rules the REE distribution in suspended solids from thermal waters
- Dissolution of airborne particles influences the REE distribution in shallow lake waters



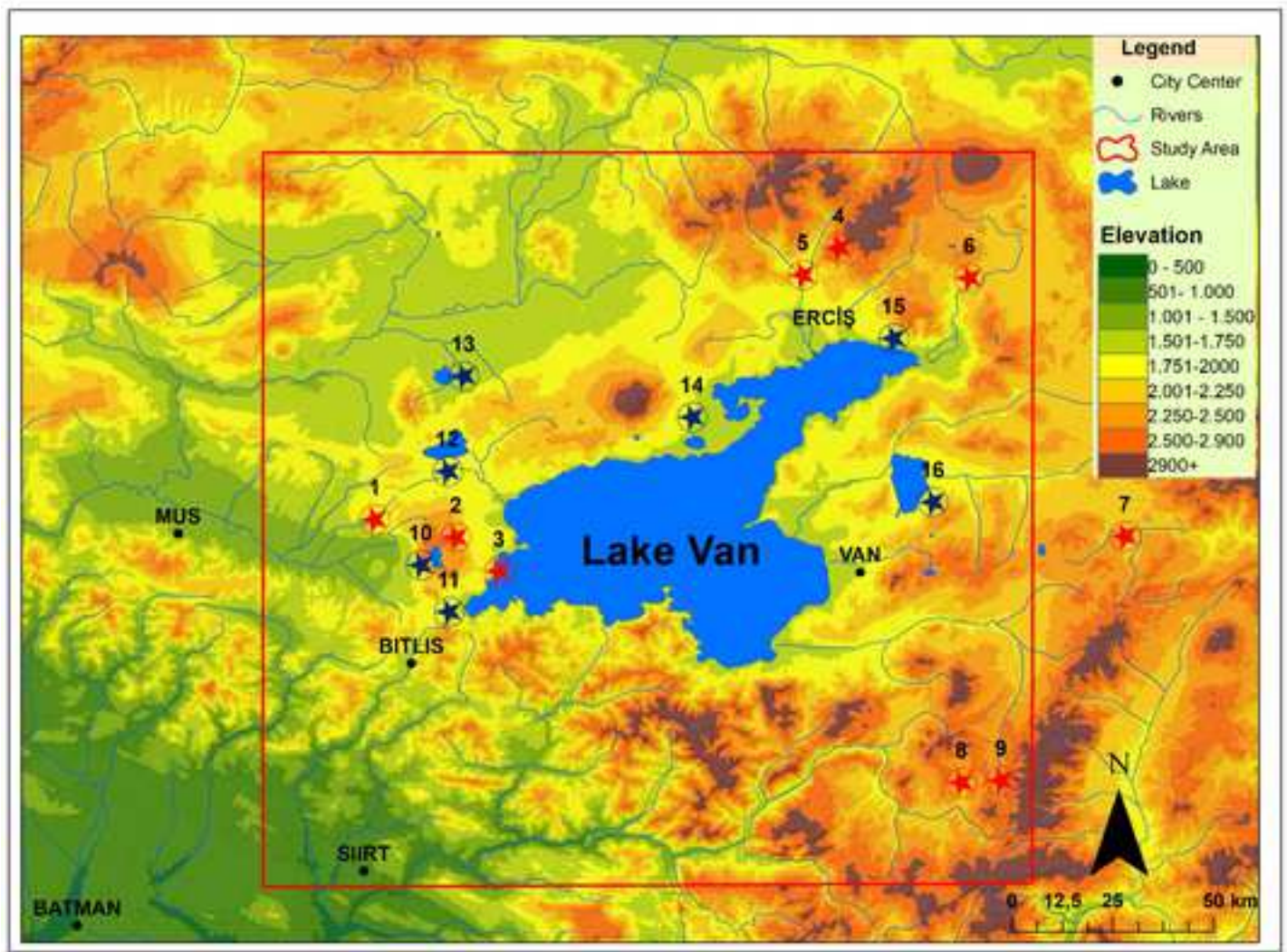


Fig. 1

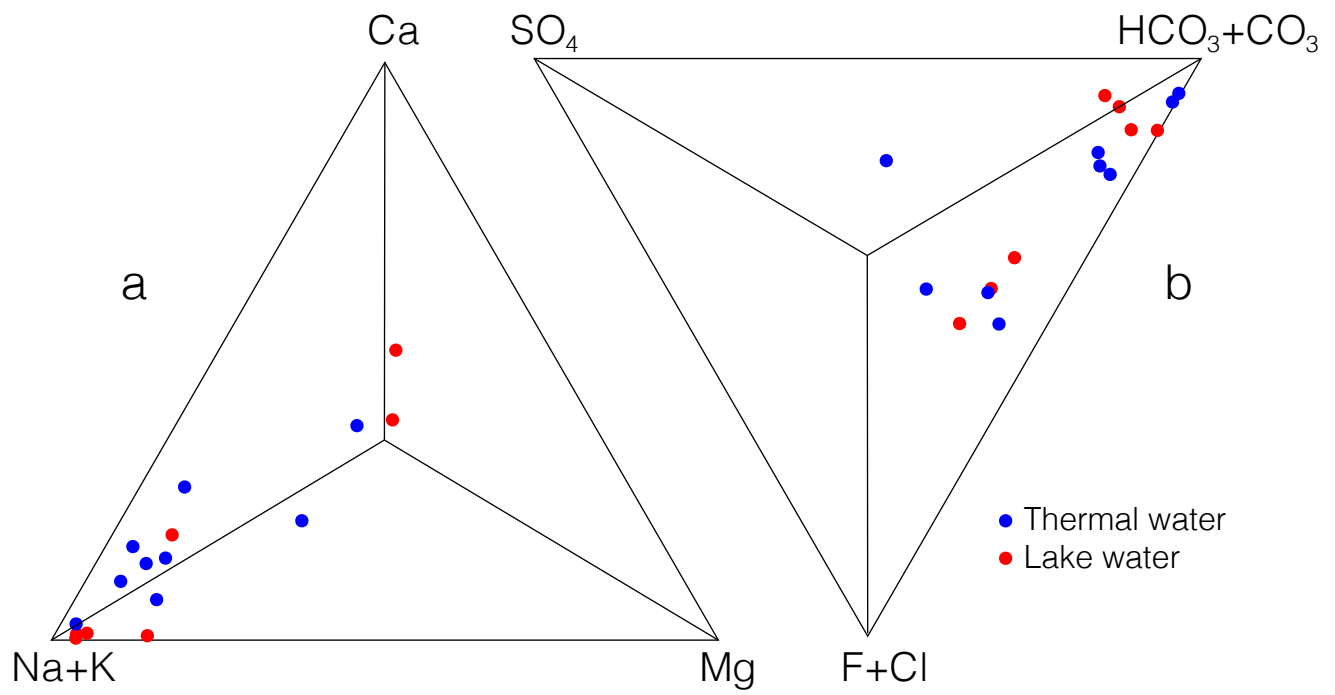


Fig. 2

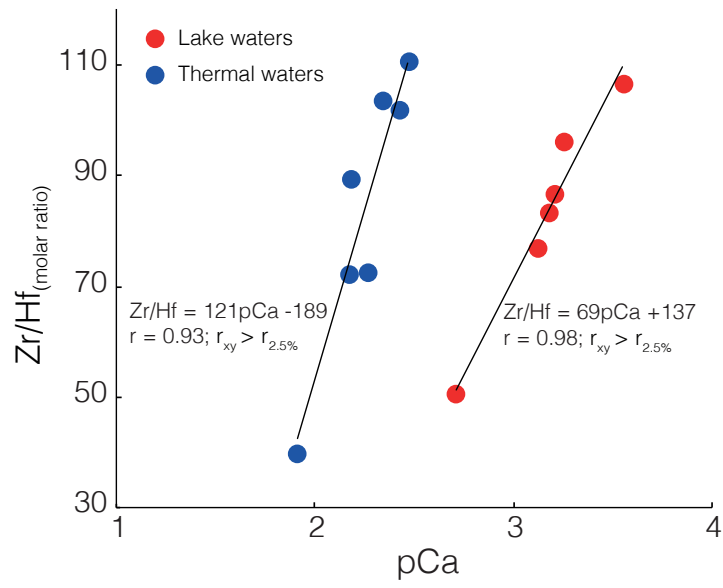


Fig. 5

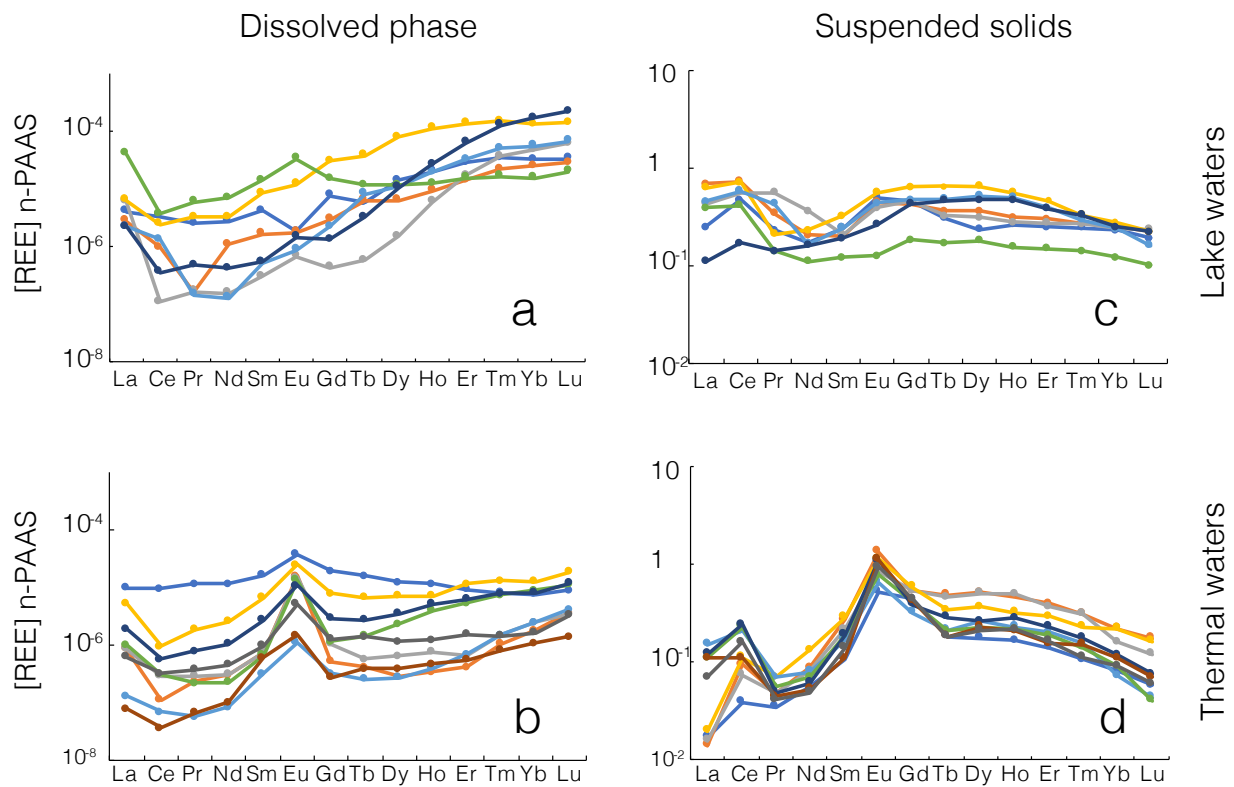


Fig. 6

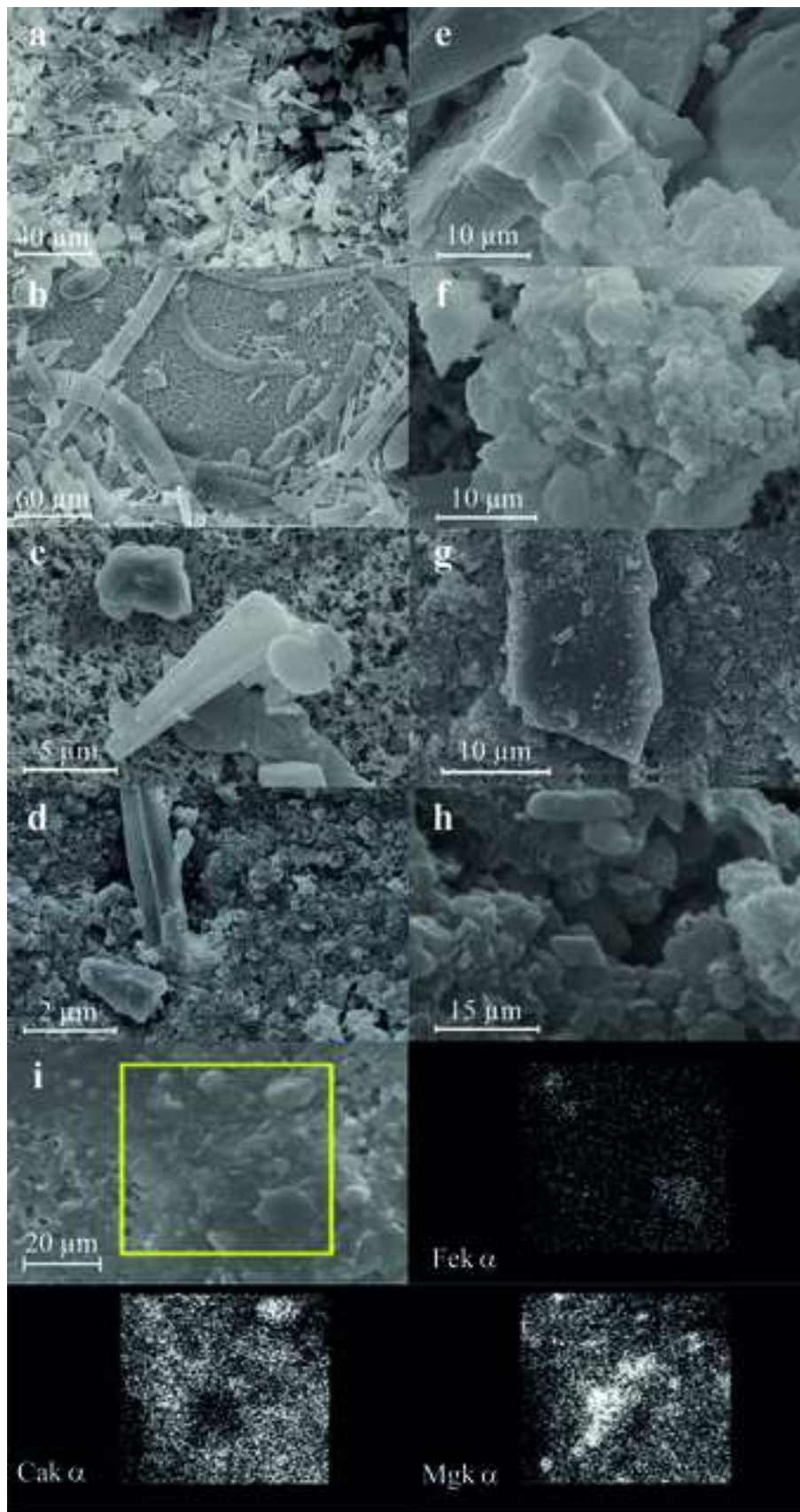


Fig. 3

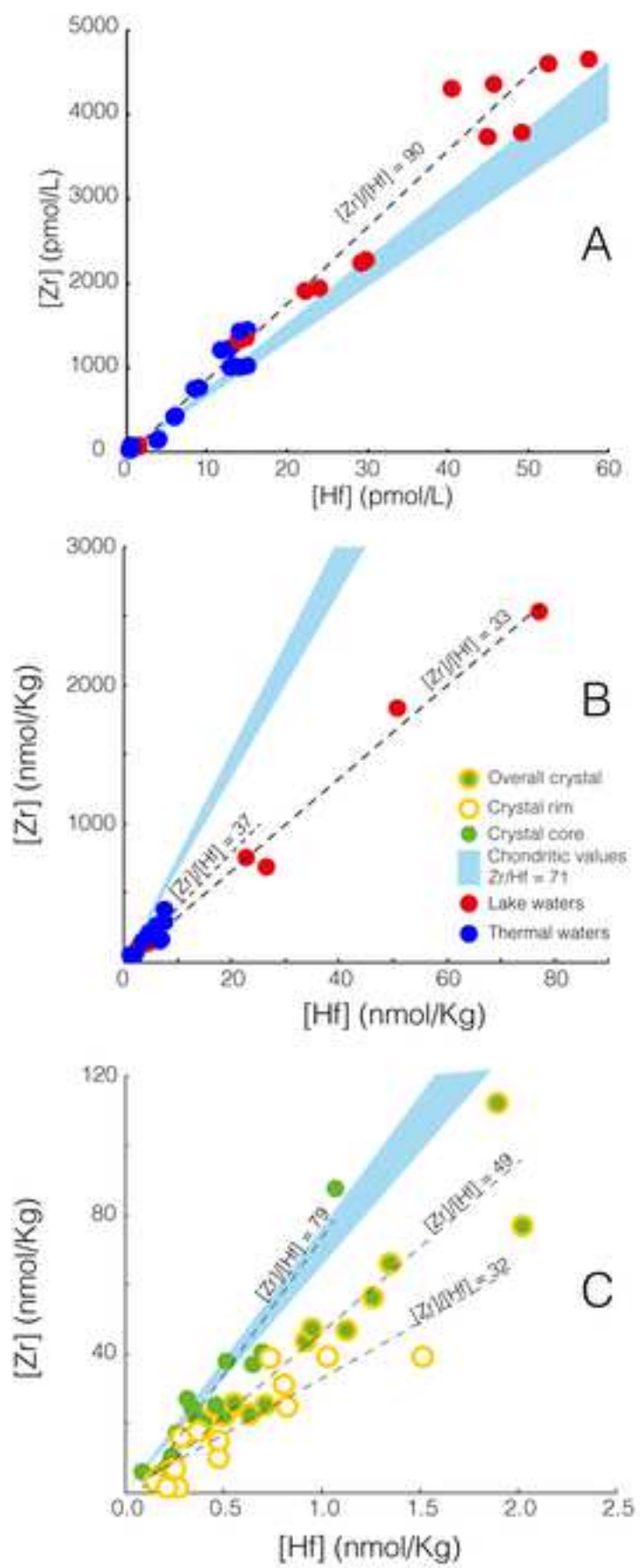


Fig. 4

Table 2

	Mg-calcite	Calcite	Dolomite	Goethite	Quartz
	VSL	58	17	13	12
	VEL	35	40	7	15
	VTL	35	46	6	13
Lake waters	VERL	80	12		8
	VNL	54	35		11
	VHL	68		17	15
	VNML	72	8	13	7

Table 1

Sample number	Location	Longitude (E)	Latitude (N)	Altitude (m)
VSL	Sodali Lake	42.950	38.815	1650
VEL	Van lake Ercis	43.455	38.999	1646
VTL	Van Lake Tatvan	42.336	38.522	1646
VERL	Ercek Lake	43.530	38.647	1803
VNL	Nazik Lake	42.294	38.852	1816
VHL	Hacli Lake	42.314	39.008	972
VNML	Nemrut Lake	42.228	38.623	2247
VSCT	Çaybagi Thermal	44.164	38.517	2070
VBDT	Dereici (Baskale) Thermal	43.951	37.838	2260
VBZT	Zereni (Baskale) Thermal	44.089	37.798	2180
VEHT	Hasan Apdal (Ercis) Thermal	43.388	39.225	1890
VGT	Gülsuhan (Tatvan) Thermal	42.394	38.591	1648
VCT	Caldiran Thermal	43.863	39.120	2050
VGCT	Cukur (Guroymak) Thermal	42.027	38.655	1480
VNMT	Nemrut Thermal	42.246	38.625	2250
VETT	Taskapi (Ercis) Thermal	43.430	39.270	1930



UNIVERSITÀ DEGLI STUDI DI PALERMO

Dipartimento di Scienze della Terra e del Mare (DiSTeM)

COD. FISC. 80023730825 ~ P.IVA 00605880822

Palermo, November 14, 2020

CONFLICT OF INTEREST DECLARATION

I subscribed Prof. Paolo Censi, as corresponding author of the paper: “*Zirconium and Hafnium fractionation and distribution of Rare Earth Elements in neutral-alkaline waters: Case study of Lake Van hydrothermal system, Turkey*”

DECLARE

that do not occur conflicts of interest in the submission of the above mentioned paper to *Journal of Geochemical Exploration*.



UNIVERSITÀ DEGLI STUDI DI PALERMO

Dipartimento di Scienze della Terra e del Mare (DiSTeM)

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All the authors contribute to the same extent to the study leading to the submission of the manuscript entitled “Zirconium and hafnium fractionation and distribution of Rare Earth Elements in neutral–alkaline waters: Case study of Lake Van hydrothermal system, Turkey”

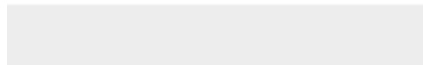
Sincerely yours,

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1 Zirconium and hafnium fractionation and distribution of Rare Earth Elements in neutral–alkaline
2 waters: Case study of Lake Van hydrothermal system, Turkey

3
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12
13 Abstract

14 We investigated the distribution of Zr, Hf, and rare earth elements (REE) as the sum of lanthanides
15 plus Y) in the hydrothermal system in the Lake Van area of south-eastern Turkey. This system is
16 characterised by water with variable pH in alkaline conditions resulting from hydrothermal CO₂
17 upraise and neoformation of calcite minerals in near equilibrium with the interacting waters. Zr, Hf,
18 and REE determinations were carried out for aqueous phases and suspended solids in lake water and
19 surrounding thermal springs. We found that dissolved Hf is partitioned relative to Zr during calcite
20 formation and that such fractionation is a function of the Ca²⁺ activity in warm water. The observed
21 Zr-Hf fractionation is explained by coulombic interactions that occur between suspended solid
22 particles and dissolved phases at the calcite-water interface. There, the surfaces of carbonate
23 minerals demonstrated greater reactivity towards aqueous Hf-bearing species relative to Zr-
24 complexes. This evidence involves a coulombic mechanism of reactivity at the calcite-water
25 interface because Hf complexes are negatively charged while Zr compounds are uncharged. Thus,

26 authigenic calcite can behave as a suitable host for dissolved metal ion species to adsorb on crystal
27 surfaces to remediate waste waters from mine drainage.

28

29 Keywords

30 Zr/Hf ratio; REE; Alkaline lakes; Turkey

31 **1 Introduction**

32 Zirconium, Hf, and rare-earth elements (REE) are used in specialised technology industries
33 notwithstanding their limited production (Chakhmouradian and Wall, 2012; Jones III et al., 2017).
34 Demand for these elements is high leading an effort by the mineral exploration industry to
35 understand genetic processes and discover new resources. Geochemically, these elements were
36 considered to be ‘immobile’ in subsurface conditions while some works indicate easy water
37 transport under subsurface hydrothermal conditions (Michard and Albarede, 1986; Michard, 1989).
38 Accordingly, REE enrichment has been observed in deposits formed from carbonisation in alkaline
39 environments where Zr concentrations were enriched by a factor of 3 to 4 (Estrade et al., 2015). The
40 behaviour of these strategic elements under surface and subsurface conditions is often neglected
41 because of difficulties in accurate analytical determination at the level of pmol/kg in natural water.
42 In a recent survey, the dissolved Zr/Hf ratio was different from that found in the rock minerals
43 (Censi et al., 2020; Cidu et al., 2013; Inguaggiato et al., 2015, 2016; Parisi et al., 2017; Tepe and
44 Bau, 2015; Zuddas et al., 2017, 2018) suggesting that the subsurface cycle is important in the
45 accumulation of these elements. In fact, the Zr/Hf molar value in rock minerals was 71 ± 6 which
46 corresponds to the so-called ‘chondritic’ signature (Jochum et al., 1986), while natural water had a
47 wider range of Zr/Hf (Zuddas et al., 2018). The supergene geochemical cycle of Zr and Hf
48 participates in the accumulation and stability of these elements through the process of dissolution
49 and precipitation of the minerals. Because the composition of subsurface and surface waters is often
50 near equilibrium with carbonate minerals like calcite, calcite is expected to play a significant role in
51 the supergene behaviour of these elements. Although carbonates are among the most common
52 minerals in equilibrium with natural water, the formation of neogenic calcite also determined the
53 change in REE distribution observed by Moller and De Lucia (2020) in controlled laboratory
54 investigations. To explain Zr-Hf decoupling during calcite crystallisation and the concurrent
55 behaviour of REEs, we investigated the distribution of these elements in a natural hydrothermal
56 system characterised by the formation of calcite under the neutral and alkaline water system found at

57 Lake Van and characterised by low thermal and lake waters with pH ranging from 6 to 10, resulting
58 from a variable CO₂ flux (Degens and Kurtman, 1978; Moller and Bau, 1993). The Lake Van system
59 is characterised by neogenic precipitation in the form of Ca-Mg carbonates (Moller and Bau, 1993).

60 **2 Material and Methods**

61 Lake Van is located in south-eastern Turkey (Fig. 1) which formed during the collision
62 between European and Arabian plates. The sequence starts with Paleozoic metamorphic rocks
63 overlaid by Permian limestones and schists followed by sediments of marine and continental origin
64 (Upper Cretaceous-Pliocene) closed by basic and felsic volcanic-pyroclastic rocks (Oligocene-
65 Recent) (Degens and Kurtman, 1978, Yilmaz, 1990). Heat flow data suggest the presence of a
66 magmatic body beneath the lake bottom, possibly related to Nemrut Volcano and sustaining input of
67 CO₂ into Lake Van waters (Moller and Bau, 1993). Presently, the Lake Van represents the largest
68 soda lake worldwide resulting from a combined volcanogenic hydrothermal CO₂ input and seasonal
69 evaporitic conditions (Moller and Bau, 1993).

70 **2.1 Field sampling**

71 Field study and water sampling were conducted in November 2018 and consisted of 16
72 samples collected from lake water and thermal springs. Coordinates and geologic units of the
73 collected samples are given in Table 1. Lake water was sampled at a depth of 15–20 cm.
74 Physicochemical parameters (pH, redox potential (Eh), temperature, and electrical conductivity)
75 were measured with an Eh oxytrode Pt probe (HamiltonTM) using a standard reference solution
76 buffer at 0.475±0.005 V. Accuracy of determinations was ±10 mV. For the major ion
77 determinations, water samples were filtered in the field through 0.45 µm Millipore® cellulose
78 acetate filters into 50 mL polyethylene bottles and acidified with ultrapure HNO₃. Water samples for
79 trace and rare element analyses were collected in 2,000 mL Nalgene bottles at each sampling site.

80 **2.2 Water analysis and thermodynamic modelling**

81 In the laboratory, samples were filtered (Millipore® manifold filter, 47 mm diameter with a
82 0.45 µm pore size cellulose nitrate membrane), then 5% ultrapure HNO₃ acid solution was added to

83 reach pH \approx 2. Zirconium, Hf, and REEs in aqueous samples were analysed according to the
 84 methods of Raso et al. (2013). Briefly, an excess of FeCl₃ (1%) solution was added to each sample
 85 (1 L), and a suitable volume of NH₄OH (25%) solution was added to attain a pH of 8 to induce
 86 precipitation of solid Fe(OH)₃. During this process, Zr, Hf, and REEs were held onto the surface of
 87 the crystallising solid. To ensure that crystallisation of Fe(OH)₃ was complete, the solution was left
 88 in a closed flask for 48–72 h in a stirrer. Next, the iron concentration was measured to assess
 89 recovery and was consistently greater than 95%. Precipitated Fe(OH)₃, together with Zr, Hf, and
 90 REEs, was collected on a membrane filter (Millipore® manifold filter, 47 mm diameter with
 91 0.45 μ m pore size). The solid filtrate was then dissolved in 3 M HCl, and the obtained solution was
 92 diluted to 1 M HCl prior to analysis using a Quadrupole-ICP-MS (Agilent 7800 series) with an
 93 external calibration procedure. The overall strategy and a detailed evaluation of the analytical errors
 94 were reported in Raso et al. (2013). The overall enrichment factor measured for Zr, Hf, and REE
 95 concentrations corresponded to 66.7.

96 Lanthanum, Ce, and Eu anomalies were calculated using equations 1–3 (Alibo and Nozaki,
 97 1999):

$$98 \quad \frac{La}{La^*} = \frac{[La]_n}{3*[Pr]_n + 2*[Nd]_n}, \quad (1)$$

$$99 \quad \frac{Ce}{Ce^*} = \frac{2*[Ce]_n}{[La]_n + [Pr]_n}, \text{ and} \quad (2)$$

$$100 \quad \frac{Eu}{Eu^*} = \frac{2*[Eu]_n}{[Sm]_n + [Gd]_n}, \quad (3)$$

101 where the subscript n indicates shale-normalised elemental concentrations calculated relative to the
 102 Post Archean Australian Shale (PAAS, Taylor and McLennan, 1995).

103 All chemicals used during laboratory manipulations were ultrapure grade. Ultrapure water
 104 (resistivity of 18.2 M Ω cm or better) was obtained from an Arium® mini system (Sartorius, Italy).
 105 HNO₃ 65%, NH₄OH (20%) and HCl (30%) solutions were purchased from J.T. Baker (ULTREX II)
 106 chemicals. Working standard solutions of the elements studied were prepared daily by stepwise
 107 dilution of standard multi-element stock solutions from DBH, Merck, or CPI International (1,000 \pm 5

108 mg/L) in a 1 M HCl medium. All labware was polyethylene, polypropylene, or Teflon® and the
109 calibration of all volumetric equipment was verified. Assessment of the analytical precision of Zr,
110 Hf, and REEs was performed on five aliquots (500 mL each) of NASS-6 (seawater certified
111 reference material for trace metals) distributed by the National Research Council of Canada. These
112 certified reference materials were treated as water samples according to the procedure mentioned
113 above and the concentrations obtained were compared with those previously reported by Raso et al.
114 (2013) and Lemaitre et al. (2014). Analysed values are reported in Supplementary Data 1.

115 Thermodynamic modelling of the chemical composition of the water performed was using
116 PHREEQC software and adding additional Lawrence Livermore National Laboratory (LLNL)
117 database (version 3.0.6, Parkhurst and Appelo, 2010).

118 **2.3 Suspended solids**

119 X-ray diffraction investigations were carried out with a Philips PW14 1373 X-ray
120 spectrometer (Cu-K α radiation 2 Θ range 10–40°, step size 0.02°, and a 1-min step time). The X-ray
121 spectra were analysed by the Rietveld method (Program DiffracPlus TOPAS®, Version 4.0, Bruker
122 AXS Inc., Karlsruhe, Germany) using parameters for the Rietveld refinement method obtained from
123 the Inorganic Crystal Structure Database (ICSD) program. This method consists of fitting the
124 experimental XRD spectrum to the theoretical spectrum (Young, 1993). A detailed explanation of
125 the method is reported in Scarlett and Madsen (2006).

126 Scanning electron microscopy (SEM) of suspended particles was carried out on materials collected
127 after filtration onto cellulose nitrate membranes. Dried solids were mounted on an aluminium stub
128 and gold coated. The SEM analyses were carried out using a LEO 440 SEM equipped with an EDS
129 system OXFORD ISIS Link and Si (Li) PENTAFET detector at the SIDERCEM SRL laboratory
130 (Caltanissetta, Italy). Chemical analyses were carried out by digesting 100 mg of each sample in 10
131 mL of a 1:1 HNO₃:H₂O₂ mixture in a sealed Teflon® TFM bomb using a microwave oven (CEM
132 MARS 5 device). After digestion, the TFM bomb was rinsed, and the solution was filtered onto

133 previously acid-cleaned 0.45 μm Millipore® filters to remove silicate residue. The filtered solution
134 was diluted to 50 mL and stored for chemical analyses.

135 **3 Results and Discussion**

136 **3.1 System description**

137 The concentration of major elements, pH, Eh, electric conductivity, and temperature are
138 reported in the Supplementary Data 2. Water samples had ionic strength ranging from 0.024 to 0.222
139 M in thermal water and between 0.004 and 0.443 M in lake water. Thermal water had pH between
140 5.8 and 7.1 and Eh between 0.01 and 0.17 V, while lake water had pH between 8 and 10 and Eh
141 from -0.29 and 0.01 V. Figure 2a shows that lake water preferentially falls in the field of alkaline
142 elements (only two samples have higher Ca^{2+} and Mg^{2+} concentrations) while thermal water fell
143 between alkaline and Ca^{2+} -rich field (few samples fell in the Mg^{2+} corner).

144 The SEM images of the suspended particulates revealed that suspended particulate matter in
145 lake water consisted of authigenic carbonates associated with biogenic detritus and lithic fragments
146 (Fig. 3a-f), while suspended solids in thermal waters were authigenic carbonates and Fe-rich
147 particles and lithogenic fragments (Fig. 3g-i). Iron-bearing particulates in the samples were often
148 mixed with authigenic calcite and dolomite (see elemental maps in Fig. 3i). The X-ray diffraction
149 analyses confirmed the presence of carbonate minerals identified as calcite and low-Mg calcite
150 representing up to 80% of the authigenic minerals (Table 2). Other phases collected were Fe-
151 oxyhydroxides and detritic silicate particles reaching 10–15% of the total mass.

152 Using the chemical composition of the collected samples, we estimated the saturation state
153 with respect to the mineral in equilibrium and found both lake water and thermal water were
154 saturated or oversaturated with respect to calcite or low-Mg calcite. Some water was also saturated
155 with respect to dolomite and goethite in agreement with the results of the X-ray determinations of
156 the collected suspended matter.

157 **3.2 Zr and Hf behaviour**

158 Zr and Hf concentrations in both aqueous and suspended particulate matter are reported in
159 Supplementary Data 3 and 4, respectively. In the aqueous phase, the concentration of Zr and Hf was
160 not homogenous. Lake waters had Zr concentrations between 79 and 4958 pmol/L while Hf
161 concentrations fell between 2 and 53 pmol/L, and thermal waters had Zr concentrations between 32
162 and 1.4 pmol/L; whereas Hf concentrations fell between 0.5 and 14 pmol/L. The concentrations of
163 Zr and Hf in thermal waters were 1 to 3 orders of magnitude lower than those found in lake water;
164 however, the concentration of Zr as a function of Hf was linearly correlated with an intercept equal
165 to zero and parametrising a single slope of 90 ± 2 (Fig. 4a). This value was higher than the Zr/Hf
166 value of 71 ± 6 for crustal materials and chondrites (Jochum et al., 1986) indicating that the calcite
167 precipitation fractionates the Zr/Hf ratio in water. In the suspended solids from lake water, Zr
168 concentration ranged from 99 to 2533 nmol/kg, whereas Hf fell between 2.6 and 77.1 nmol/kg.
169 Suspended solids from thermal water had Zr concentrations between 20.7 and 346.7 nmol/kg and Hf
170 between 0.6 and 6.8 nmol/kg. Similarly, Zr and Hf in solids collected from thermal water were lower
171 compared to suspended solids found in lake water. Figure 4B shows the concentration of Zr as a
172 function of Hf in neogenic calcite from lake water and describes a linear slope of 33.5 similar to
173 calcite in equilibrium with thermal water where the slope was 37.

174 These results show that the Zr/Hf ratio in water is different from that in crustal rocks and that
175 Hf is preferentially enriched in authigenic carbonate minerals. Preferential enrichment of Hf relative
176 to Zr in newly-formed suspended carbonates was unexpected and may result from processes
177 occurring at the surface-water interface. The calcite-water interface is generally represented by an
178 electric double-layer of weak mineral solubility buffered by CO_3^{2-} ions with the ζ potential of the
179 {104} carbonate surface independent of pH (Herberling et al., 2011). Laboratory measurements
180 indicate the positive surface potential (up to 24 mV) when the Ca^{2+} concentration in solution is
181 between 10^{-2} and 10^{-4} M (Al-Mahrouqui et al., 2017). Byrne (2002) made a critical review of
182 stability constants for Zr and Hf hydroxyl complexes based on the work of Sillen and Martell (1964),
183 Baes and Mesmer (1976), Ekberg et al. (2004), and Brown et al. (2005) and estimated that the most

184 abundant Zr and Hf complexes in aqueous phase were $[\text{Zr}(\text{OH})_4]^0$ and $[\text{Hf}(\text{OH})_5]^-$, respectively
185 (Koschinsky and Hein, 2003). Reporting Zr/Hf ratios as a function of pCa:

186
$$pCa = -\text{Log}(a_{Ca^{2+}})(4)$$

187 (Fig. 5), we found two distinct correlation lines, one for thermal water and one for lake water,
188 suggesting that negatively-charged $[\text{Hf}(\text{OH})_5]^-$ complexes were preferentially scavenged onto {104}
189 positive carbonate surfaces and were, in turn, removed from the aqueous phase. This potential
190 mechanism explains the Zr/Hf fractionation observed during calcite precipitation under pH
191 conditions between 7 and 10 and should correspond to the higher binding energy between Hf and
192 {104} calcite surface relative to Zr as estimated by theoretical ab initio calculations (Pinto et al.,
193 2017). Our results support the Zr/Hf fractionation observed by Firdaus et al. (2011, 2018) during Fe-
194 oxyhydroxide formation. Fe-oxyhydroxide was in fact positively charged under weak alkaline
195 conditions (Koschinsky and Hein, 2003). On the contrary, Zr-Hf fractionation did not occur onto
196 surfaces of Mn-oxyhydroxides that were negatively charged at the pH up to acidic values (Tripathy
197 et al., 2001). The Zr-Hf fractionation here observed during calcite formation and interpreted as
198 water-interface interaction is consistent with the Zr-Hf fractionation found in stromatolites (Censi et
199 al., 2015), halite (Censi et al., 2017a, 2020), and other groundwater environments (Censi et al.,
200 2017b; Inguaggiato et al., 2015, 2016).

201 **3.3 REE behaviour**

202 Dissolved REE concentrations in lake water and thermal water in the Lake Van system
203 ranged between 7135 and 76140 pmol/L and from 168 to 17426 pmol/L, respectively
204 (Supplementary Data 3). In lake waters the shale normalised pattern (Fig. 6a) reveals the presence of
205 La positive anomaly as well an ‘ascending’ feature leading to enrichment in heavy REE (HREE,
206 from Ho to Lu) while thermal waters have a single slight medium REE (MREE, from Sm to Dy)
207 enrichment (Fig. 6b). In both types of water, Eu anomalies were 0.3 and 4.8 without evidence of a
208 positive Ce anomaly ($0.1 \leq \text{Ce}/\text{Ce}^* \leq 1$). The progressive increase of REEs observed in the
209 normalised patterns in these waters may result from the values of the stability constants of di-

210 carbonate and carbonate REE-complex ions (Lee and Byrne, 1993; Liu and Byrne, 1998) potentially
211 producing $[\text{REE}(\text{CO}_3)_2]^-$ complexes (Johannesson et al., 1994; Johannesson and Lyon, 1994; Kerrich
212 et al., 2002; Moller and Bau, 1993). Our data showed a higher abundance of La and therefore
213 inconsistent with earlier REE measurements from Van Lake waters (Moller and Bau, 1993). The
214 water level of Lake Van has dropped due to climate change over the last 30 years while the input of
215 lithogenic solids increased. The main difference between the REE distribution recognised in lake
216 Van waters and that reported by Moller and Bau (1993) is the lack of positive Ce anomaly in our
217 samples. This may be caused by a larger dissolution of lithogenic particles typical of continental and
218 coastal waters. The dissolution of lithogenic solids usually provides La enrichment in the aqueous
219 phase that can conceal the concurrent Ce enrichment completely or partially (Censi et al., 2007;
220 Greaves et al., 1999; Grenier et al., 2018).

221 Suspended particulate material collected from the lake and thermal waters had a total
222 concentration of REEs ranging between 508.2 and 835.2 nmol/kg and from 102.3 to 278.5 nmol/kg,
223 respectively (Supplementary Data 4). These values are 3 to 4 orders of magnitude higher than the
224 concentrations found in related waters. The shale-normalised REE patterns of suspended calcites
225 collected in lake waters (Fig. 6c) and thermal waters (Fig. 6d) display positive Ce anomalies ranging
226 from 1.1 to 3.1, an enrichment in MREEs with positive Eu anomalies ($0.8 < \text{Eu}/\text{Eu}^* < 4.2$), and
227 slight progressive depletion of HREEs. Suspended solids from thermal waters showed ‘bulged’ REE
228 patterns (Fig. 6d) while solids from lake waters had flatter patterns (Fig. 6C). The REE distribution
229 in shale-normalised patterns usually show MREE enrichment mainly ruled by crystal-chemical
230 reasons, as demonstrated by the review of distribution coefficient values reported by Moller and De
231 Lucia (2020). In suspended solids from thermal water, the MREE enrichment was more evident than
232 in materials collected from lake water. This evidence is probably justified by the concurrent presence
233 of a detrital fraction in the ‘bulge’ of suspended solids from lake water with a ‘flatter’ shale-like
234 trend in the normalised concentration (Taylor and McLennan, 1988).

235 The behaviour of Y-Ho twin is different in lake and thermal waters. In lakes, Y and Ho
236 concentrations ranged from 3845 to 62279 pmol/L and 36 to 686 pmol/L respectively, whereas in
237 thermal springs, Y ranges up to 3577 pmol/L, whereas Ho fell between 2 and 68 pmol/L. The related
238 Y/Ho ratios spanned from 51 to 111 in lake water and between 53 and 162 in thermal water. These
239 ratios fell between chondritic and super-chondritic values. On the contrary, in carbonates from
240 suspended solids, the Y/Ho ratio ranged from 32.4 to 49.6 in lake water and from 28.6 and 47.8 in
241 thermal water. Then Y/Ho values in suspended solids spanned between slight subchondritic to
242 chondritic values ($Y/Ho = 52 \pm 5$, Jochum et al., 1986).

243 The role played by the formation of calcite in the distribution of REEs in alkaline waters was
244 highlighted by the super-chondritic value of the Y/Ho ratio. Our study revealed that Ho was 2 to 3
245 orders of magnitude higher than Y during calcite formation causing enrichment. The preferential
246 enrichment of Ho in calcite compared to Y observed in our study may reflect the larger value of the
247 Misono softness parameter of Ho^{3+} relative to Y^{3+} (Thompson et al., 2013) due to the different
248 external electronic configuration of these metal ions (Qu et al., 2009).

249 The different fractionation of these elements found during the formation of calcite also
250 reflects other processes. In thermal water, the intensity of the water-rock interaction may cause
251 MREE enrichment (Fig. 6c). Dissolution of lithogenic rocks may be the source of MREE (Bau,
252 1999; Censi et al., 2019; Greaves et al., 1999; Haley et al., 2004; Inguaggiato et al., 2016; Lin et al.
253 2019) and explain MREE enrichment observed in authigenic hydrothermal carbonates (Jakubowicz
254 et al., 2015; Phan et al., 2019; Wang et al., 2020), phosphates (Censi et al., 2007; Hannigan and
255 Sholkovitz, 2001; Reynard et al., 1999; Zhang et al., 2016), sulphates (Censi et al., 2014; 2018; Kagi
256 et al., 1993; Playà et al., 2007; Toulkeridis et al., 1998), and Fe-oxyhydroxides (Bau, 1999;
257 Davranche et al., 2011).

258 **4 Implications and Conclusions**

259 The results of our investigation of water oversaturated with calcite indicate that calcite plays
260 a crucial role in constraining Zr-Hf fractionation and the geochemical behaviour of REEs in alkaline

261 lake waters. The strong reactivity of positively charged {104} surfaces of calcite coupled with the
262 aqueous Zr and Hf speciation lead to preferential accumulation of Hf in calcite relative to Zr. This
263 indication agrees with the larger affinity of Hf towards solid surfaces relative to Zr already
264 evidenced onto the surfaces of Fe-oxyhydroxides. But the Hf removal onto the carbonate surfaces is
265 ruled by pCa, whereas onto the Fe-oxyhydroxide surfaces is ruled by pH. Therefore, changes of the
266 Zr/Hf ratio in natural waters relative to the chondritic signature indicate that these elements are
267 involved in processes occurring at the interface with crystallising authigenic minerals.

268 Differences in the external electronic configuration of Zr and Hf is similar to that between Y
269 and Ho due to differences in softness/hardness between Zr and Hf and between Y and Ho that, in
270 turn, influence hydrolysis during formation of Zr and Hf hydroxyl complexes. This process could be
271 a crucial aspect of the fractionation of these metal ions during calcite formation. Differences in
272 softness between Y and Ho lead increase stability of Ho-O relative to Y-O bonds on solid surfaces
273 and rule Y-Ho fractionation at the solid-liquid interface.

274 Based on previous researches addressed to the larger Hf affinity relative to Zr towards crystal
275 surfaces of halite, potash salts and Fe-oxyhydroxides, this study contributes to demonstrate that the
276 larger surface reactivity of Hf relative to Zr is a widespread phenomenon through the sedimentary
277 processes. The evidence of this phenomenon is detected measuring the Zr/Hf ratio in natural waters
278 and in authigenic minerals.

279

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284

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458

459 TABLE CAPTIONS

460 Table 1 - Sampling sites and geographical localisation of studied waters.

461 Table 2 - Mineralogical composition of studied suspended particulates in lake waters. Values are
462 given in weight %.

463

464 FIGURE CAPTIONS

465 Figure 1 – Geological sketch map of the studied area.

466 Figure 2 – Major element concentrations analysed in studied waters.

467 Figure 3 – SEM images of suspended solids in studied water samples filtered with membrane
468 porosity 0.45 μm . Fig. 3a, 3b, 3c, and 3d: Different amounts of build-up in suspended solids mixed
469 to authigenic carbonates in lake waters from different sites. Fig. 3e and 3f: Some particulars of well-
470 developed calcite crystals together with biologic remnants in lake waters. Fig. 3g and 3h: Authigenic
471 solids and lithic fragments are suspended in thermal springs. Fig. 3i: Fine-grained Ca-Mg carbonates
472 and Fe-bearing solids in suspended particles from thermal springs. The open square indicates the
473 area where data to depict X-ray Fluorescence maps of Fe, Ca, and Mg were collected.

474 Figure 4 – a: Zr and Hf concentration measured in studied water samples and the linear trend
475 depicted by these analyses. b: Zr and Hf concentration measured in suspended solids in studied
476 water samples. c: Zr and Hf concentration measured in halite crystallising from the Dead Sea brine
477 (data from Censi et al., 2020). Measured Zr and Hf concentrations are compared with the range of
478 Zr/Hf values characteristic of crustal rocks and chondrite ($\text{Zr}/\text{Hf} = 71 \pm 6$, Jochum et al., 1986).

479 Figure 5– Zr/Hf values measured in the dissolved phase of the lake and thermal waters reported
480 towards Ca^{2+} concentration (mol/L) expressed as pCa (see equation 4).

481 Figure 6 – Shale-normalised REE pattern calculated in the dissolved phase of the lake (a) and
482 thermal waters (b). These features are also depicted in terms of MREE/HREE and LREE/MREE
483 ratios (c). Shale-normalised REE pattern calculated for suspended solids in the lake (d) and thermal
484 waters (e). These features are also reported in terms of MREE/HREE and LREE/MREE ratios (f).